

**United States Air Force
611th Air Support Group/
Civil Engineering Squadron**



Elmendorf AFB, Alaska

Risk Assessment

**Barter Island Radar Installation,
Alaska**

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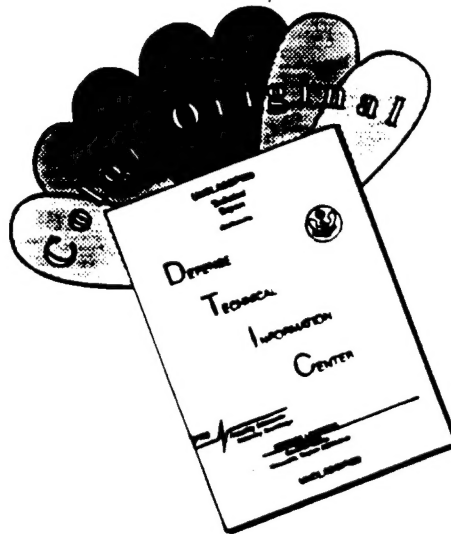
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Alaska**

Prepared by:

ICF Technology Incorporated

05 MAY 1995

PREFACE

This report presents the findings of Risk Assessments at sites located at the Barter Island radar installation in northern Alaska. The sites were characterized based on sampling and analyses conducted during Remedial Investigation activities performed during August and September 1993. This report was prepared by ICF Technology Incorporated.

This report was prepared between January 1995 and May 1995. Mr. Samer Karmi of the Air Force Center for Environmental Excellence was the Alaska Restoration Team Chief for this task.

Approved:

Thomas McKinney
Program Director
ICF Technology Incorporated

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LIST OF ACRONYMS AND ABBREVIATIONS

ADD	Average Daily Dose
Air Force	United States Air Force
ANWR	Alaska National Wildlife Refuge
ARAR	Applicable or Relevant and Appropriate Requirements
ATSDR	Agency for Toxic Substances and Disease Registry
BCF	Bioconcentration Factors
CDI	Chronic Daily Intake
COE	U.S. Army Corps of Engineers
COCs	Chemicals of Concern
DEW	Distant Early Warning
DRPH	Diesel Range Petroleum Hydrocarbons
ECAO	Environmental Criterion Assessment Office of EPA
EPA	U.S. Environmental Protection Agency
ERA	Ecological Risk Assessment
HI	Hazard Index
ha	Hectare
HEAST	Health Effects Assessment Summary Tables
HSDB	Hazardous Substance Data Bank
IRIS	Integrated Risk Information System
IRP	Installation Restoration Program
IS	Onsite Dietary Intake
LADD	Lifetime Average Daily Dose
LOAEL	Lowest-Observed Adverse Effect Level

LIST OF ACRONYMS AND ABBREVIATIONS (CONTINUED)

LRRS	Long Range Radar Station
MDEP	Massachusetts Department of Environmental Protection
MOGAS	Motor Vehicle Gasoline
MSL	Mean Sea Level
NOAEL	No Observed Adverse Effect Level
NOEL	No Observed Effect Level
NAS	National Academy of Science
PAHs	Polycyclic Aromatic Hydrocarbons
PCBs	Polychlorinated Biphenyls
RBSL	Risk-Based Screening Level
RfD	Reference Dose
RIs	Remedial Investigations
RI/FS	Remedial Investigation/Feasibility Study
RME	Reasonable Maximum Exposure
SIF	Scaling Factor
SF	Slope Factor
TPH	Total Petroleum Hydrocarbon
TRVs	Toxicity Reference Values
VOCs	Volatile Organic Compounds
UCL	Upper Confidence Limit
UF	Uncertainty factors
USFWS	U.S. Fish and Wildlife Service
USGS	United States Geological Survey

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1.0 INTRODUCTION

This document contains the baseline human health risk assessment and the ecological risk assessment (ERA) for the Barter Island Distant Early Warning (DEW) Line radar installation. Fourteen sites at the Barter Island radar installation underwent remedial investigations (RIs) during the summer of 1993. The presence of chemical contamination in the soil, sediments, and surface water at the installation was evaluated and reported in the Barter Island Remedial Investigation/Feasibility Study (RI/FS) [United States Air Force (Air Force) 1994a]. The analytical data reported in the RI/FS form the basis for the human health and ecological risk assessment. The primary contaminants of concern at the 14 sites are diesel and gasoline from past spills and/or leaks. The general location of the Barter Island radar installation is shown in Figure 1-1. The 14 sites investigated and the types of samples collected at each site are presented in Table 1-1.

The purpose of the risk assessment is to evaluate the human and ecological health risks that may be associated with chemicals released to the environment at the 14 sites investigated during the RIs. The risk assessment characterizes the probability that measured concentrations of hazardous chemical substances will cause adverse effects in humans or the environment in the absence of remediation. The risk assessment will be used to determine if remediation (site cleanup) is necessary and also to rank sites for remedial action. Additionally, it will be used as a model for the risk assessment to be performed at the other DEW Line installations (Bullen Point, Oliktok Point, Point Lonely, Barrow Point, Wainwright, and Point Lay) and the Cape Lisburne radar installation.

1.1 ORGANIZATION OF REPORT

Section 1.0 contains introductory information regarding the installation location and conditions, and a summary outline of the approach to the human health and ecological risk assessments. Section 2.0 is the Baseline Human Health Risk Assessment and Section 3.0 is the Ecological Risk Assessment. Section 2.0, Baseline Human Health Risk Assessment, is composed of:

- **Selection of Site Contaminants.** Presents the chemicals of concern (COCs) for human health and describes how they were selected for this risk assessment.
- **Exposure Assessment.** Identifies the pathways by which potential human exposures could occur, and estimates the magnitude, frequency, and duration of those exposures.
- **Toxicity Assessment.** Summarizes the toxicity of the selected COCs and the relationship between magnitude of exposure and the development of adverse health effects.

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TABLE 1-1. SITES EVALUATED AT BARTER ISLAND DEW LINE INSTALLATION

SITE NAME	SITE ID NUMBER	SOIL	SEDIMENTS	SURFACE WATER
Old Landfill	LF01	X	X	X
POL Catchment	LF03	X	X	X
Current Landfill	LF04	X	X	X
Contaminated Ditch	SD08	X	X	X
Old Runway Dump	LF12	X	NA	NA
Heated Storage Building	SS13	X	X	X
Garage	SS14	X	X	X
Weather Station Building	SS15	X	X	NA
White Alice Facility	SS16	X	X	NA
POL Tanks	ST17	X	X	NA
Fuel Tanks	ST18	X	X	NA
Old Dump Site	LF19	X	X	X
Bladder Diesel Spill	SS20	X	X	X
JP-4 Spill	SS21	X	X	NA

X Chemical analyses were performed on these media.

NA No chemical analysis was performed.

- **Risk Characterization.** Integrates the toxicity and exposure assessments to estimate the potential risks to human health from exposure to chemicals in environmental media.
- **Risk Assessment Uncertainty.** Describes the potential shortcomings in the data and the methods used to develop the risk assessment, and the uncertainties in the interpretation of the data and the risk characterization results.

Section 3.0, the Ecological Risk Assessment, is composed of:

- **Selection of Site Contaminants.** Presents the COCs for ecological receptors and describes how they were selected for the ERA.

- **Ecological Exposure Assessment.** Identifies the potential receptors and representative species, habitat suitability, and exposure pathways.
- **Ecological Toxicity Assessment.** Describes the potential effects of site contaminants on the representative species.
- **Risk Characterization for Ecological Receptors.** Evaluates the likelihood of adverse effects on ecological receptors.
- **Ecological Uncertainty Analysis.** Describes the potential shortcomings in the data and the methods used to develop the ERA, and the uncertainties in the interpretation of the data and the ecological risk characterization results.

Appendix A contains the human health risk assessment spreadsheets used to estimate chemical intake, noncancer hazard, and excess lifetime cancer risk. Appendix B consists of toxicology profiles. The estimated exposure calculations for ecological receptors are presented in Appendix C. Appendix D contains the RI analytical data for all sites from which the contaminants of concern were selected upon which the human health and ecological risk assessments are based.

1.2 RISK ASSESSMENT GUIDANCE DOCUMENTS

The following guidance documents were used to develop the human health and ecological risk assessments:

- *Risk Assessment Guidance for Superfund: Volume 1, Human Health Evaluation Manual (Part A)* [U.S. Environmental Protection Agency (EPA) 1989a]
- Region 10 Supplemental Risk Assessment Guidance for Superfund (EPA 1991a)
- *Risk Assessment Guidance for Superfund: Volume 2, Environmental Evaluation Manual* (EPA 1989b)
- General Guidance for Ecological Risk Assessment at Air Force Bases (MITRE 1990)
- Handbook to Support the Installation Restoration Program (IRP) Statements of Work (Air Force 1991).
- Framework for Ecological Risk Assessment (EPA 1992a)

1.3 INSTALLATION DESCRIPTION AND ENVIRONMENTAL SETTING

The Barter Island DEW Line installation is located on a 4,353-acre island at 70°08'N, 143°35'W, approximately 75 miles west of the Canadian border on the Arctic Coastal Plain of Alaska,

adjacent to the native village of Kaktovik. The area location of the Barter Island radar installation is shown on Figure 1-2 and a site plan is provided in Figure 1-3. The island consists of low-lying tundra on the northern border of the Alaska National Wildlife Refuge (ANWR). The maximum elevation on Barter Island is 55 feet above mean sea level and drainage is radially away from the high points. The Barter Island radar installation is situated adjacent to the northern coast, on a relatively flat area below a gradual slope. Upgradient from the installation is a freshwater lake used as a drinking supply for the installation and Kaktovik village. Kaktovik, population approximately 225, is located east of the station, adjacent to Kaktovik Bay. The community of Kaktovik relocated in 1947, 1952, and 1964 to accommodate the establishment and expansion of the Barter Island radar installation. Land use prior to the construction of the radar installation was a mix of natural undisturbed tundra and areas where Kaktovik village was situated.

Barter Island radar installation, also known as BAR-M, was the prototype DEW Line station and has been in active use since 1952. The installation operates and maintains a radar and communication system for detection of potential enemy aircraft entering U.S. airspace. Approximately seven contract personnel currently are stationed at the installation. The radar installation has been upgraded over the years with modern electronic equipment and in recent years has been called a Long Range Radar Station (LRRS). The installation consists of two module trains, a power plant, garages, warehouses, a hangar, a weather building, and radar and communication structures.

Module train "A" houses the electronics equipment work areas and the radar tower, limited personnel quarters, administration offices, a mechanical room with emergency boiler and fuel storage, and a personnel support module with water storage, shower, and toilets. Adjacent to this structure, and connected by corridors, are the power plant and vehicle maintenance buildings.

Train "B" is the main living and personnel support area. The dining and recreation areas are located in the center of this module train. Mechanical rooms are distributed throughout the living quarters. The solid waste incinerator is connected by a corridor to module "B".

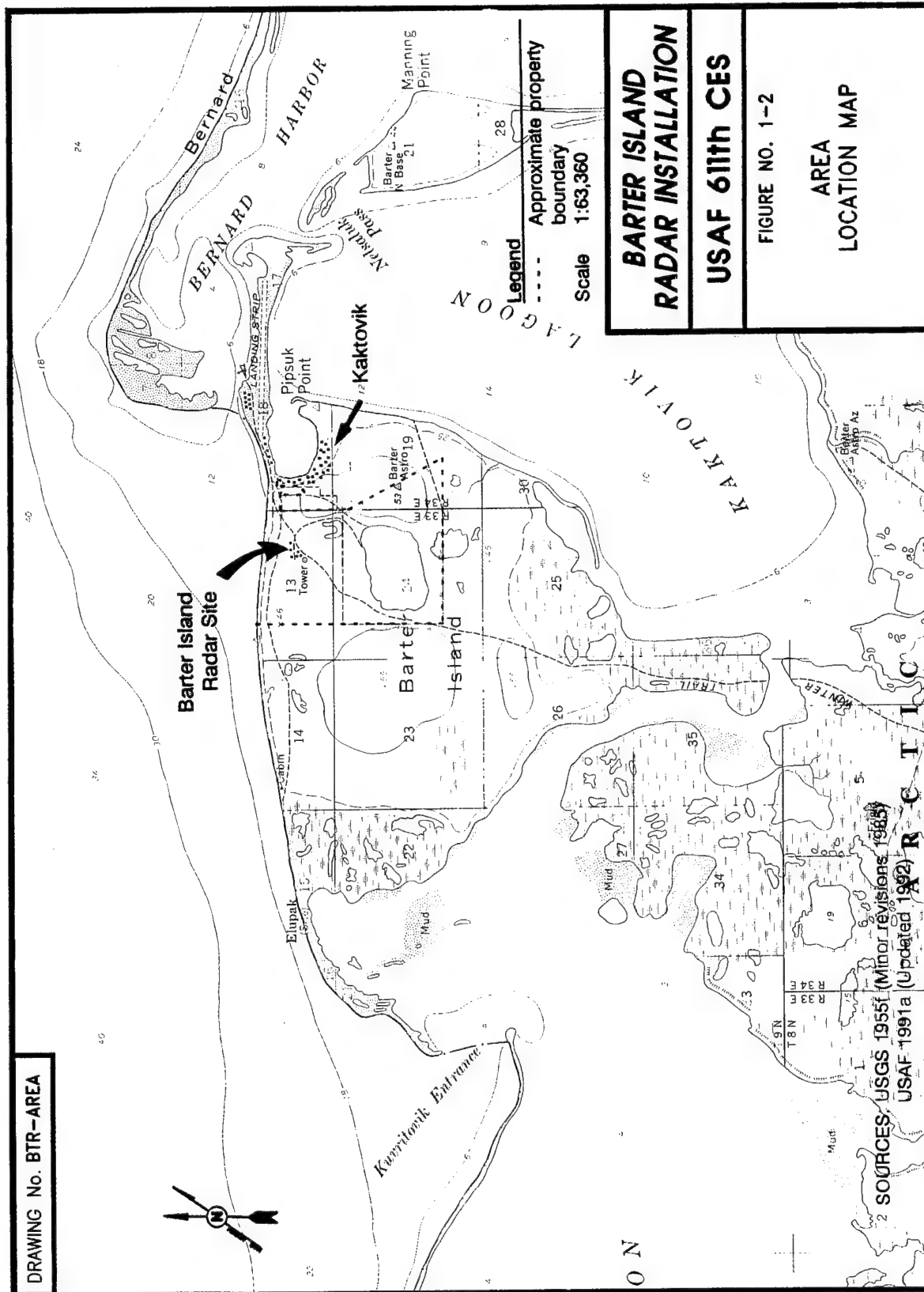
The rotating radar is in the radome. The radome, platform, plenum, and stairwell/cable chase are supported by steel columns and trusses on the west end of module train "A". There are also two communications billboards (known as "White Alices") and a radio relay building located south of the module trains.

Aircraft facilities include a lighted gravel runway and hangar. The runway is 4,820 feet long with a 360-foot overrun on the east end and a 480-foot overrun on the west end. The total length of the runway, with overruns, is 5,660 feet.

Precipitation at the Barter Island averages 7 inches per year, which includes 45 inches of snow. Average daily minimum and maximum temperatures in summer are 30°F and 46°F, respectively. In winter, these temperatures are -20°F and 6°F, respectively. Prevailing winds are easterly and average near 13 mph. The habitat on the island consist of Arctic tundra with several large and small lakes and a few streams that flow to the ocean. The island is predominantly covered by a thin tundra mat, beneath which is a layer of sand and loess (wind blown silt) approximately 2

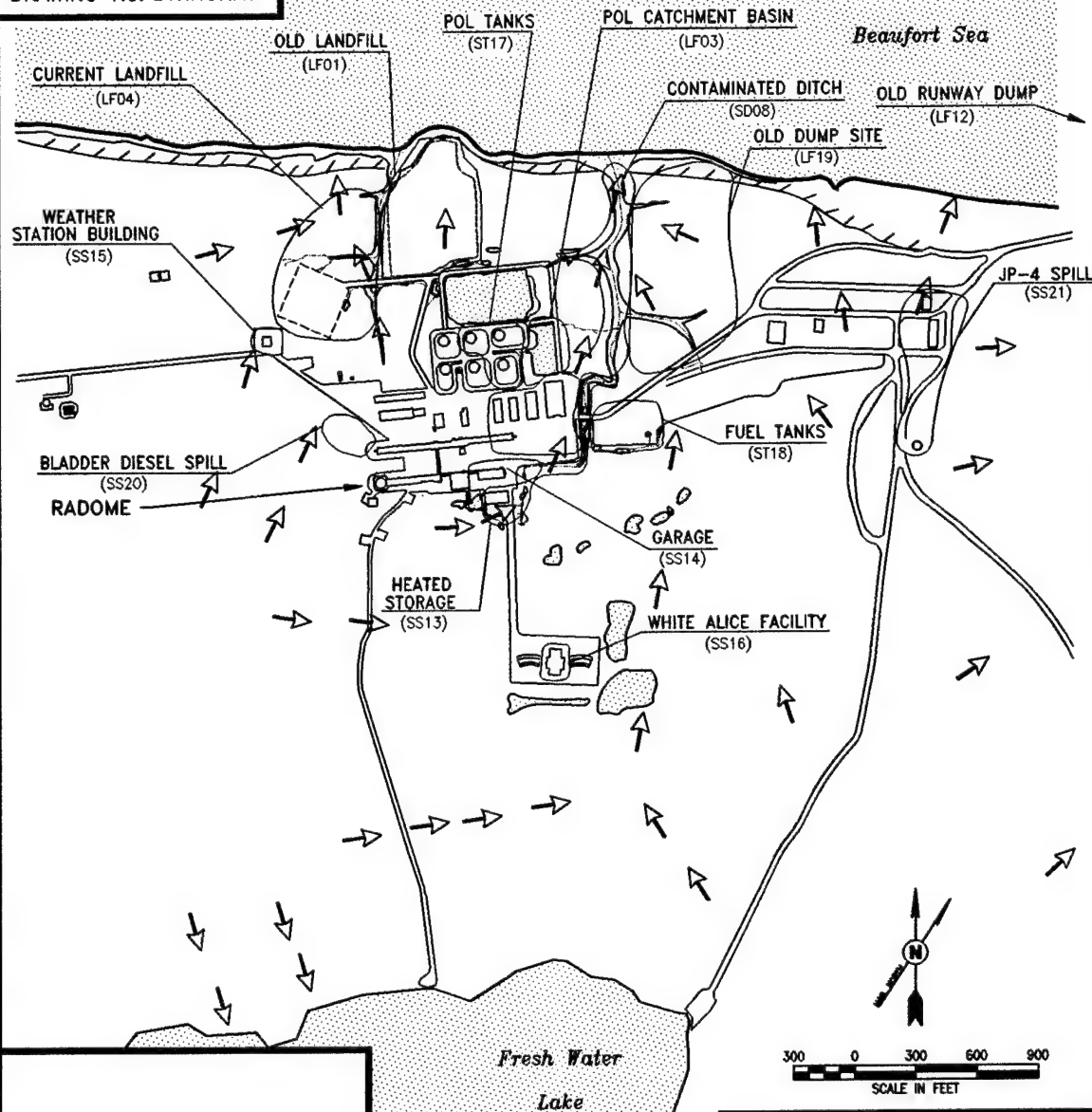
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LEGEND

- BUILDINGS, STRUCTURES
- ROADS
- TUNDRA
- SURFACE WATER
- CULVERT
- GRAVEL PAD BOUNDARY
- SURFACE DRAINAGE
- RI AREAS OF CONCERN

**BARTER ISLAND
RADAR INSTALLATION**

USAF 611th CES

FIGURE NO. 1-3

INSTALLATION
SITE PLAN

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to 3 feet thick. Underlying these deposits are lenses and layers of marine and alluvial clay, silt, sand, and sandy gravel of the Meade River Unit of the Gubik Formation. Permafrost in the area is up to 1,300 feet thick. A detailed description of the geology, soil, hydrology, meteorology, and biology of Barter Island is presented in the Barter Island RI/FS report (Air Force 1995a).

The Barter Island radar installation was investigated to evaluate possible contamination related to Air Force activities and historical waste disposal practices at the sites. Fourteen sites at the Barter Island radar installation were determined to be of potential concern based on previous IRP sampling activities, literature search, pre-survey and reconnaissance trips, interviews with station personnel, and information on disposal practices at DEW Line stations. The sites were investigated during RI/FS activities to confirm the presence or absence of chemical contamination; define the extent and magnitude of confirmed chemical releases; gather adequate data to determine the magnitude of potential risks to human health and the environment; and gather adequate data to identify and select the appropriate remedial actions for those sites where apparent risks exceed acceptable limits. The remainder of this Section describes the 14 sites at which sampling and analyses were conducted during RI activities.

1.3.1 Old Landfill (LF01)

The Old Landfill site (LF01) is located adjacent to the Beaufort Sea at the northernmost boundary of the Barter Island installation (Figure 1-3). The landfill was operational between 1956 and 1978 and is approximately two to three acres in size. Historically, the landfill received all waste from the station and the nearby village of Kaktovik. It reportedly received household waste, human and animal waste, drums, and other maintenance wastes. Station personnel stated that compaction, grading, removal of drums, installation of a gravel cap, and a general cleanup of exposed waste was conducted in 1992. A seawall was constructed that currently prevents erosion of the landfill by coastal wave processes.

1.3.2 POL Catchment (LF03)

The POL Catchment site (LF03) is a rectangular shaped tundra area surrounded by a gravel berm. The site is located north of the module trains and directly east of the POL Tanks (Figure 1-3). The POL Catchment serves as a secondary containment unit for petroleum hydrocarbons released from the POL Tanks (ST17) as a result of tank leaks and/or fuel spills. The POL Tank area is a bulk fuel storage area for Arctic grade diesel fuels.

1.3.3 Current Landfill (LF04)

The Current Landfill site (LF04) is located north of the module trains, southwest of the Old Landfill (LF01), and west of the access road (Figure 1-3). The Current Landfill is approximately two acres in size and receives wastes generated by the station. It also received waste from the village of Kaktovik from 1978 to 1992 when the village constructed its own landfill. It reportedly received household waste, human and animal waste, drums, and other maintenance waste. The disposal of wastes at this site by station personnel is in accordance with appropriate regulations; the use of the site by Kaktovik community residents was uncontrolled. The northern portion of the landfill was investigated because it appeared that uncontrolled disposal had occurred in that area.

1.3.4 Contaminated Ditch (SD08)

The Contaminated Ditch site (SD08) is located approximately 100 yards northeast of the main facility structures (Figure 1-3). The Contaminated Ditch is a large, deep, naturally eroded gully running to the north and discharging to the Beaufort Sea. The ditch was suspected of containing petroleum hydrocarbons from a ruptured fuel line near the warehouse area north of the module trains. Station personnel reported that a general cleanup had been conducted at the ditch and exposed metal debris removed.

1.3.5 Old Runway Dump (LF12)

The Old Runway Dump (LF12) site is located on the northeast corner of Barter Island, east of the runway (Figure 1-3). It is a two-acre area suspected of receiving wastes generated during the construction of the station and for some short period thereafter. The site may have received construction debris, old vehicles, drums, and other waste generated during this period. The landfill has been closed since 1957 and was reportedly cleaned up between 1979 and 1980.

1.3.6 Heated Storage Building (SS13)

The Heated Storage Building site (SS13) is located southeast of the module trains and the power house (Figure 1-3). The floor drains within the building discharged directly to tundra areas below. The drains may have received oils and other waste automotive fluids. The floor drains were sealed in July 1993 by the Air Force to prevent future release of contaminants from the Heated Storage Building to the tundra.

1.3.7 Garage (SS14)

The Garage (SS14) site is located south of the module trains and north of the Heated Storage Building (Figure 1-3). The floor drains in this building discharged directly to the tundra beneath the structures. The drains may have received vehicle maintenance waste. The floor drains were sealed in July 1993 by the Air Force to prevent future release of contaminants from the Garage to the tundra.

1.3.8 Weather Station Building (SS15)

The Weather Station Building site (SS15) is located northwest of the module trains and southwest of the Current Landfill (LF04) (Figure 1-3). A 1,200-gallon, aboveground diesel fuel storage tank is located at the northeast corner of the building. The diesel tank has leaked over the years and a stained area was observed just below the tank fittings.

1.3.9 White Alice Facility (SS16)

The White Alice Facility (SS16) site is located 1,600 feet south of the module trains (Figure 1-3). Currently inactive, it previously was a transmission and receiving unit for the station. The site consists of a radio relay building and two large White Alice "billboards" that look like outdoor

movie screens. It was suspected that dielectric fluids containing polychlorinated biphenyls (PCBs) were discharged to the surface soils in small quantities during maintenance of facility equipment.

1.3.10 POL Tanks (ST17)

The POL Tanks (ST17) site, an active bulk fuel storage area for Arctic grade diesel fuels, is located north of the module trains and south of the sewage lagoon (Figure 1-3). The POL Tanks were investigated as a possible source area of the POL Catchment (LF03). The site consists of six large, approximately 200,000 gallon, aboveground tanks and associated piping and pump house.

1.3.11 Fuel Tanks (ST18)

The Fuel Tanks (ST18) site is located approximately 300 feet east of module train B (Figure 1-3), and consists of six 10,000 gallon aboveground fuel tanks. Four tanks containing motor vehicle gasoline (MOGAS) and two tanks containing diesel are being investigated as a potential area of a past fuel spills and/or leaks. The site is active and is used for fueling installation vehicles and heavy equipment.

1.3.12 Old Dump Site (LF19)

The Old Dump Site (LF19) site consists of several acres of mostly tundra located northeast of the module trains and east of the Contaminated Ditch (Figure 1-3). The village of Kaktovik was located at this site from 1952 to 1964. Contaminants were suspected to be present at this site due to drum storage leaks and spills and previous waste management practices.

1.3.13 Bladder Diesel Spill (SS20)

The Bladder Diesel Spill (SS20) site is located west of module train B (Figure 1-3). The Bladder Diesel Spill area was historically a storage area for Arctic grade diesel fuels. Site personnel indicated the possibility of a past fuel spill in this area (Air Force 1993e) so, although the bladder diesel tank has been removed, the area adjacent to the tank site was suspected of containing petroleum hydrocarbons.

1.3.14 JP-4 Spill (SS21)

The JP-4 Spill site (SS21) is located approximately 1,300 feet east of the main facility (Figure 1-3). This reported fuel spill was from a cut in an approximately six-inch diameter JP-4 fuel line approximately 100 yards below the JP-4 fuel tank. The fuel line runs from the JP-4 tank to the hangar area. Site personnel indicated that a village diesel spill occurred upgradient and some of the product from the village spill may have migrated over this site.

1.4 APPROACH TO HUMAN HEALTH RISK ASSESSMENT

The Barter Island DEW Line installation presents a unique challenge in development of a human health risk assessment. Many of the conventional assumptions applied in risk assessments do not apply to the North Slope of Alaska. Barter Island is remote and sparsely populated. Native residents, largely Inupiat, follow a lifestyle that includes a significant subsistence component; much of their food consists of mammals (whales and caribou), aquatic life (Arctic char), and birds (ptarmigan, duck) that are abundant in this area of the Arctic. The climate is generally harsh, and the soil and surface water are frozen for approximately nine months of the year.

The general approach to the human health risk assessment is to quantify the excess lifetime cancer risk or the noncancer hazard for the site contaminants detected at each of the 14 sites at the installation. The maximum concentration of each chemical detected is used instead of an arithmetic mean or 95th percentile upper confidence limit (UCL) because contamination was detected infrequently and found to be generally of low concentration. Incorporating nondetects into the calculation of an average or UCL when the frequency of positive detects is low tends to yield low and unreliable estimates of contamination. Use of the maximum concentration yields a conservative estimate of risk or hazard.

To the extent possible, site-specific information is incorporated into the development of the exposure assumptions. The harsh climate naturally serves to limit exposure to contaminated soil, sediment, and surface water.

Excess lifetime cancer risk and noncancer hazard are calculated for the soil or sediment ingestion and water ingestion pathways. Other pathways were eliminated from consideration as described in Section 2.2, the Human Health Risk Exposure Assessment.

1.5 APPROACH TO ECOLOGICAL RISK ASSESSMENT

The objective of the ERA is to estimate potential impacts to aquatic and terrestrial plants and animals at the Barter Island radar installation. The MITRE guidance (1990) suggests that ERAs should "estimate the potential for occurrence of adverse effects that are manifested as changes in the diversity, health and behavior" of ecosystems. MITRE proposes that this can be accomplished by:

- Estimating the health risk to individual species,
- Evaluating the health of the community of exposed species, and
- Determining the potential adverse effects of contamination over several life cycles of the species under study.

Because this is a screening level assessment, the scope of this ERA is limited to the first task: estimating the health risk to individual species. If a potential health risk to individual species is identified, further work may be recommended to evaluate the community and life cycle effects. It is important to note that the health risk to an individual species is different from the health risk to an individual within a species. The former refers to population level biology, where the

individual is not considered a relevant endpoint. The latter assesses the risks to an individual. In this assessment, the individual is considered only in the case of threatened or endangered species.

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2.0 BASELINE HUMAN HEALTH RISK ASSESSMENT

The purpose of the baseline human health risk assessment for the Barter Island DEW Line installation is to provide a basis for developing a risk management plan, including remedial action alternatives based on data from the RI/FS. The RI at the Barter Island installation included investigation of 14 sites. The risk assessment develops numerical estimates of cancer risk and noncancer hazard for each site where sufficient information is available. Where information is not adequate to quantify noncancer hazard or cancer risk for a given COC, a qualitative discussion of the toxicity of that COC is provided in the Toxicity Profiles (Appendix B).

The risk assessment addresses issues unique to this location as described in the introduction. It follows the conventional approach, however, in that it is comprised of five sections:

- Identification of COCs - in which the chemicals detected in environmental samples are compared to regulatory benchmark concentrations and concentrations considered to be applicable or relevant and appropriate requirements (ARAR),
- Exposure assessment - in which the frequency, duration, and magnitude of potential exposures to the COCs are estimated,
- Toxicity assessment - in which the toxicology of the COCs is assessed,
- Risk characterization - in which the potential for adverse health effects in humans as a result of exposure to the COCs is quantified (as appropriate) and discussed, and
- Uncertainty assessment - in which the general sources of uncertainty in the assessment and the site-specific sources of uncertainty are discussed.

2.1 IDENTIFICATION OF CONTAMINANTS OF CONCERN

Chemicals of potential concern to human health were selected for each site at the Barter Island facility based on comparison of chemical concentrations to risk-based screening level (RBSL), naturally-occurring background concentrations, ARARs, and safe levels of essential human nutrients (e.g., calcium, magnesium, sodium and potassium).

This section discusses the evaluation of data prior to screening (2.1.1), describes and presents equations for calculating RBSLs (2.1.2), identifies chemicals that are essential human nutrients (2.1.3), describes the collection of background samples (2.1.4) and then discusses the selection of COCs (2.1.5). The last segment (2.1.6) identifies the COCs at each site.

2.1.1 Evaluation of Analytical Data

Before screening for chemicals of potential concern, the results of the RI sampling program were sorted by medium (e.g., soil, sediment, and surface water) and reviewed for quality. The review

included an evaluation of the analytical methods used, the sample quantitation limits, qualified data, and a comparison to background levels and blanks. Chemicals present in samples and in blanks at similar concentrations (i.e., laboratory and field contaminants) were not included in the detailed analysis.

Analytical data were reviewed for completeness, comparability, representativeness, precision, and accuracy. In addition, data validation qualifiers were considered in assessing the quality of the data. As outlined in the Risk Assessment Guidance for Superfund (EPA 1989a), samples qualified with an "R" were considered unsuitable for use and the samples eliminated from the data set. "J" values were included in the risk assessment at the reported concentration, e.g., a "12J" on the lab report would have been incorporated as a "12."

2.1.2 Risk-Based Screening Levels

An RBSL is a chemical concentration in a particular medium expected to yield a given risk or hazard quotient (e.g., 10^{-7} risk; 0.1 hazard quotient) under a given set of conditions. For Barter Island, the RBSLs were calculated for soil based on reasonable maximum exposure (RME) parameters for DEW Line workers and native northern adults and children. In developing the RBSLs, the most recent toxicity factors available from the Integrated Risk Information System (IRIS) and the Health Effects Assessment Summary Tables (HEAST) were used. IRIS and HEAST are databases of toxicity information for human health risk assessment maintained by the Environmental Criterion Assessment Office of the EPA. The information presented on IRIS represents the Agency's consensus regarding the toxicity of chemicals released to the environment.

2.1.2.1 Formulae for Calculating RBSLs. The RBSL concentrations were derived using EPA Region 10 guidance (1991a). The equations presented by EPA (1991a) are also presented in the Risk Assessment Guidance for Superfund Volume I, Part B (EPA 1991b). Exposure assessment and risk characterization algorithms for human health risk assessments use site-specific contaminant concentration data, factors describing exposure, and toxicity dose-response values (e.g., reference doses or carcinogen slope factors). These risk assessment algorithms are solved for the concentration term to derive the RBSL for soil and ground or surface water. The algorithms are summarized as follows:

$$\text{Risk} = C \times \left(\frac{\text{CR} \times \text{EFD}}{\text{BW} \times \text{AT}} \right) \times \text{SF} \quad \text{or} \quad \text{or HQ} = C \times \left(\frac{\text{CR} \times \text{EFD}}{\text{BW} \times \text{AT}} \right) / \text{RfD} \quad \text{EQUATION 1, 2}$$

C = Concentration

CR = Contact Rate

EFD = Exposure Frequency and Duration

BW = Body Weight

AT = Average Time

SF = Slope Factor

RfD = Reference Dose

HQ = Hazard Quotient.

Risk-based screening levels are calculated using a specified target cancer risk or HQ with the toxicity and exposure factors. The EPA Region 10 guidance (1991a) recommends that a 1×10^{-7} target risk and a target noncancer hazard quotient of 0.1 be used for soil and a 1×10^{-6} risk and 0.1 HQ be used for ground or surface water. The lower target cancer risk is used for screening

soil because additional pathways, such as dermal contact and inhalation, are not accounted for by the calculations (EPA 1991a).

Equations (1) and (2) shown above are rearranged to solve for the concentration term (i.e., the RBSL):

$$C = \text{Risk} / \left(\left(\frac{CR \times EFD}{BW \times AT} \right) \times SF \right) \quad \text{or} \quad C = HQ / \left(\left(\frac{CR \times EFD}{BW \times AT} \right) / RfD \right) \quad \text{EQUATION 3, 4}$$

Surface Water Ingestion Equations. Using standard default exposure factors (EPA 1989b) for water ingestion, the equation for cancer risk from drinking water ingestion becomes:

$$\text{Risk} = C (\mu\text{g/L}) \times 0.001 \text{ mg}/\mu\text{g} \times \left(\frac{2 \text{ L/day} \times 350 \text{ day/year} \times 30 \text{ year}}{70 \text{ kg} \times 70 \text{ year} \times 365 \text{ day/year}} \right) \times SF_o \quad \text{EQUATION 5}$$

This can be rearranged to solve for risk-based screening level, for example with a target cancer risk of 10^{-6} :

$$C (\mu\text{g/L}) = 10^{-6} \times 1,000 \mu\text{g}/\text{mg} / \left[\left(\frac{2 \text{ L/day} \times 350 \text{ day/year} \times 30 \text{ year}}{70 \text{ kg} \times 70 \text{ year} \times 365 \text{ day/year}} \right) \times SF_o \right] \quad \text{EQUATION 6}$$

For non-carcinogens, the equation for HQ for drinking water ingestion is:

$$HQ = C (\mu\text{g/L}) \times 0.001 \text{ mg}/\mu\text{g} \times \left(\frac{2 \text{ L/day} \times 350 \text{ day/year} \times 30 \text{ year}}{70 \text{ kg} \times 70 \text{ year} \times 365 \text{ day/year}} \right) / RfD_o \quad \text{EQUATION 7}$$

The equation for concentration representing HQ of 1 from ingestion is:

$$C (\mu\text{g/L}) = 1 \times 1,000 \mu\text{g}/\text{mg} / \left[\left(\frac{2 \text{ L/day} \times 350 \text{ day/year} \times 30 \text{ year}}{70 \text{ kg} \times 70 \text{ year} \times 365 \text{ day/year}} \right) / RfD_o \right] \quad \text{EQUATION 8}$$

Soil Ingestion Equations. The equation for calculating carcinogenic risk from soil ingestion, combining child and adult exposure, is as follows:

$$\text{Risk} = C (\text{mg}/\text{kg}) \times 0.000001 \text{ kg}/\text{mg} \times \quad \text{EQUATION 9}$$

$$\left[\left(\frac{200_c \text{ mg/day} \times 350_c \text{ day/year} \times 6 \text{ year}}{15_c \text{ kg} \times 365 \text{ day/year}} \right) + \left(\frac{100_a \text{ mg/day} \times 350_a \text{ day/year} \times 24 \text{ year}}{70_a \text{ kg} \times 365 \text{ day/year}} \right) \right] / 70 \text{ year} \times SF_o$$

This can be rearranged to solve for concentration at a target cancer risk of 10^{-7} :

$$C \text{ (mg/kg)} = 10^{-7} \times 1,000,000 \text{ kg/mg/}$$

EQUATION 10

$$\left[\left(\frac{200_c \text{ mg/day} \times 350_c \text{ day/year} \times 6 \text{ year}}{15_c \text{ kg} \times 365 \text{ day/year}} \right) + \left(\frac{100_a \text{ mg/day} \times 350_a \text{ day/year} \times 24 \text{ year}}{70_a \text{ kg} \times 365 \text{ day/year}} \right) \right] / 70 \text{ year} \times SF_o$$

For non-carcinogens in soil, HQ is calculated:

$$HQ = C \text{ (mg/kg)} \times 0.000001 \text{ kg/mg} \times$$

EQUATION 11

$$\left[\left(\frac{200_c \text{ mg/day} \times 350_c \text{ day/year} \times 6 \text{ year}}{15_c \text{ kg} \times 365 \text{ day/year}} \right) + \left(\frac{100_a \text{ mg/day} \times 350_a \text{ day/year} \times 24 \text{ year}}{70_a \text{ kg} \times 365 \text{ day/year}} \right) / 30 \text{ year} \right] / RfD_o$$

For non-carcinogens, a concentration representing HQ of 0.1 is calculated by:

$$C \text{ (mg/kg)} = 0.1 \times 1,000,000 \text{ mg/kg} /$$

EQUATION 12

$$\left[\left(\left(\frac{200_c \text{ mg/day} \times 350_c \text{ day/year} \times 6 \text{ year}}{15_c \text{ kg} \times 365 \text{ day/year}} \right) + \left(\frac{100_a \text{ mg/day} \times 350_a \text{ day/year} \times 24 \text{ year}}{15_a \text{ kg} \times 365 \text{ day/year}} \right) \right) / 30 \text{ year} \right] / RfD_o$$

2.1.3 Screening of Chemicals by Comparing Maximum Detected Concentrations of Essential Human Nutrients

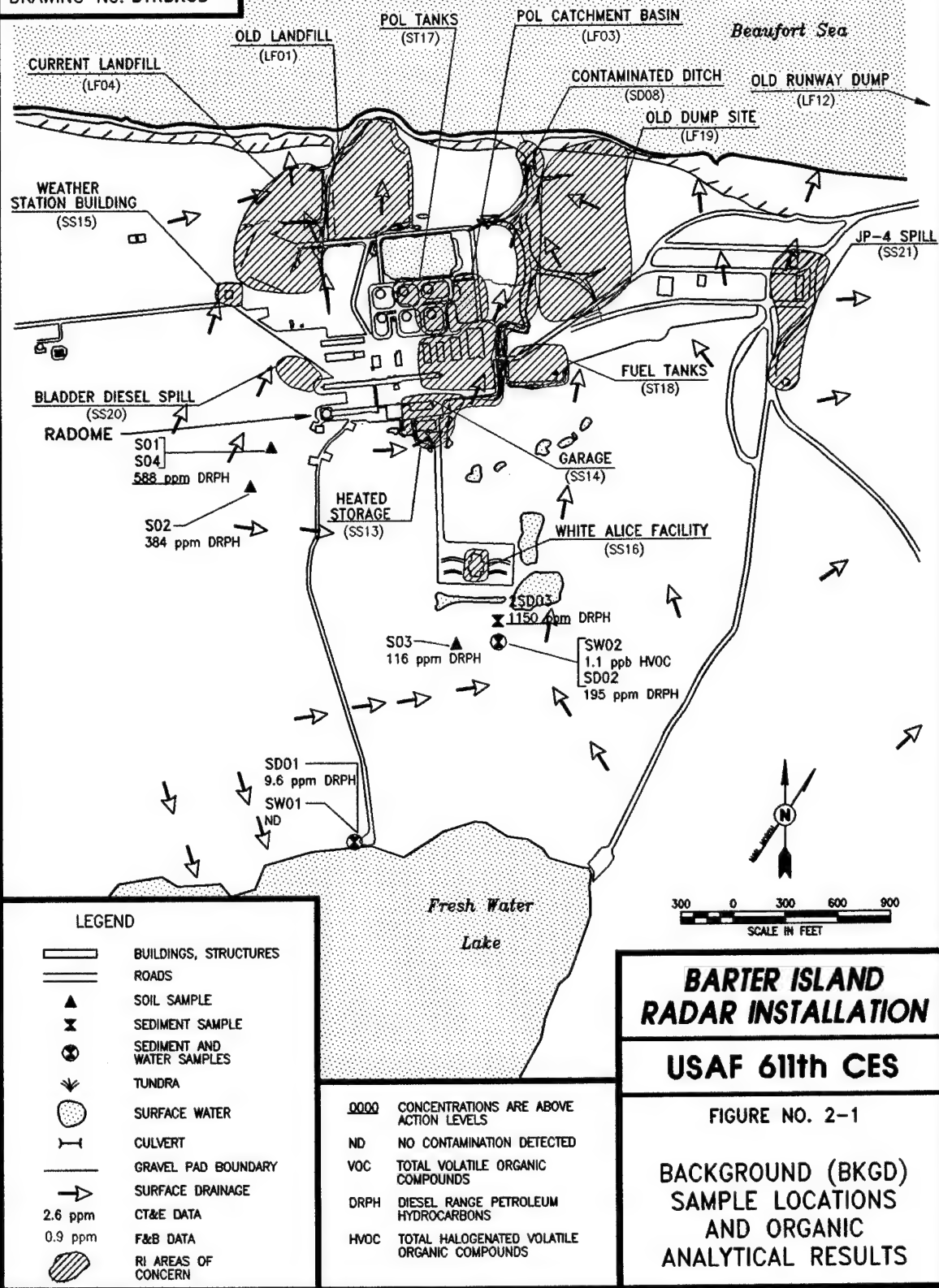
Based on EPA Region 10 guidance (1991a), calcium, magnesium, potassium, iron, and sodium are considered to be essential human nutrients and were eliminated from the human health risk assessment at the screening stage. These chemicals are often detected but are not toxic in humans except at extremely high doses. No quantitative toxicity information is available for these elements from EPA sources.

2.1.4 Concentrations of Organic and Inorganic Constituents in Background Samples

A total of eight samples were collected upgradient of the radar installation to determine the concentrations of naturally occurring organic and inorganic constituents in soil/sediment and surface water (Figure 2-1). Although some naturally occurring compounds were detected in some of the soil/sediment background samples in the Diesel Range Petroleum Hydrocarbons (DRPH) analyses, the organic concentration in background samples is assumed to be non-detect. This conservative approach was used because it is not possible to determine what degree, if any, the DRPH detected in site samples were naturally occurring compounds.

In order to obtain a representative range of inorganic (metal) concentrations in soil/sediments and surface waters of the North Slope, 44 samples (29 soil/sediment and 15 water) from seven North Slope radar installations were collected to establish the naturally occurring background concentrations. The seven installation include Barter Island, Bullen Point, Oliktok Point, Point

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Lonely, Point Barrow, Point Lay, and Wainwright. Approximately four soil/sediment and two surface water background samples were collected and analyzed for metals at each of the seven radar installations to determine the background concentrations of inorganic (metal) analytes across similar coastal environments of the North Slope.

2.1.5 Selection of Chemicals of Concern

Soil and Sediment. The maximum concentrations of the chemicals detected in soil or sediment samples at Barter Island and not considered to be essential human nutrients were compared, on a site-by-site basis, to the corresponding RBSLs, background concentrations, and where available, federal or state ARARs. A chemical was selected as a COC for soil and sediment if the maximum concentration at which the chemical was measured exceeded the corresponding background concentration and the RBSL (based either on cancer risk or noncancer hazard) or an ARAR (Table 2-1). Thus, for example, the maximum concentration of DRPH at the Contaminated Ditch (SD08), 2,260 mg/kg, exceeds the background range and the state ARAR of 500 mg/kg. Therefore, DRPH is selected as a COC for the soils at the Contaminated Ditch.

The COCs at each site were compared to background concentrations, RBSLs and ARARs in Table 2-1. The chemicals retained as COCs exceed background concentrations, and either the RBSL or ARAR. The COCs selected at each site are discussed in Sections 2.1.5.1 to 2.1.5.14.

Surface Water. The maximum concentrations of the chemicals detected in surface water samples at Barter Island were compared, on a site-by-site basis, to the corresponding RBSLs, background concentrations, and where available, federal or state ARARs. A chemical was selected as a COC for surface water if the maximum concentration at which the chemical was measured exceeded the corresponding background concentration and the RBSL (based either on cancer risk or noncancer hazard) or an ARAR (Table 2-1). Thus, for example, the maximum concentration of benzene at the POL Catchment (LF03), 2.7 µg/L, exceeds the background concentration of <1 µg/L (not detected) and the RBSL based on cancer risk of 0.617 µg/L. Therefore, benzene was selected as a COC for the surface water at the POL Catchment.

The COCs at each site were compared to background concentrations, RBSLs and ARARs in Table 2-1. The chemicals retained as COCs exceed background concentrations, and either the RBSL or ARAR.

Chemicals Without RBSLs or ARARs. For several chemicals detected in soil and water samples from the Barter Island installation, RfDs or slope factors are not available for calculating RBSLs and ARARs have not been established. Thus, no screening level concentration is available with which to select COCs. These chemicals are largely components of petroleum products and include the substituted benzenes and substituted toluenes.

Exclusion of these contaminants from the quantitative evaluation of risk is not anticipated to result in a significant underestimation of risk. They are considered to be less toxic than other chemicals expected to contribute most of the noncancer hazard or cancer risk at a given site, e.g., petroleum hydrocarbons, benzene, and tetrachloroethane.

TABLE 2-1. IDENTIFICATION OF CONTAMINANTS OF CONCERN: COMPARISON OF MAXIMUM CONCENTRATIONS TO RISK-BASED SCREENING LEVELS, ARARS, AND BACKGROUND EVALUATION OF CHEMICALS DETECTED

SITE	MATRIX	CHEMICAL DETECTED	MAXIMUM CONCENTRATION	UNITS	BACKGROUND RANGE	RBSL		ARAR	CHEMICAL OF CONCERN
						CANCER	NON-CANCER		
Old Landfill (LF01)	Soil	RRPH	530	mg/kg	<480	--	--	2,000 ^a	NO
		1,4-Dichlorobenzene	0.044	mg/kg	<0.020-<0.500	2.67	--	--	NO
		Naphthalene	0.105	mg/kg	<0.025-<3.50	--	100	--	NO
		1,2,4-Trichlorobenzene	0.046	mg/kg	<0.025-<0.500	--	270	--	NO
		1,2,4-Trimethylbenzene	0.041	mg/kg	<0.025-<0.500	--	--	--	NO
		Xylenes	0.033	mg/kg	<0.040-<1.000	--	54,000	--	NO
		Aluminum	3,600	mg/kg	1,500-25,000	--	--	--	NO
		Barium	76	mg/kg	27-390	--	1,890	--	NO
		Calcium	10,600	mg/kg	360-59,000	--	--	--	NO
		Chromium	6.6	mg/kg	<4.3-47	--	135	--	NO
		Copper	14	mg/kg	<2.7-45	--	999	--	NO
		Iron	12,000	mg/kg	5,400-35,000	--	--	--	NO
		Lead	29	mg/kg	<5.1-22	--	--	500 ^b	NO
		Magnesium	2,700	mg/kg	360-7,400	--	--	--	NO
		Manganese	170	mg/kg	25-290	--	3,780	--	NO
		Nickel	8.6	mg/kg	4.2-46	--	540	--	NO
		Potassium	610	mg/kg	<300-2,200	--	--	--	NO
		Sodium	160	mg/kg	<160-680	--	--	--	NO
		Vanadium	11	mg/kg	6.3-59	--	189	--	NO
		Zinc	55	mg/kg	9.2-95	--	8,100	--	NO
	Water	p-Isopropyltoluene	1.7	µg/L	<1	--	--	--	NO
		Toluene	56	µg/L	<1	--	96.5	1,000 ^c	NO
		Aluminum	180	µg/L	<100-350	--	--	--	NO

TABLE 2-1. IDENTIFICATION OF CONTAMINANTS OF CONCERN: COMPARISON OF MAXIMUM CONCENTRATIONS TO RISK-BASED SCREENING LEVELS, ARARS, AND BACKGROUND EVALUATION OF CHEMICALS DETECTED (CONTINUED)

SITE	MATRIX	CHEMICAL DETECTED	MAXIMUM CONCENTRATION	UNITS	BACKGROUND RANGE	RBSL		ARAR	CHEMICAL OF CONCERN
						CANCER	NON-CANCER		
Old Landfill (LF01) (Continued)	Water (Continued)	Barium	120	µg/L	<50-93	--	256	2,000 ^d	NO
		Calcium	190,000	µg/L	4,100-88,000	--	--	--	NO
		Iron	15,000	µg/L	<100-2,800	--	--	--	NO
		Magnesium	78,000	µg/L	<5,000-54,000	--	--	--	NO
		Manganese	1,500	µg/L	<50-510	--	18.3	--	YES
		Potassium	110,000	µg/L	<5,000	--	--	--	NO
		Sodium	440,000	µg/L	8,400-450,000	--	--	--	NO
		DRPH	28,600	mg/kg	9.55-1,150	--	--	500 ^a	YES
POL Catchment (LF03)	Soil	GRPH	78.5	mg/kg	<0.400-<9	--	--	100 ^a	NO
		RRPH	180	mg/kg	<480	--	--	2,000	NO
		n-Butylbenzene	3.70	mg/kg	<0.025-<0.500	--	--	--	NO
		sec-Butylbenzene	1.62	mg/kg	<0.025-<0.500	--	--	--	NO
		Ethylbenzene	0.982	mg/kg	<0.020-<0.500	--	2,700	--	NO
		Isopropylbenzene	0.409	mg/kg	<0.025-<0.500	--	--	--	NO
		p-Isopropyltoluene	1.69	mg/kg	<0.025-<0.500	--	--	--	NO
		Methylene Chloride	0.225	mg/kg	<.020-<0.500	8.53	1,620	--	NO
		2-Methylnaphthalene	3.44	mg/kg	<0.230-<3.50	--	--	--	NO
		n-Propylbenzene	0.920	mg/kg	<0.025-<0.500	--	--	--	NO
		Tetrachloroethane	5.42	mg/kg	<0.020-<0.500	1.23	270	--	YES
		Toluene	0.116	mg/kg	<0.020-<0.500	--	5,400	--	NO
		1,2,4-Trimethylbenzene	0.638	mg/kg	<0.025-<0.500	--	--	--	NO
		1,3,5-Trimethylbenzene	0.536	mg/kg	<0.025-<0.500	--	--	--	NO
		Xylenes	5.30	mg/kg	<0.040-<1,000	--	54,000	--	NO

TABLE 2-1. IDENTIFICATION OF CONTAMINANTS OF CONCERN: COMPARISON OF MAXIMUM CONCENTRATIONS TO RISK-BASED SCREENING LEVELS, ARARS, AND BACKGROUND EVALUATION OF CHEMICALS DETECTED (CONTINUED)

SITE	MATRIX	CHEMICAL DETECTED	MAXIMUM CONCENTRATION	UNITS	BACKGROUND RANGE	RBSL		ARAR	CHEMICAL OF CONCERN
						CANCER	NON-CANCER		
POL Catchment (LF03) (Continued)	Water	DRPH	1770	µg/L	<200	--	292	--	YES
		GRPH	367	µg/L	<20	50	730	--	YES
		Benzene	2.7	µg/L	<1	0.617	--	5 ^e	YES
		n-Butylbenzene	2.4	µg/L	<1	--	--	--	NO
		sec-Butylbenzene	1.0	µg/L	<1	--	--	--	NO
		Ethylbenzene	19	µg/L	<1	--	158	700 ^c	NO
		Isopropylbenzene	2.9	µg/L	<1	--	--	--	NO
		p-Isopropyltoluene	2.1	µg/L	<1	--	--	--	NO
		Naphthalene	35	µg/L	<1	--	--	--	NO
		n-Propylbenzene	3.6	µg/L	<1	--	--	--	NO
		1,2,4-Trimethylbenzene	19	µg/L	<1	--	--	--	NO
		1,3,5-Trimethylbenzene	13	µg/L	<1	--	--	--	NO
		Xylenes	38.4	µg/L	<2	--	7,300	10,000 ^c	NO
		DRPH	60	mg/kg	9.55-1,150	--	--	500	NO
Current Landfill (LF04)	Soil	GRPH	8	mg/kg	<0.400-<9	--	--	100	NO
		Toluene	0.026	mg/kg	<0.020-<0.500	--	5,400	--	NO
		Xylenes	0.2	mg/kg	<0.040-<1.000	--	54,000	--	NO
		Aluminum	7,500	mg/kg	1,500-25,000	--	--	--	NO
		Barium	120	mg/kg	27-390	--	1,890	--	NO
		Calcium	10,000	mg/kg	360-59,000	--	--	--	NO
		Chromium	3.4	mg/kg	<4.3-47	--	135	--	NO
		Copper	18	mg/kg	<2.7-45	--	999	--	NO
		Iron	17,000	mg/kg	5,400-35,000	--	--	--	NO

TABLE 2-1. IDENTIFICATION OF CONTAMINANTS OF CONCERN: COMPARISON OF MAXIMUM CONCENTRATIONS TO RISK-BASED SCREENING LEVELS, ARARS, AND BACKGROUND EVALUATION OF CHEMICALS DETECTED (CONTINUED)

SITE	MATRIX	CHEMICAL DETECTED	MAXIMUM CONCENTRATION	UNITS	BACKGROUND RANGE	RBSL		ARAR	CHEMICAL OF CONCERN
						CANCER	NON-CANCER		
Current Landfill (LF04) (Continued)	Soil (Continued)	Lead	6.1	mg/kg	<5.1-22	--	--	500	NO
		Magnesium	3,800	mg/kg	360-7,400	--	--	--	NO
		Manganese	380	mg/kg	25-290	--	3,780	--	NO
		Nickel	16	mg/kg	4.2-46	--	540	--	NO
		Sodium	870	mg/kg	<160-680	--	--	--	NO
		Vanadium	21	mg/kg	6.3-59	--	189	--	NO
		Zinc	65	mg/kg	9.2-95	--	8,100	--	NO
	Water	Dichlorofluoromethane	3.7	µg/L	<1	--	--	--	NO
		p-Isopropyltoluene	1.1	µg/L	<1	--	--	--	NO
		Trichloroethene	36	µg/L	<1	--	--	5 ^e	YES
		Barium	150	µg/L	<50-93	--	256	2,000	NO
		Calcium	120,000	µg/L	4,100-88,000	--	--	--	NO
		Iron	21,000	µg/L	<100-28,000	--	--	--	NO
		Magnesium	41,000	µg/L	<5,000-54,000	--	--	--	NO
Contaminated Ditch (SD08)	Soil	Manganese	1,800	µg/L	<50-510	--	18.3	--	YES
		Potassium	12,000	µg/L	<5,000	--	--	--	NO
		Sodium	100,000	µg/L	8,200-450,000	--	--	--	NO
		DRPH	2,260	mg/kg	9.55-1,150	--	--	500	YES
		GRPH	171	mg/kg	<0.400-<9	--	--	100	YES
		Ethylbenzene	3.01	mg/kg	<0.020-<0.500	--	2,700	--	NO
		Xylenes	12.53	mg/kg	<0.040-<1.000	--	54,000	--	NO
		Aluminum	2,500	mg/kg	1,500-25,000	--	--	--	NO
		Barium	44	mg/kg	27-390	--	1,890	--	NO

TABLE 2-1. IDENTIFICATION OF CONTAMINANTS OF CONCERN: COMPARISON OF MAXIMUM CONCENTRATIONS TO RISK-BASED SCREENING LEVELS, ARARS, AND BACKGROUND EVALUATION OF CHEMICALS DETECTED (CONTINUED)

SITE	MATRIX	CHEMICAL DETECTED	MAXIMUM CONCENTRATION	UNITS	BACKGROUND RANGE	RBSL		ARAR	CHEMICAL OF CONCERN
						CANCER	NON-CANCER		
Contaminated Ditch (SD08) (Continued)	Soil (Continued)	Beryllium	3.2	mg/kg	<2.6-64	0.0149	135	--	YES
		Cadmium	2.8	mg/kg	<3.0-<36	--	27	--	NO
		Calcium	12,300	mg/kg	360-59,000	--	--	--	NO
		Chromium	5.3	mg/kg	<4.3-47	--	135	--	NO
		Cobalt	5.7	mg/kg	--	--	--	--	NO
		Copper	5.8	mg/kg	<2.7-45	--	999	--	NO
		Iron	12,000	mg/kg	5,400-35,000	--	--	--	NO
		Magnesium	2,600	mg/kg	360-7,400	--	--	--	NO
		Manganese	250	mg/kg	25-290	--	3,780	--	NO
		Nickel	7.6	mg/kg	4.2-46	--	540	--	NO
		Potassium	440	mg/kg	<300-2,200	--	--	--	NO
		Selenium	57	mg/kg	<7.8-<170	--	135	--	NO
		Sodium	80	mg/kg	<160-680	--	--	--	NO
		Vanadium	8.4	mg/kg	6.3-59	--	189	--	NO
		Zinc	27	mg/kg	9.2-95	--	8,100	--	NO
		cis-1,2-Dichloroethene	1.5	µg/L	<1	--	36.5	70 ^c	NO
		Aluminum	1,900	µg/L	<100-350	--	--	--	NO
	Water	Barium	110	µg/L	<50-93	--	256	2,000	NO
		Calcium	71,000	µg/L	4,100-88,000	--	--	--	NO
		Iron	4,200	µg/L	<100-2,800	--	--	--	NO
		Magnesium	31,000	µg/L	<5,000-54,000	--	--	--	NO
		Manganese	170	µg/L	<50-510	--	18.3	--	NO
		Sodium	120,000	µg/L	8,200-450,000	--	--	--	NO

TABLE 2-1. IDENTIFICATION OF CONTAMINANTS OF CONCERN: COMPARISON OF MAXIMUM CONCENTRATIONS TO RISK-BASED SCREENING LEVELS, ARARS, AND BACKGROUND EVALUATION OF CHEMICALS DETECTED (CONTINUED)

SITE	MATRIX	CHEMICAL DETECTED	MAXIMUM CONCENTRATION	UNITS	BACKGROUND RANGE	RBSL		ARAR	CHEMICAL OF CONCERN
						CANCER	NON-CANCER		
Old Runway Dump (LF12)	Soil	Aluminum	1,700	mg/kg	1,500-25,000	--	--	--	NO
		Barium	16	mg/kg	27-390	--	1,890	--	NO
		Calcium	2,200	mg/kg	360-59,000	--	--	--	NO
		Chromium	4.5	mg/kg	<4.3-47	--	135	--	NO
		Copper	7.3	mg/kg	<2.7-45	--	999	--	NO
		Iron	7,500	mg/kg	5,400-35,000	--	--	--	NO
		Magnesium	1,100	mg/kg	360-7,400	--	--	--	NO
		Manganese	67	mg/kg	25-290	--	3,780	--	NO
		Nickel	5.3	mg/kg	4.2-46	--	540	--	NO
		Sodium	410	mg/kg	<160-680	--	--	--	NO
		Vanadium	5.1	mg/kg	6.3-59	--	189	--	NO
		Zinc	16	mg/kg	9.2-95	--	8,100	--	NO
		DRPH	3,580	mg/kg	9.55-1,150	--	--	500	YES
		GRPH	423	mg/kg	<0.400-<9	--	--	100	YES
Heated Storage Building (SS13)	Soil	RRPH	2,400	mg/kg	<480	--	--	2,000	YES
		Ethylbenzene	0.283	mg/kg	<0.020-<0.500	--	2,700	--	NO
		Toluene	0.632	mg/kg	<0.020-<0.500	--	5,400	--	NO
		1,3,5-Trimethylbenzene	7.68	mg/kg	<0.025-<0.500	--	--	--	NO
		Xylenes	2.53	mg/kg	<0.040-<1.000	--	54,000	--	NO
		Aroclor 1254	2.72	mg/kg	<0.020-<0.100	--	0.54	10 ¹	YES
		Aluminum	3,000	mg/kg	1,500-25,000	--	--	--	NO
		Barium	43	mg/kg	27-390	--	1,890	--	NO
		Cadmium	2.7	mg/kg	<3.0-<36	--	27	--	NO

TABLE 2-1. IDENTIFICATION OF CONTAMINANTS OF CONCERN: COMPARISON OF MAXIMUM CONCENTRATIONS TO RISK-BASED SCREENING LEVELS, ARARS, AND BACKGROUND EVALUATION OF CHEMICALS DETECTED (CONTINUED)

SITE	MATRIX	CHEMICAL DETECTED	MAXIMUM CONCENTRATION	UNITS	BACKGROUND RANGE	RBSL		ARAR	CHEMICAL OF CONCERN
						CANCER	NON-CANCER		
Heated Storage Building (SS13) (Continued)	Soil (Continued)	Calcium	14,000	mg/kg	360-59,000	--	--	--	NO
		Chromium	17	mg/kg	<4.3-47	--	135	--	NO
		Copper	12	mg/kg	<2.7-45	--	999	--	NO
		Iron	9,600	mg/kg	5,400-35,000	--	--	--	NO
		Lead	33	mg/kg	<5.1-22	--	--	500	NO
		Magnesium	4,900	mg/kg	360-7,400	--	--	--	NO
		Manganese	96	mg/kg	25-290	--	3,780	--	NO
		Nickel	7.9	mg/kg	4.2-46	--	540	--	NO
		Potassium	580	mg/kg	<300-2,200	--	--	--	NO
		Sodium	60	mg/kg	<160-680	--	--	--	NO
		Vanadium	14	mg/kg	6.03-59	--	189	--	NO
		Zinc	500	mg/kg	9.2-95	--	8,100	--	NO
	Water	DRPH	5,760	µg/L	<200	--	292	--	YES
		GRPH	6.9	µg/L	<20	50	730	--	NO
		Benzene	6.9	µg/L	<1	0.617	--	5	YES
		Chloromethane	4.2	µg/L	<1	6.54	--	--	NO
		1,2-Dichloroethane	9.1	µg/L	1.3B-3.2B	0.934	--	5 ^e	NO ^a
		Ethylbenzene	3.2	µg/L	<1	--	158	700	NO
		Naphthalene	1.6	µg/L	<1	--	--	--	NO
		Tetrachloroethane	12	µg/L	<1	1.63	36.5	5 ^c	YES
		1,2,4-Trimethylbenzene	1.4	µg/L	<1	--	--	--	NO
		1,3,5-Trimethylbenzene	1.3	µg/L	<1	--	--	--	NO
		Xylenes	12.8	µg/L	<2	--	7,300	10,000	NO

TABLE 2-1. IDENTIFICATION OF CONTAMINANTS OF CONCERN: COMPARISON OF MAXIMUM CONCENTRATIONS TO RISK-BASED SCREENING LEVELS, ARARS, AND BACKGROUND EVALUATION OF CHEMICALS DETECTED (CONTINUED)

SITE	MATRIX	CHEMICAL DETECTED	MAXIMUM CONCENTRATION	UNITS	BACKGROUND RANGE	RBSL		ARAR	CHEMICAL OF CONCERN
						CANCER	NON-CANCER		
Heated Storage Building (SS13) (Continued)	Water (Continued)	Aluminum	210	µg/L	<100-350	--	--	--	NO
		Barium	82	µg/L	<50-93	--	256	2,000	NO
		Calcium	130,000	µg/L	4,100-88,000	--	--	--	NO
		Iron	12,000	µg/L	<100-2,800	--	--	--	NO
		Magnesium	46,000	µg/L	<5,000-54,000	--	--	--	NO
		Manganese	540	µg/L	<50-510	--	18.3	--	YES
		Potassium	17,000	µg/L	<5,000	--	--	--	NO
		Sodium	110,000	µg/L	8,200-450,000	--	--	--	NO
		DRPH	12,400	mg/kg	9.55-1,150	--	--	500	YES
		GRPH	700	mg/kg	<0.400-<9	--	--	100	YES
Garage (SS14)	Soil	RRPH	27,000	mg/kg	<480	--	--	2,000	YES
		Benzene	1.4	mg/kg	<0.020-<0.500	2.2	--	0.5 ^a	YES
		n-Butylbenzene	4.22	mg/kg	<0.025-<0.500	--	--	--	NO
		sec-Butylbenzene	1.28	mg/kg	<0.025-<0.500	--	--	--	NO
		tert-Butylbenzene	0.256	mg/kg	<0.025-<0.500	--	--	--	NO
		cis-1,2-Dichloroethene	0.069	mg/kg	<0.025-<0.500	--	270	--	NO
		Ethylbenzene	11	mg/kg	<0.020-<0.500	--	2,700	--	NO
		Isopropylbenzene	0.681	mg/kg	<0.025-<0.500	--	--	--	NO
		p-Isopropyltoluene	2.47	mg/kg	<0.025-<0.500	--	--	--	NO
		Naphthalene	46	mg/kg	<0.025-<3.50	--	100	--	NO
		n-Propylbenzene	1.17	mg/kg	<0.025-<0.500	--	--	--	NO

TABLE 2-1. IDENTIFICATION OF CONTAMINANTS OF CONCERN: COMPARISON OF MAXIMUM CONCENTRATIONS TO RISK-BASED SCREENING LEVELS, ARARS, AND BACKGROUND EVALUATION OF CHEMICALS DETECTED (CONTINUED)

SITE	MATRIX	CHEMICAL DETECTED	MAXIMUM CONCENTRATION	UNITS	BACKGROUND RANGE	RBSL		ARAR	CHEMICAL OF CONCERN
						CANCER	NON-CANCER		
Garage (SS14) (Continued)	Soil (Continued)	Tetrachloroethane	0.023	mg/kg	<0.020-<0.500	1.23	270	--	NO
		Toluene	2.3	mg/kg	<0.020-<0.500	--	5,400	--	NO
		1,2,4-Trimethylbenzene	14.7	mg/kg	<0.025-<0.500	--	--	--	NO
		1,3,5-Trimethylbenzene	9.32	mg/kg	<0.025-<0.500	--	--	--	NO
		Xylenes	47	mg/kg	<0.040-<1.000	--	54,000	--	NO
		Isophrone	1.49	mg/kg	<0.230-<3.50	67.4	5,400	--	NO
		2-Methylnaphthalene	14.5	mg/kg	<0.230-<3.50	--	--	--	NO
		Phenanthrene	4.79	mg/kg	<0.230-<3.50	--	--	--	NO
		Fluoranthene	2.28	mg/kg	<0.230-<3.50	--	1,080	--	NO
		bis(2-Ethylhexyl)Phthalate	4.6	mg/kg	<0.230-<3.50	4.57	540	--	YES
		Aluminum	2,100	mg/kg	1,500-25,000	--	--	--	NO
		Barium	29	mg/kg	27-390	--	1,890	--	NO
		Calcium	9,300	mg/kg	360-59,000	--	--	--	NO
		Chromium	53	mg/kg	<4.3-47	--	135	--	NO
		Copper	16	mg/kg	<2.7-45	--	999	--	NO
		Iron	6,500	mg/kg	5,400-35,000	--	--	--	NO
		Lead	231	mg/kg	<5.1-22	--	--	500	NO
		Magnesium	4,500	mg/kg	360-7,400	--	--	--	NO
		Manganese	63	mg/kg	25-290	--	3,780	--	NO
		Nickel	5.6	mg/kg	4.2-46	--	540	--	NO

TABLE 2-1. IDENTIFICATION OF CONTAMINANTS OF CONCERN: COMPARISON OF MAXIMUM CONCENTRATIONS TO RISK-BASED SCREENING LEVELS, ARARS, AND BACKGROUND EVALUATION OF CHEMICALS DETECTED (CONTINUED)

SITE	MATRIX	CHEMICAL DETECTED	MAXIMUM CONCENTRATION	UNITS	BACKGROUND RANGE	RBSL		ARAR	CHEMICAL OF CONCERN
						CANCER	NON-CANCER		
Garage (SS14) (Continued)	Soil (Continued)	Potassium	310	mg/kg	<300-2,200	--	--	--	NO
		Sodium	110	mg/kg	<160-680	--	--	--	NO
		Vanadium	7.5	mg/kg	6.3-59	--	189	--	NO
		Zinc	200	mg/kg	9.2-95	--	8,100	--	NO
	Water	Xylenes	8	µg/L	<2	--	7,300	10,000	NO
		Barium	74	µg/L	<50-93	--	256	2,000	NO
		Calcium	110,000	µg/L	4,100-88,000	--	--	--	NO
		Iron	6,000	µg/L	<100-2,800	--	--	--	NO
		Magnesium	34,000	µg/L	<5,000-54,000	--	--	--	NO
		Manganese	490	µg/L	<50-510	--	18.3	--	NO
		Potassium	9,100	µg/L	<5,000	--	--	--	NO
		Sodium	140,000	µg/L	8,200-450,000	--	--	--	NO
Weather Station Building (SS15)	Soil	DRPH	8,420	mg/kg	9.55-1,150	--	--	500	YES
		GRPH	1,020	mg/kg	<0.400-<9	--	--	100	YES
		Ethylbenzene	1.09	mg/kg	<0.020-<0.500	--	2,700	--	NO
		Naphthalene	0.083	mg/kg	<0.025-<3.50	--	100	--	NO
		Toluene	0.137	mg/kg	<0.020-<0.500	--	5,400	--	NO
		1,3,5-Trimethylbenzene	0.084	mg/kg	<0.025-<0.500	--	--	--	NO
		Xylenes	2.34	mg/kg	<0.040-<1,000	--	54,000	--	NO
White Alice Facility (SS16)	Soil	Aroclor 1254	52	mg/kg	<0.020-<0.100	--	0.54	10	YES

TABLE 2-1. IDENTIFICATION OF CONTAMINANTS OF CONCERN: COMPARISON OF MAXIMUM CONCENTRATIONS TO RISK-BASED SCREENING LEVELS, ARARS, AND BACKGROUND EVALUATION OF CHEMICALS DETECTED (CONTINUED)

SITE	MATRIX	CHEMICAL DETECTED	MAXIMUM CONCENTRATION	UNITS	BACKGROUND RANGE	RBSL		ARAR	CHEMICAL OF CONCERN
						CANCER	NON-CANCER		
POL Tanks (ST17)	Soil	DRPH	1,670	mg/kg	9.55-1,150	--	--	500	YES
		GRPH	295	mg/kg	<0.400-<9	--	--	100	YES
		Ethylbenzene	0.4	mg/kg	<0.020-<0.500	--	2,700	--	NO
		Toluene	1.0	mg/kg	<0.020-<0.500	--	5,400	--	NO
		1,2,4-Trimethylbenzene	0.511	mg/kg	<0.025-<0.500	--	--	--	NO
		Xylenes	0.6	mg/kg	<0.040-<1,000	--	54,000	--	NO
Fuel Tanks (ST18)	Soil	DRPH	490	mg/kg	9.55-1,150	--	--	500	NO
		GRPH	6	mg/kg	<0.400-<9	--	--	100	NO
		RRPH	190	mg/kg	<480	--	--	2,000	NO
		Benzene	0.1	mg/kg	<0.020-<0.500	2.2	--	0.5	NO
		Ethylbenzene	0.2	mg/kg	<0.020-<0.500	--	2,700	--	NO
		Toluene	0.1	mg/kg	<0.020-<0.500	--	5,400	--	NO
		1,2,4-Trimethylbenzene	0.128	mg/kg	<0.025-<0.500	--	--	--	NO
		1,3,5-Trimethylbenzene	0.048	mg/kg	<0.025-<0.500	--	--	--	NO
Old Dump Site (LF19)	Soil	Xylenes	0.4	mg/kg	<0.040-<1,000	--	54,000	--	NO
		DRPH	580	mg/kg	9.55-1,150	--	--	500	YES
		GRPH	22	mg/kg	<0.400-<9	--	--	100	NO
		RRPH	5,800	mg/kg	<480	--	--	2,000	YES
		Aluminum	3,500	mg/kg	1,500-25,000	--	--	--	NO
		Barium	31	mg/kg	27-390	--	1,890	--	NO
		Calcium	12,000	mg/kg	360-59,000	--	--	--	NO
		Chromium	7.4	mg/kg	<4.3-47	--	135	--	NO
		Copper	7.9	mg/kg	<2.7-45	--	999	--	NO

TABLE 2-1. IDENTIFICATION OF CONTAMINANTS OF CONCERN: COMPARISON OF MAXIMUM CONCENTRATIONS TO RISK-BASED SCREENING LEVELS, ARARS, AND BACKGROUND EVALUATION OF CHEMICALS DETECTED (CONTINUED)

SITE	MATRIX	CHEMICAL DETECTED	MAXIMUM CONCENTRATION	UNITS	BACKGROUND RANGE	RBSL		ARAR	CHEMICAL OF CONCERN
						CANCER	NON-CANCER		
Old Dump Site (LF19) (Continued)	Soil (Continued)	Iron	8,800	mg/kg	5,400-35,000	--	--	--	NO
		Magnesium	1,900	mg/kg	360-7,400	--	--	--	NO
		Manganese	160	mg/kg	25-290	--	3,780	--	NO
		Nickel	7.8	mg/kg	4.2-46	--	540	--	NO
		Potassium	410	mg/kg	<300-2,200	--	--	--	NO
		Sodium	70	mg/kg	<160-680	--	--	--	NO
		Vanadium	9.3	mg/kg	6.3-59	--	189	--	NO
		Zinc	21	mg/kg	9.2-95	--	8,100	--	NO
		Aluminum	160	µg/L	<100-350	--	--	--	NO
		Barium	140	µg/L	<50-93	--	256	2,000	NO
		Calcium	170,000	µg/L	4,100-88,000	--	--	--	NO
		Iron	18,000	µg/L	<100-2,800	--	--	--	NO
		Magnesium	75,000	µg/L	<5,000-54,000	--	--	--	NO
Bladder Diesel Spill (SS20)	Soil	Manganese	70	µg/L	<50-510	--	18.3	--	NO
		Potassium	8,700	µg/L	<5,000	--	--	--	NO
		Sodium	320,000	µg/L	8,200-450,000	--	--	--	NO
		Naphthalene	0.280	mg/kg	<0.025-<3.50	--	100	--	NO
		Toluene	0.097	mg/kg	<0.020-<0.500	--	5,400	--	NO
		Phenanthrene	1.96	mg/kg	<0.230-<3.50	--	--	--	NO
		Fluoranthene	1.70	mg/kg	<0.230-<3.50	--	1,080	--	NO

TABLE 2-1. IDENTIFICATION OF CONTAMINANTS OF CONCERN: COMPARISON OF MAXIMUM CONCENTRATIONS TO RISK-BASED SCREENING LEVELS, ARARS, AND BACKGROUND EVALUATION OF CHEMICALS DETECTED (CONTINUED)

SITE	MATRIX	CHEMICAL DETECTED	MAXIMUM CONCENTRATION	UNITS	BACKGROUND RANGE	RBSL		ARAR	CHEMICAL OF CONCERN
						CANCER	NON-CANCER		
Bladder Diesel Spill (SS20) (Continued)	Soil (Continued)	Pyrene	1.26	mg/kg	<0.230-<3.50	--	810	--	NO
		Aluminum	2,100	mg/kg	1,500-25,000	--	--	--	NO
		Barium	20	mg/kg	27-390	--	1,890	--	NO
		Calcium	33,000	mg/kg	360-59,000	--	--	--	NO
		Chromium	9.3	mg/kg	<4.3-47	--	135	--	NO
		Copper	6.6	mg/kg	<2.7-45	--	999	--	NO
		Iron	6,200	mg/kg	5,400-35,000	--	--	--	NO
		Lead	11	mg/kg	<5.1-22	--	--	500	NO
		Magnesium	17,000	mg/kg	360-7,400	--	--	--	NO
		Manganese	77	mg/kg	25-290	--	3,780	--	NO
		Nickel	7.7	mg/kg	4.2-46	--	540	--	NO
		Sodium	80	mg/kg	<160-680	--	--	--	NO
		Vanadium	6.7	mg/kg	6.3-59	--	189	--	NO
		Zinc	30	mg/kg	9.2-95	--	8,100	--	NO
		Naphthalene	2.7	µg/L	<1-<10	--	--	--	NO
JP-4 Spill (SS21)	Soil	DRPH	1,300	mg/kg	9.55-1,150	--	--	500	YES
		GRPH	300	mg/kg	<0.400-<9	--	--	100	YES
		Benzene	3.9	mg/kg	<0.020-<0.500	2.2	--	0.5	YES
		n-Butylbenzene	4.20	mg/kg	<0.025-<0.500	--	--	--	NO
		sec-Butylbenzene	1.94	mg/kg	<0.025-<0.500	--	--	--	NO
		tert-Butylbenzene	0.345	mg/kg	<0.025-<0.500	--	--	--	NO
		Ethylbenzene	11.8	mg/kg	<0.020-<0.500	--	2,700	--	NO
		Isopropylbenzene	2.25	mg/kg	<0.025-<0.500	--	--	--	NO

TABLE 2-1. IDENTIFICATION OF CONTAMINANTS OF CONCERN: COMPARISON OF MAXIMUM CONCENTRATIONS TO RISK-BASED SCREENING LEVELS, ARARS, AND BACKGROUND EVALUATION OF CHEMICALS DETECTED (CONTINUED)

SITE	MATRIX	CHEMICAL DETECTED	MAXIMUM CONCENTRATION	UNITS	BACKGROUND RANGE	RBSL		ARAR	CHEMICAL OF CONCERN
						CANCER	NON-CANCER		
JP-4 Spill (SS21)	Soil (Continued)	p-Isopropyltoluene	2.56	mg/kg	<0.025-<0.500	--	--	--	NO
		Naphthalene	5.54	mg/kg	<0.025-<3.50	--	100	--	NO
		n-Propylbenzene	3.26	mg/kg	<0.025-<0.500	--	--	--	NO
		Toluene	24.0	mg/kg	<0.020-<0.500	--	5,400	--	NO
		1,2,4-Trimethylbenzene	18.1	mg/kg	<0.025-<0.500	--	--	--	NO
		1,3,5-Trimethylbenzene	9.16	mg/kg	<0.025-<0.500	--	--	--	NO
		Xylenes	69.0	mg/kg	<0.040-<1.000	--	54,000	--	NO
		2-Methylnaphthalene	1.42	mg/kg	<0.230-<3.50	--	--	--	NO

a Target cleanup levels for DRPH, GRPH, and RRPH in soil are based on ADEC Non UST guidance and do not necessarily correspond to final site specific cleanup goals.
b EPA 1991c.
c MCL, 56 FR 3526 (30 January 1991).
d MCL, 56 FR 30266 (01 July 1991).
e MCL, 52 FR 25690 (08 July 1987).
f TSCA Cleanup Level.
g 1,2-Dichloroethane was detected in the majority of blank and background samples, therefore was not included as a COC.

Exposures to Lead. Exposures to lead may cause adverse noncarcinogenic health effects; however, EPA has not developed an RfD for this chemical. Lead concentrations in soil were compared to EPA's final action level for lead in soil of 500-1,000 mg/kg. It is estimated that exposure to soil containing 500 mg of lead per kilogram soil would yield blood lead levels below 10 µg/L (a blood lead level of concern) in roughly 99 percent of young children who are not also exposed to excessive lead paint hazards or heavily contaminated soils (EPA 1989d). Lead concentrations did not exceed 500 mg/kg at any sampling location at the Barter Island facility.

2.1.5.1 Old Landfill (LF01). No COCs were identified for the soil matrix at the Old Landfill (Figure 2-2) based on a comparison of the maximum concentrations of the detected chemicals to background levels, and the RBSL and/or ARAR concentrations (Table 2-1). Manganese was the only chemical identified as a COC for surface water (Table 2-1).

2.1.5.2 POL Catchment (LF03). DRPH and tetrachloroethane were identified as COCs for the soil matrix at the POL Catchment site (Figure 2-3). The maximum concentration of DRPH exceeded the ARAR for diesel contamination of soil (Table 2-1) (ADEC 1991). The maximum concentration of tetrachloroethane exceeded the carcinogen RBSL (Table 2-1).

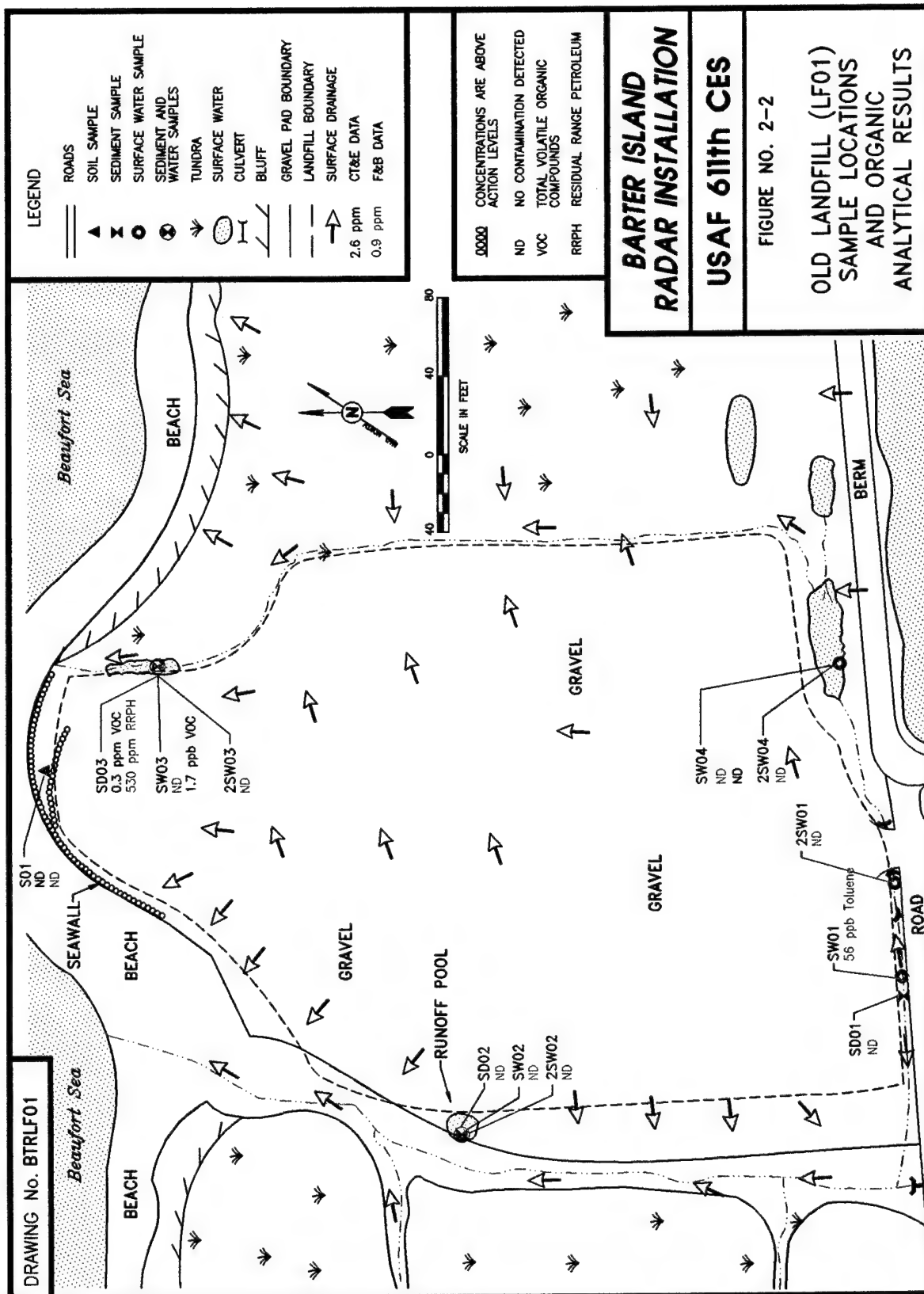
DRPH, GRPH, and benzene were identified as COCs for the surface water at the POL Catchment (Figure 2-3). The maximum concentration of DRPH exceeded the noncarcinogen RBSL for diesel contamination of surface water (Table 2-1). The maximum concentrations of GRPH and benzene exceeded their carcinogen RBSLs (Table 2-1).

2.1.5.3 Current Landfill (LF04). No COCs were identified for the soil matrix at the Current Landfill (Figure 2-4) based on a comparison of the maximum concentrations of the detected chemicals to background levels, and the RBSL and/or ARAR concentrations (Table 2-1).

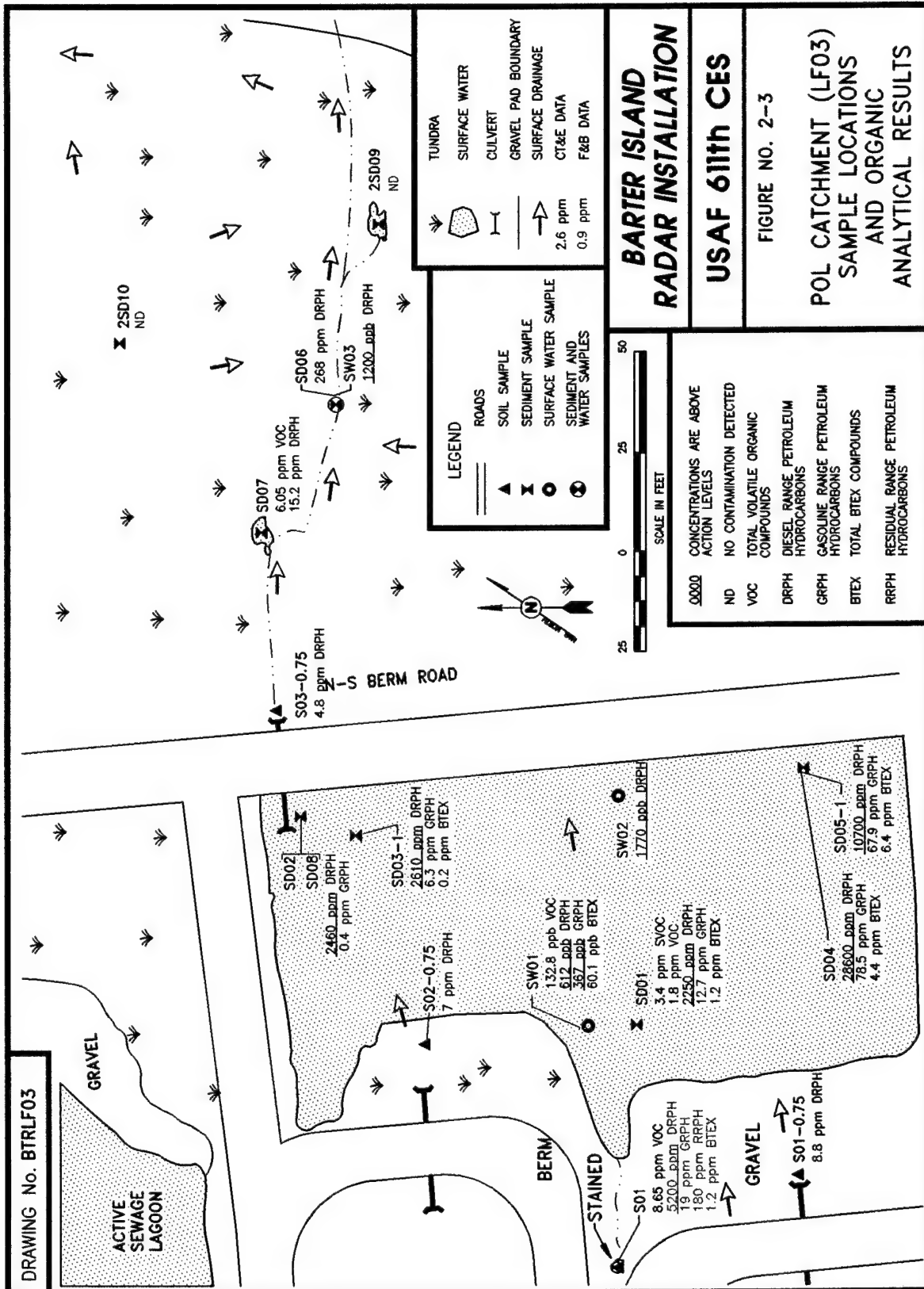
Trichloroethene and manganese were identified as COCs for the surface water at the Current Landfill (Figure 2-4). The maximum concentration of trichloroethene exceeded the drinking water MCL (Table 2-1). The maximum concentration of manganese exceeded the noncarcinogen RBSL (Table 2-1).

2.1.5.4 Contaminated Ditch (SD08). DRPH, GRPH and beryllium were identified as COCs for the soil matrix at the Contaminated Ditch (Figures 2-5 and 2-6). The maximum concentrations of DRPH and GRPH exceeded the ARARs for diesel and gasoline contamination of soil (Table 2-1) (ADEC 1991). The maximum concentration of beryllium exceeded the carcinogen RBSL (Table 2-1).

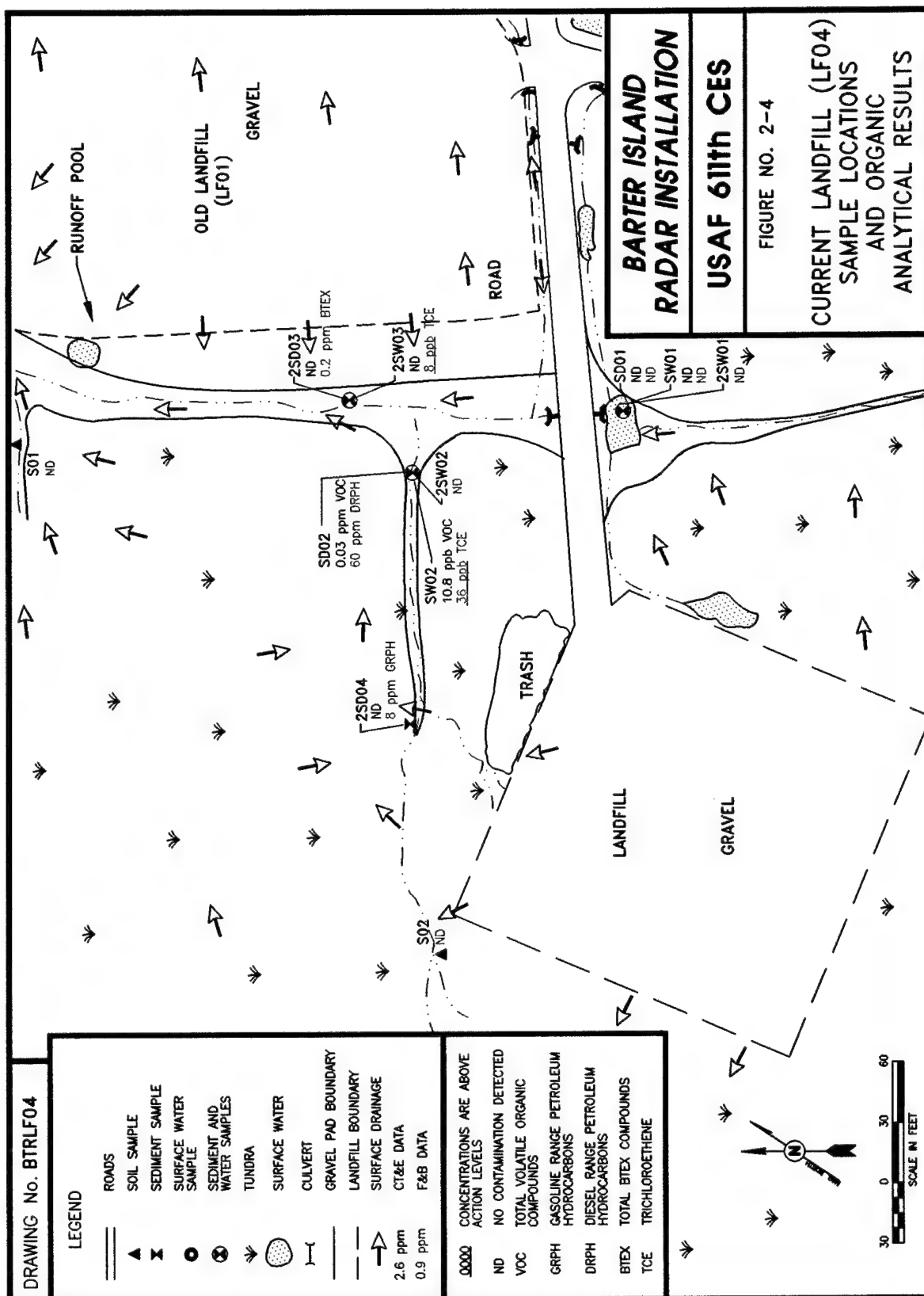
No COCs were identified for surface water at the Contaminated Ditch (Figure 2-5) based on a comparison of the maximum concentrations of the detected chemicals to background levels, and the RBSL and/or ARAR concentrations (Table 2-1).



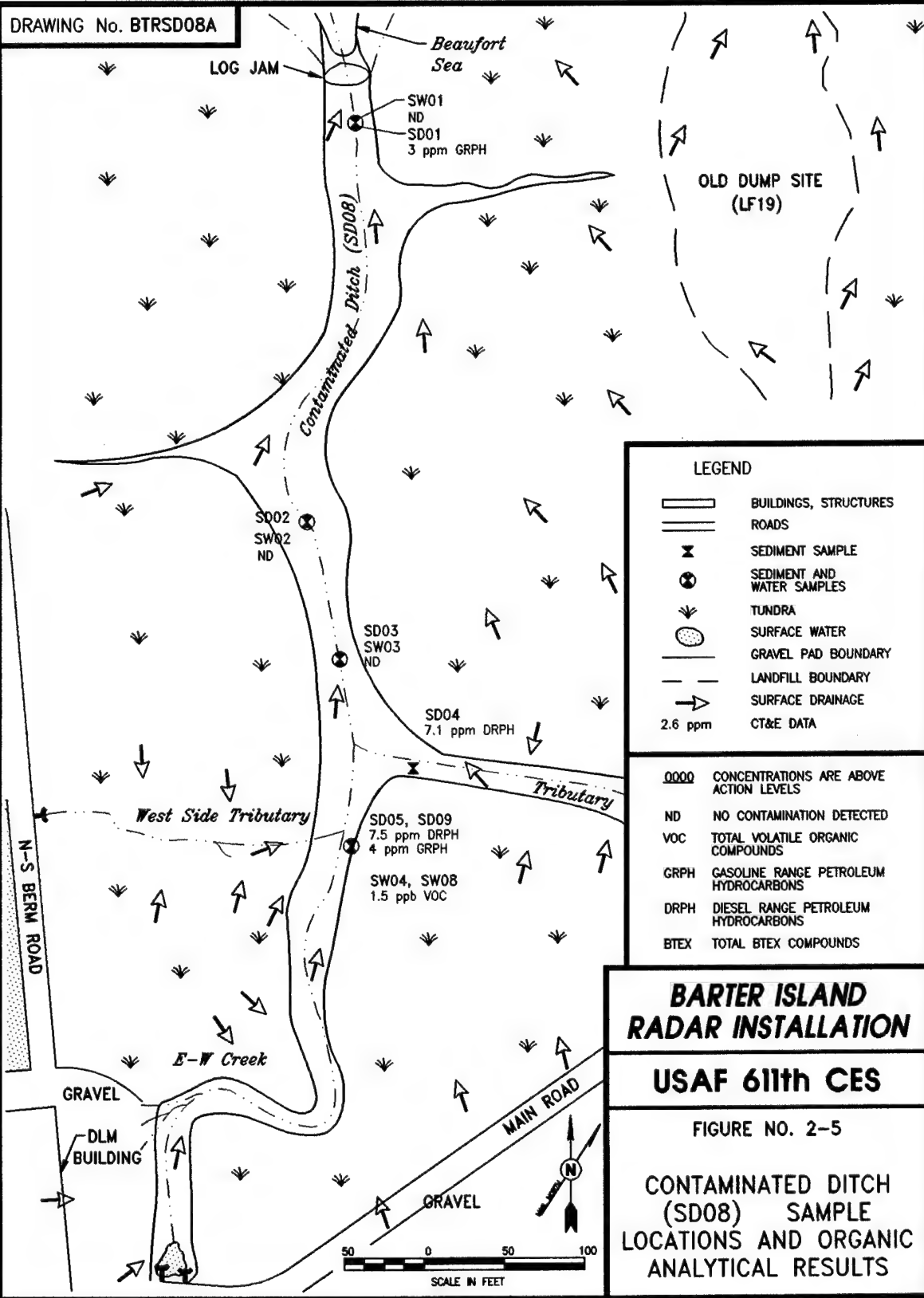
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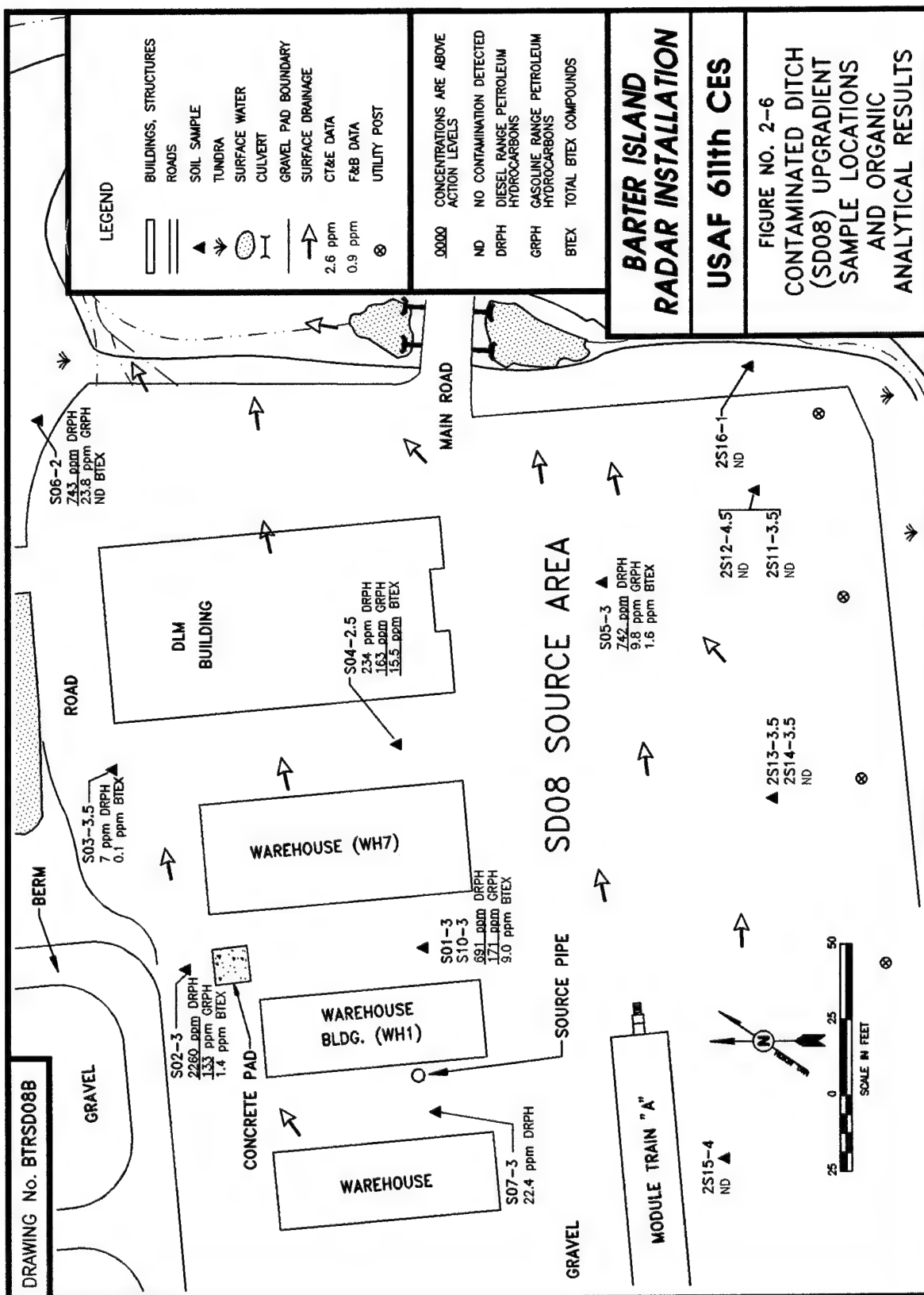
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2.1.5.5 Old Runway Dump (LF12). No COCs were identified for the soil matrix at the Old Runway Dump (Figure 2-7) based on a comparison of the maximum concentrations of the detected chemicals to background levels, and the RBSL and/or ARAR concentrations (Table 2-1).

No surface water bodies were identified at the Old Runway Dump.

2.1.5.6 Heated Storage Building (SS13). DRPH, GRPH, RRPB, and Aroclor 1254 were identified as COCs for the soil matrix at the Heated Storage Building site (Figures 2-8 and 2-9). The maximum concentrations of DRPH, GRPH, and RRPB exceeded the ARARs for petroleum hydrocarbon contamination of soil (Table 2-1) (ADEC 1991). The maximum concentration of Aroclor 1254 exceeded the noncarcinogen RBSL (Table 2-1).

DRPH, benzene, tetrachloroethane and manganese were identified as COCs for the surface water at the Heated Storage Building site (Figures 2-8 and 2-9). The maximum concentration of DRPH exceeded the noncarcinogen RBSL for diesel contamination of surface water (Table 2-1). The maximum concentrations of benzene and tetrachloroethane exceeded their carcinogen RBSLs (Table 2-1). The maximum concentration of manganese exceeded the noncarcinogen RBSL (Table 2-1). 1,2-Dichloroethane was detected at this site; however, 1,2-dichloroethane was also detected in the majority of blank and background samples at similar concentrations. Therefore it was not included as a COC.

2.1.5.7 Garage (SS14). DRPH, GRPH, RRPB, benzene, and bis(2-ethylhexyl)phthalate were identified as COCs for the soil matrix at the Garage (Figure 2-10). The maximum concentrations of DRPH, GRPH, and RRPB exceeded their ARARs for petroleum hydrocarbon contamination of soil (Table 2-1). The maximum concentration of benzene exceeded the ARAR for benzene in soil (Table 2-1) (ADEC 1991). The maximum concentration of bis(2-ethylhexyl)phthalate exceeded the carcinogen RBSL (Table 2-1).

No COCs were identified for surface water at the Garage (Figure 2-10) based on a comparison of the maximum concentrations of the detected chemicals to background levels, and the RBSL and/or ARAR concentrations (Table 2-1).

2.1.5.8 Weather Station Building (SS15). DRPH and GRPH were identified as COCs for the soil matrix at the Weather Station Building (Figure 2-11). The maximum concentrations of DRPH and GRPH exceeded the ARARs for petroleum hydrocarbon contamination of soil (Table 2-1) (ADEC 1991).

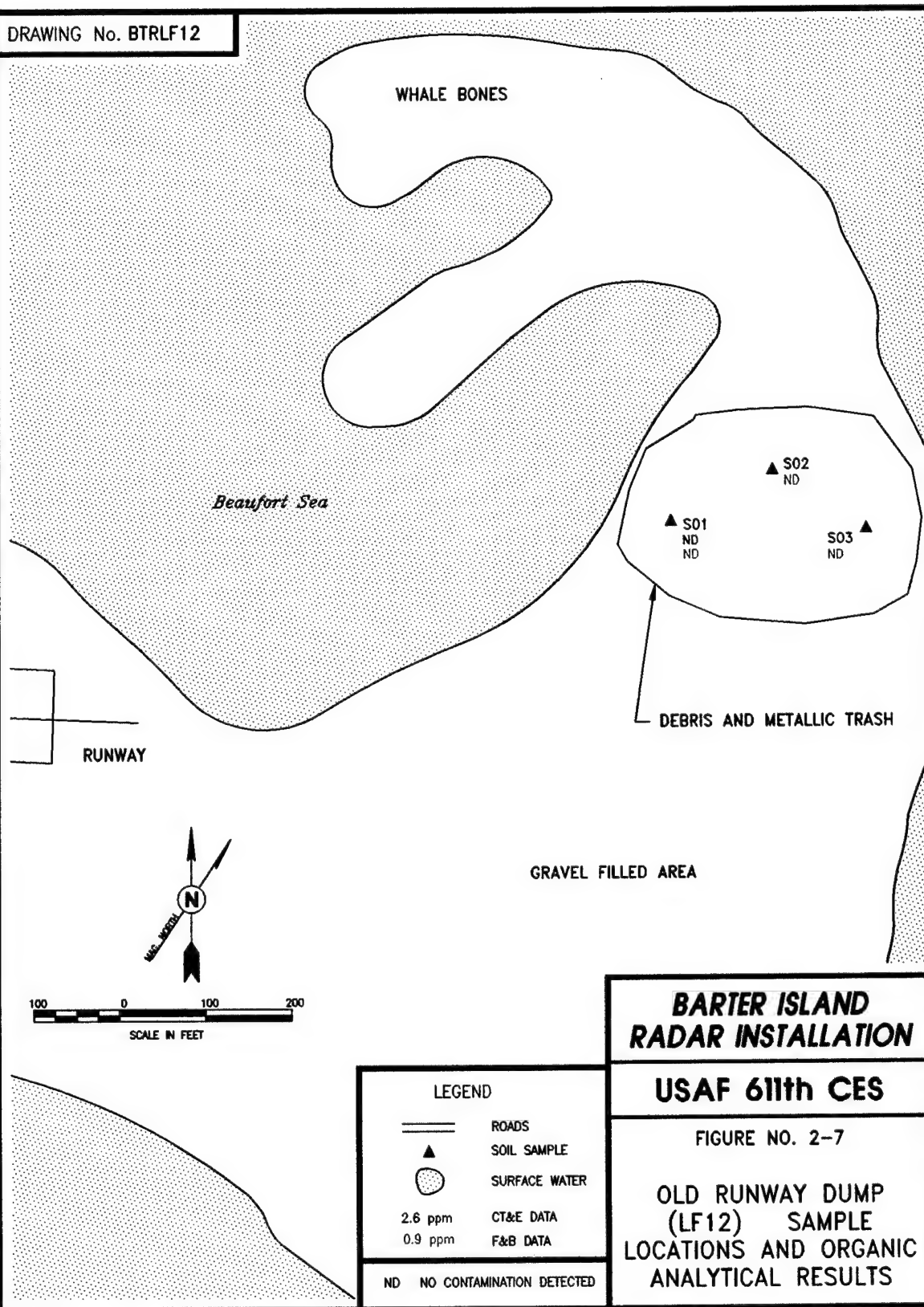
No surface water bodies were identified at the Weather Station Building.

2.1.5.9 White Alice Facility (SS16). Aroclor 1254 was identified as a COC for the soil matrix at the White Alice Facility (Figure 2-12). The maximum concentration of Aroclor 1254 exceeded the noncarcinogen RBSL (Table 2-1). No other COCs were identified for the site.

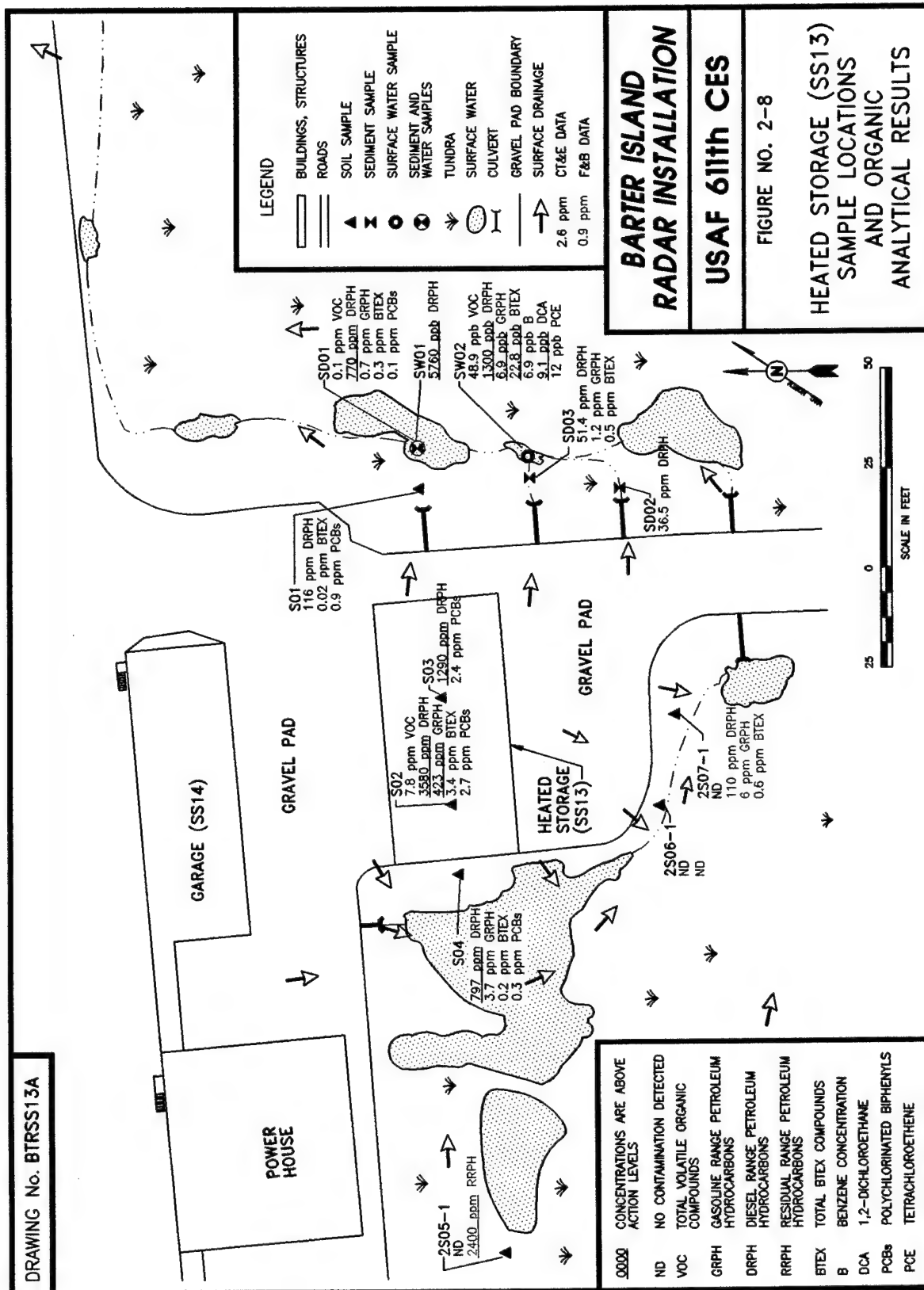
No surface water bodies were identified at the White Alice Facility.

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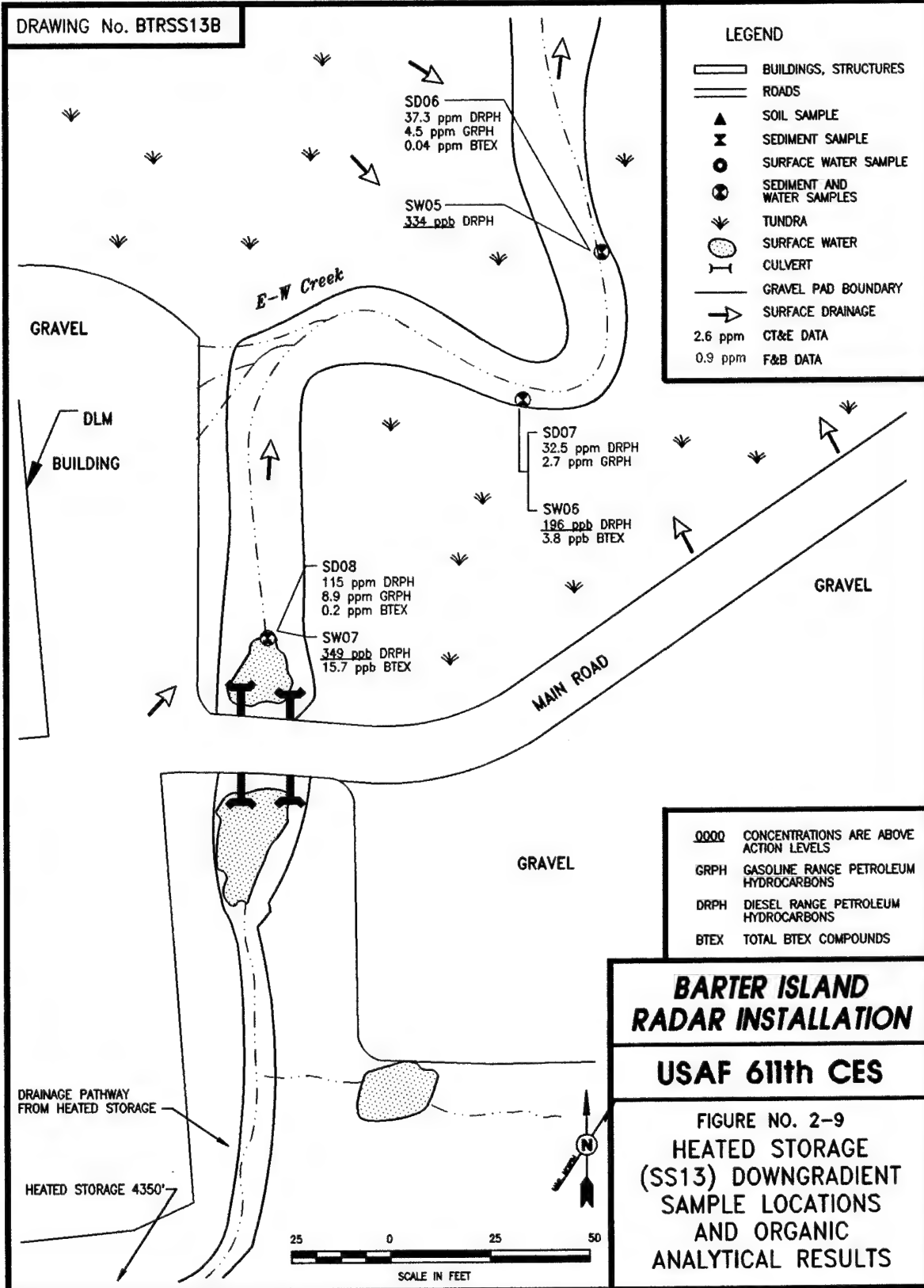


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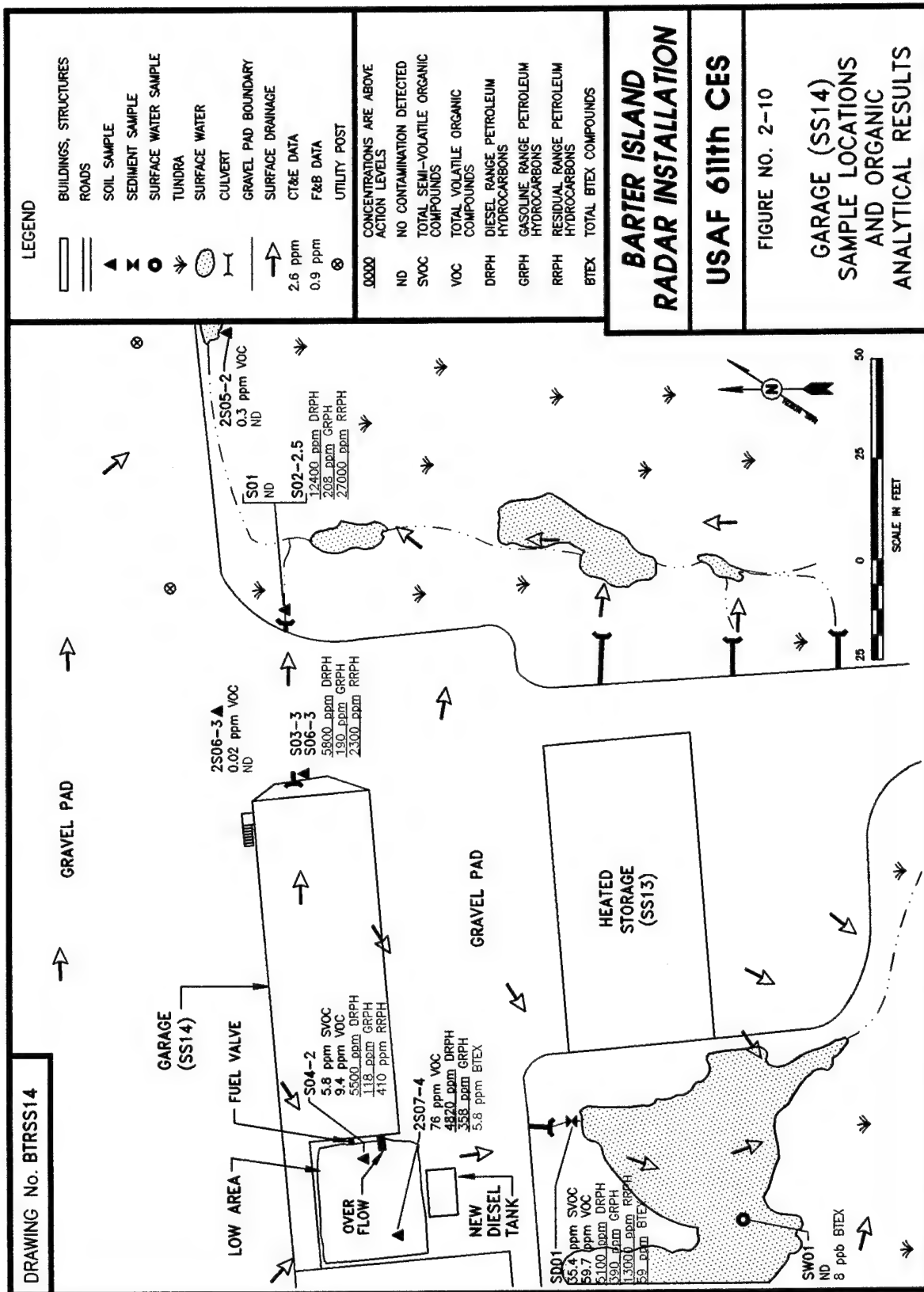


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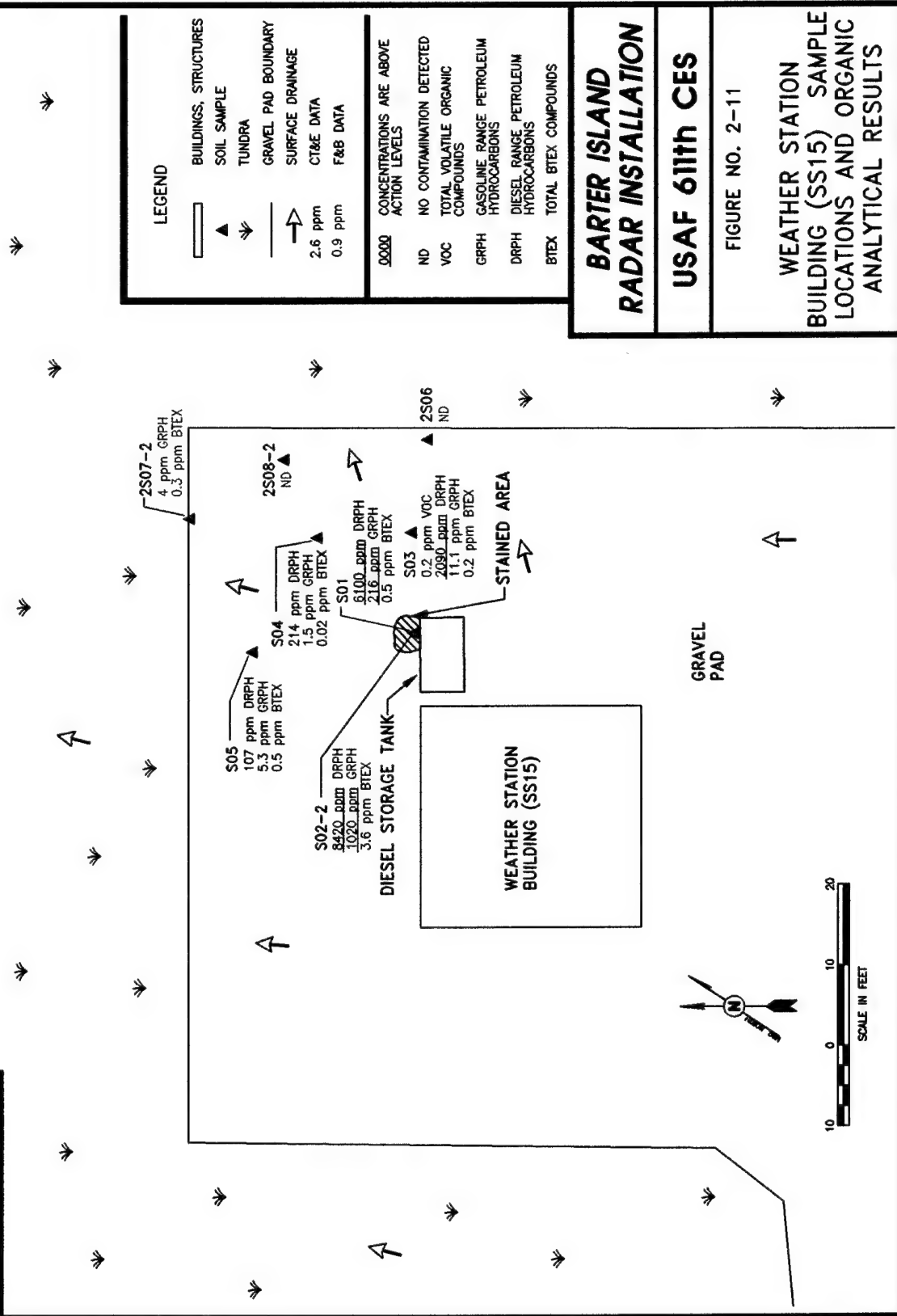


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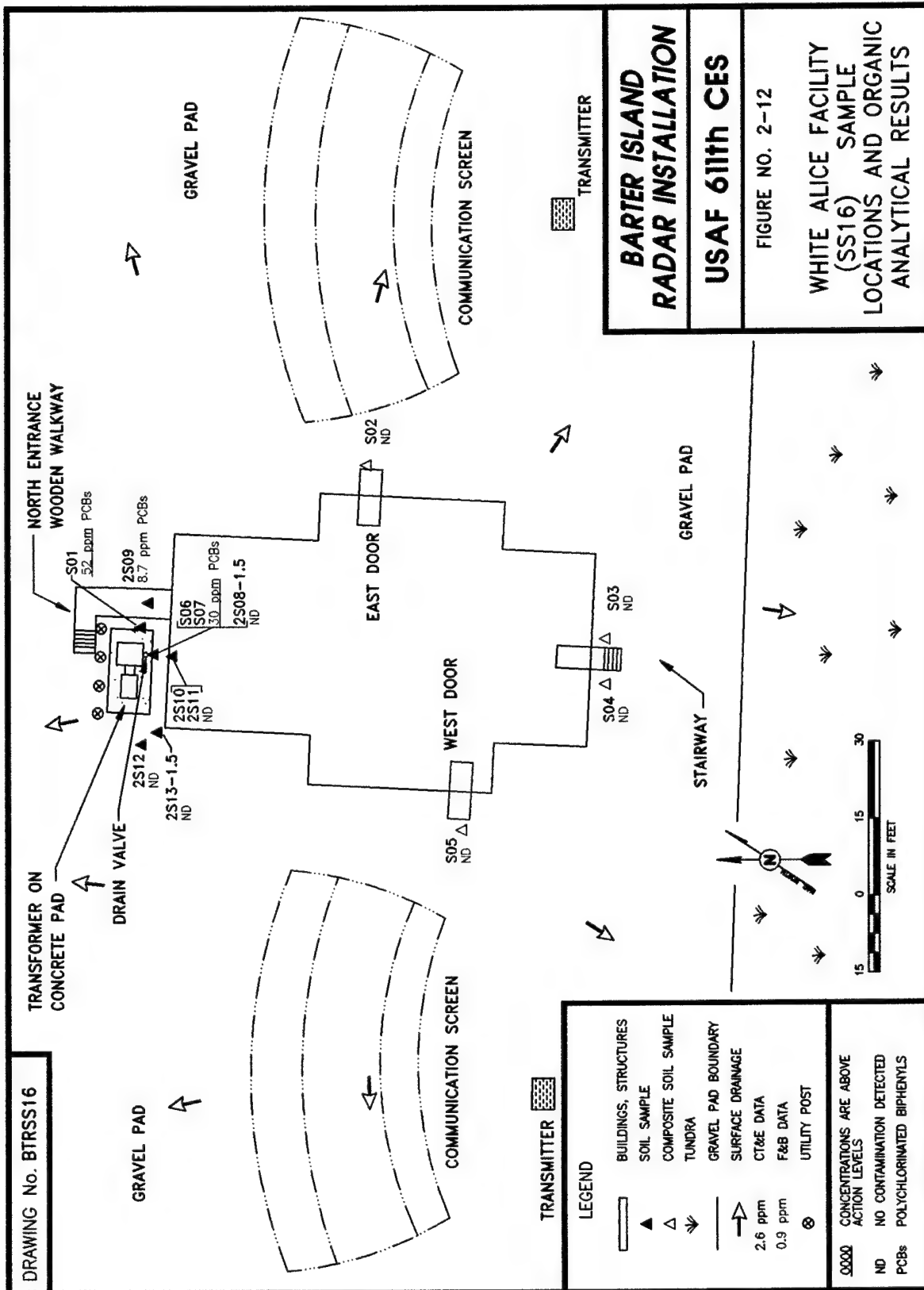


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2.1.5.10 POL Tanks (ST17). DRPH and GRPH were identified as COCs for the soil matrix at the POL Tanks (Figure 2-13). The maximum concentrations of DRPH and GRPH exceeded the ARARs for petroleum hydrocarbon contamination of soil (Table 2-1) (ADEC 1991).

No surface water bodies were identified at the POL Tanks site.

2.1.5.11 Fuel Tanks (ST18). No COCs were identified for the soil matrix at the Fuel Tanks site (Figure 2-14) based on a comparison of the maximum concentrations of the detected chemicals to their RBSL and/or ARAR concentrations (Table 2-1).

No COCs were identified for surface water at the Fuel Tanks site (Figure 2-14) because no contaminants were detected at concentrations greater than the analytical detection limits (Appendix D).

2.1.5.12 Old Dump Site (LF19). DRPH and RRPB were identified as COCs for the soil matrix at the Old Dump Site (Figure 2-15). The maximum concentrations of DRPH and RRPB exceeded the ARARs for petroleum hydrocarbon contamination of soil (Table 2-1) (ADEC 1991).

No COCs were identified for surface water at the Old Dump Site (Figure 2-15) based on a comparison of the maximum concentrations of the detected chemicals to background levels, and the RBSL and/or ARAR concentrations (Table 2-1).

2.1.5.13 Bladder Diesel Spill (SS20). No COCs were identified for the soil matrix at the Bladder Diesel Spill site (Figure 2-16) based on a comparison of the maximum concentrations of the detected chemicals to background levels and the RBSL and/or ARAR concentrations (Table 2-1).

No COCs were identified for surface water at the Bladder Diesel Spill site (Figure 2-16) because only naphthalene was detected and this chemical does not have an associated RBSL or ARAR (Table 2-1).

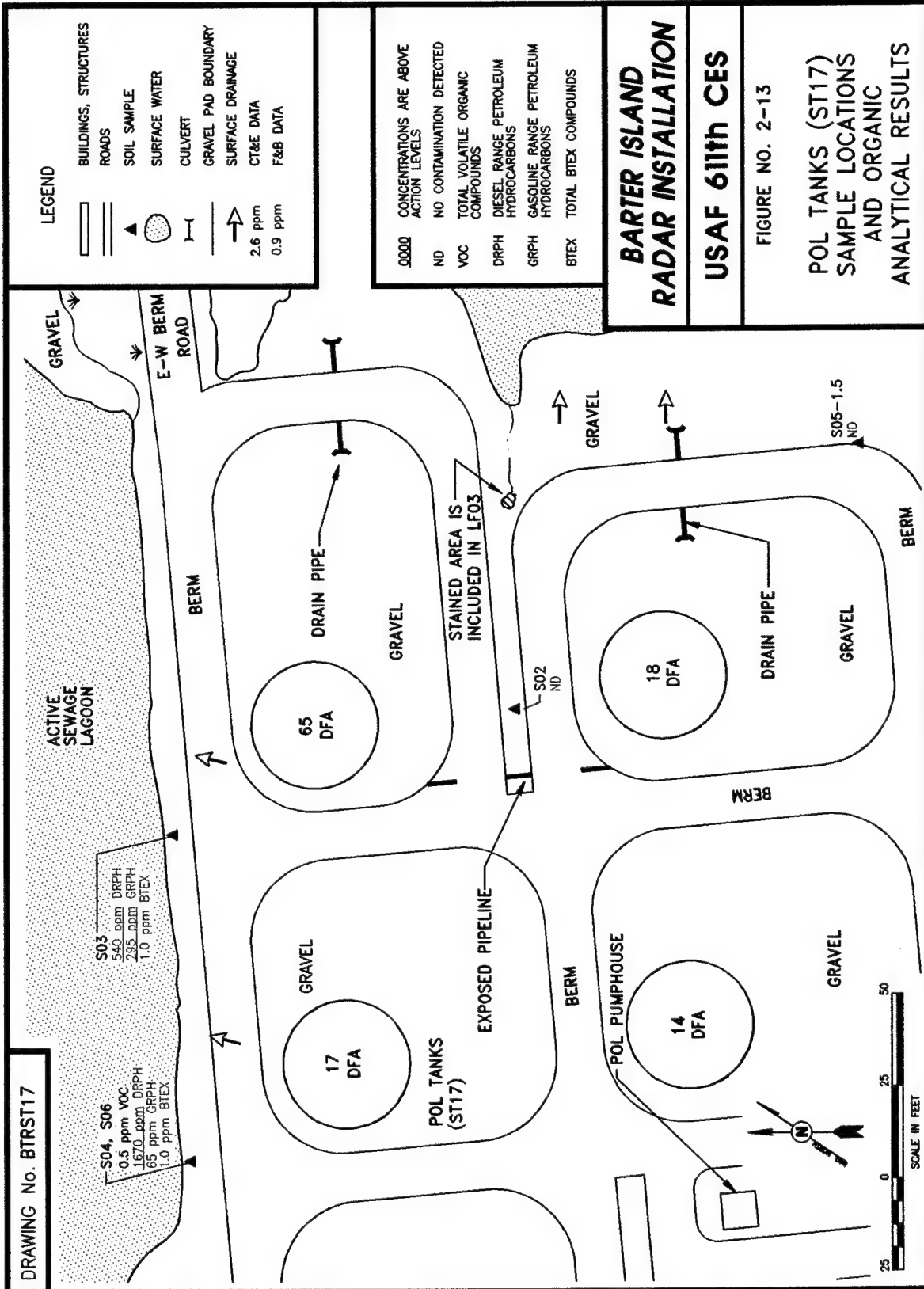
2.1.5.14 JP-4 Spill (SS21). DRPH, GRPH, and benzene were identified as COCs for the soil matrix at the JP-4 Spill site (Figure 2-17). The maximum concentrations of DRPH and GRPH exceeded the ARARs for petroleum hydrocarbon contamination of soil (Table 2-1) (ADEC 1991). The maximum concentration of benzene exceeded the carcinogen RBSL (Table 2-1).

No surface water bodies were identified at the JP-4 Spill site.

2.2 EXPOSURE ASSESSMENT

The exposure assessment section of a baseline human health risk assessment identifies and describes potential receptors and the exposure pathways by which exposure may occur, and estimates the magnitude of those exposures. This section includes an analysis of which pathways are complete (2.2.1), an estimation of the total intake of the chemicals (2.2.2) and a summary of how the average daily dose was calculated (2.2.3).

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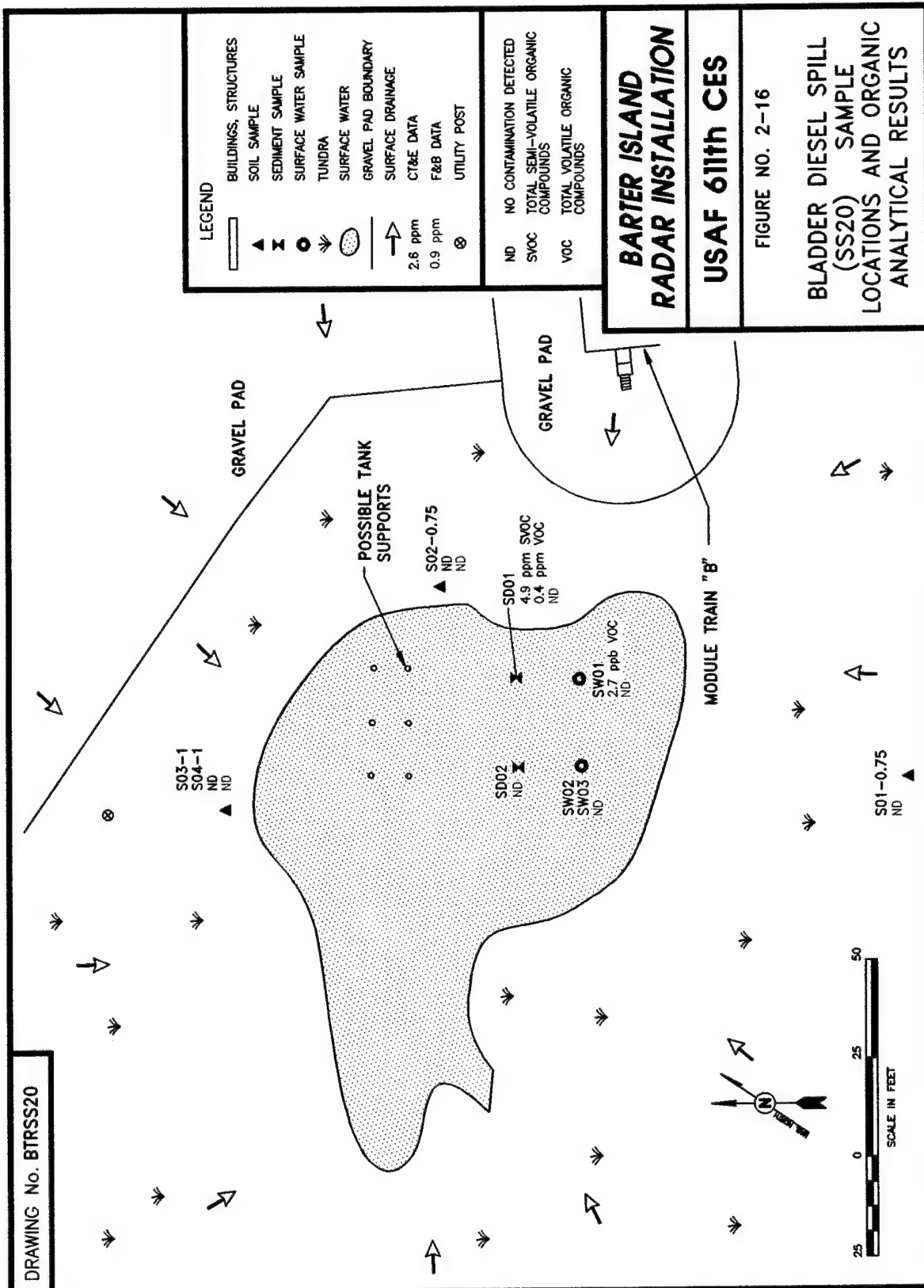


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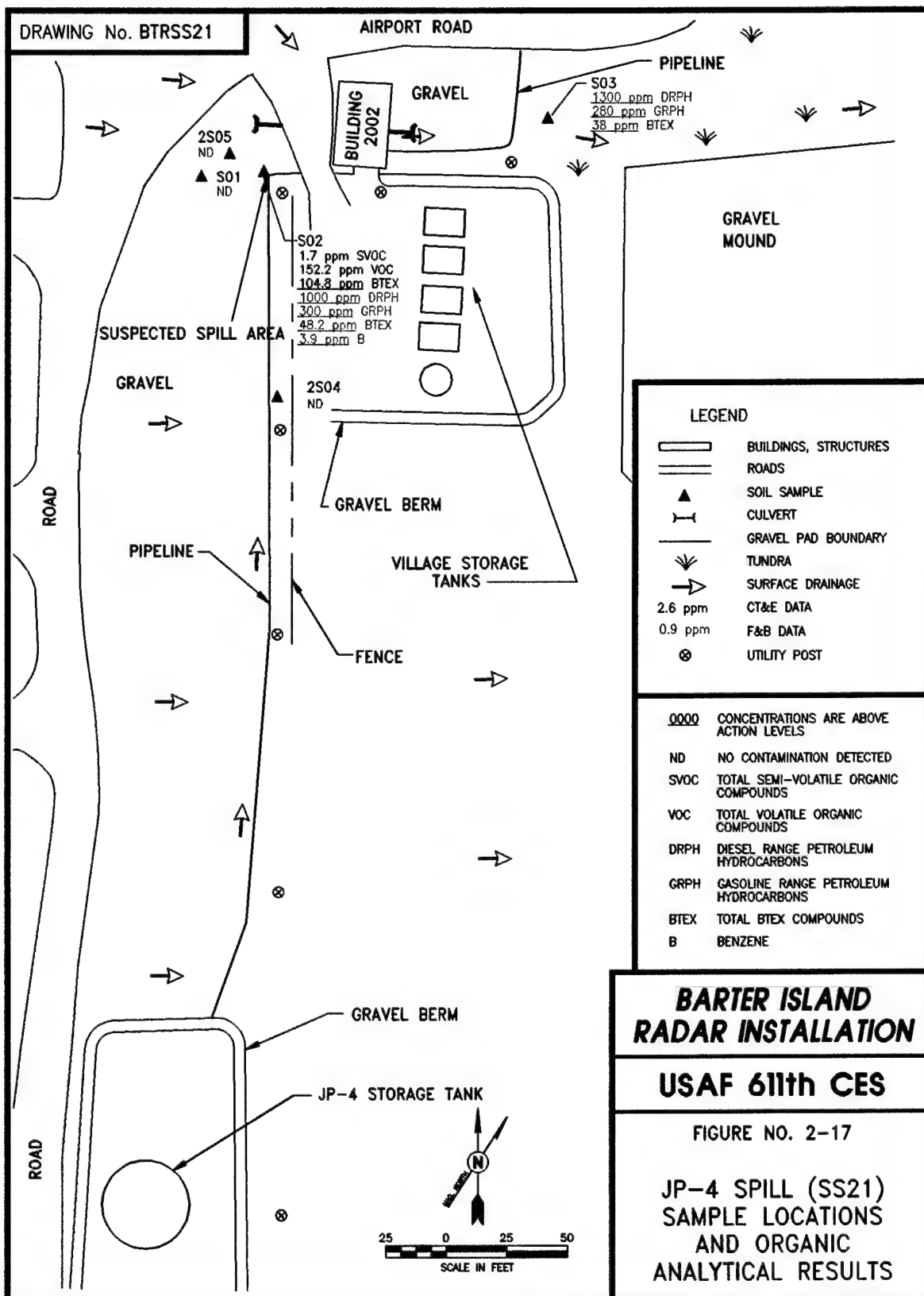


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2.2.1 Pathway Analysis

Pathway analysis involves the evaluation of the components of potential exposure pathways and a determination of whether each pathway is complete. An exposure pathway describes the course a chemical will take from a source to an exposure point where a receptor can come into contact with the chemical. A complete exposure pathway has five components:

- source of contamination;
- release mechanism;
- transport mechanism;
- exposure point; and
- receptor.

If one component of an exposure pathway does not exist, then exposure will not occur and there is no health risk. For example, if a shallow aquifer was contaminated with tetrachloroethylene, but that aquifer was not used as a water supply, no exposure point would exist and a ground water ingestion pathway would not be complete.

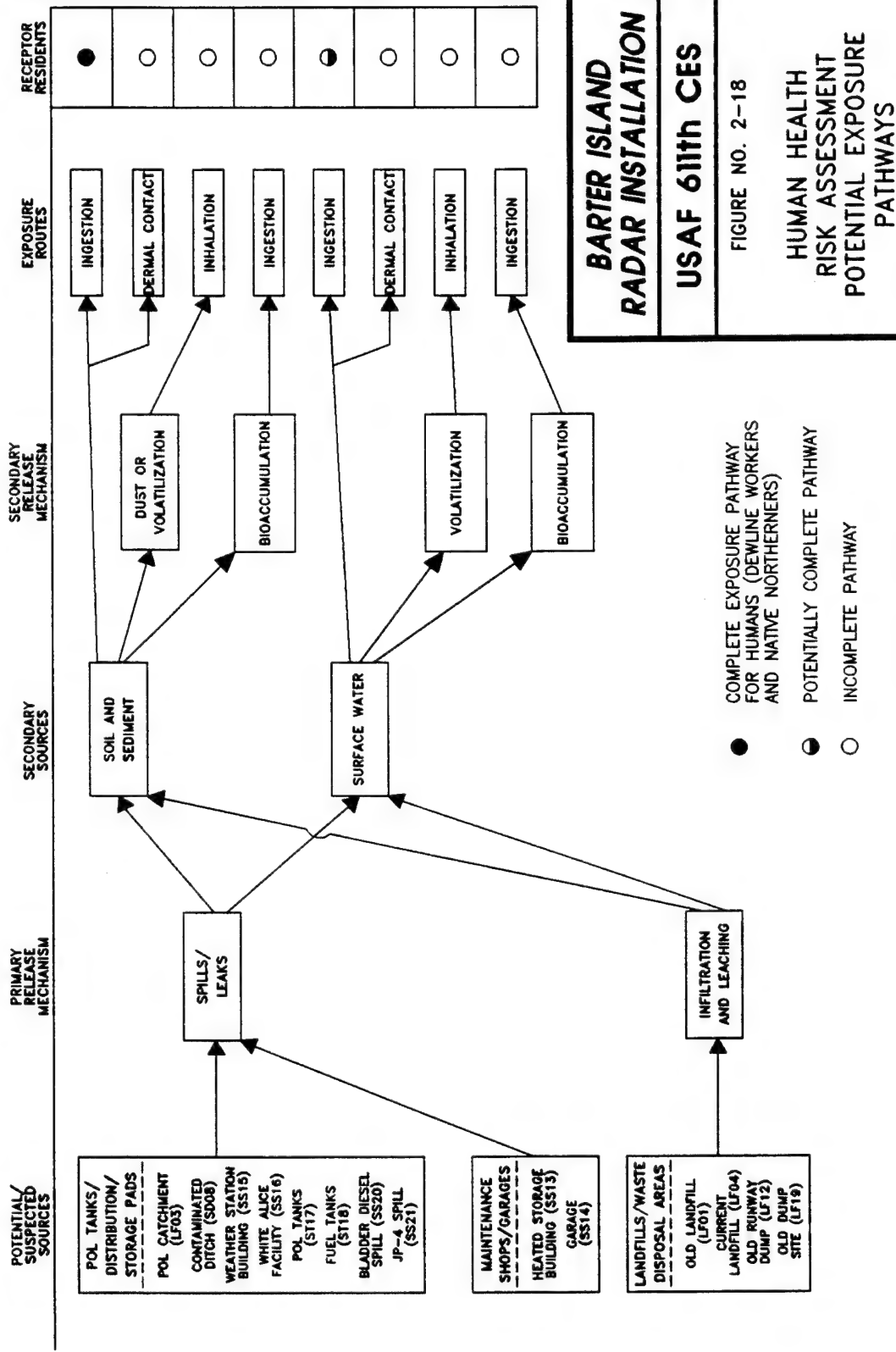
The potential exposure pathways evaluated for the Barter Island human health risk assessment are presented in Figure 2-18 and Table 2-2, and are discussed in Sections 2.2.1.1 through 2.2.1.4.

2.2.1.1 Soil and Sediment Ingestion. Barter Island installation workers and native villagers may be exposed to soil and sediment contaminated by operations at the installation. The most likely exposure routes are incidental ingestion of soil and dermal absorption of contaminants in the soil. Site-specific characteristics will limit the magnitude, frequency, and duration of exposures to soil and sediment. The ground is covered with snow and ice, which eliminates soil or sediment exposure, for approximately nine months of the year. In the summer months when snow cover is generally absent, cool temperatures (30°F to 46°F) keep both workers and villagers in heavy, long-sleeved clothing and gloves that eliminate dermal contact with and hand-to-mouth transfer of soil. Therefore, although both the incidental soil ingestion and dermal contact pathways are considered to be potentially complete, only incidental ingestion of soil or sediment will be evaluated further in this risk assessment.

The exposure assumptions used to evaluate the soil and sediment ingestion pathway are upper bound residential scenario assumptions and, therefore, probably overestimate the true hazard or risk associated with this pathway. The purpose of using residential assumptions is to evaluate the hazard or risk associated with future residential use of the Barter Island installation. Although the Air Force has no plans to retire the Barter Island radar installation, it is possible that the installation may be retired and released for civilian use, in which case residential use of the installation may be possible.

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TABLE 2-2. EXPOSURE PATHWAY ANALYSIS FOR BARTER ISLAND HUMAN HEALTH RISK ASSESSMENT

POTENTIALLY CONTAMINATED MEDIUM	POTENTIAL ROUTES OF EXPOSURE	POTENTIAL RECEPTORS	PATHWAY COMPLETE?	EXPOSED POPULATION ESTIMATE
Soil	Ingestion, dermal absorption	DEW Line workers, native villagers	Ingestion, Yes Dermal Contact, No	275
Sediments	Ingestion, dermal absorption	DEW Line workers, native villagers	Ingestion, Yes Dermal Contact, No	275
Air	Inhalation of volatiles from soil or surface water or inhalation of fugitive dust	DEW Line workers, native villagers	No, volatile concentrations in soil and surface water are very low; dust generation is not likely due to marshy vegetated landscape and high humidity.	0
Surface Water	Incidental ingestion, dermal absorption	DEW Line workers, native villagers	Maybe, drinking water supplies are either upgradient from installation or in unaffected areas; fishing occurs in unaffected areas; swimming does not occur on-site; however, incidental exposure may occur during installation operations or trespassing by native villagers.	275
Ground Water	Ingestion, dermal absorption	DEW Line workers, native villagers	No, permafrost limits presence of ground water to shallow active layer that is not used for any purpose.	0

2.2.1.2 Inhalation. Barter Island installation workers and native villagers may be exposed to site contamination by inhalation of organic compounds that have volatilized from the soil or surface water, or by inhalation of windborne dust to which contamination has adsorbed. These exposure pathways are not considered complete for the Barter Island risk assessment because snow and ice cover the site for approximately nine months of the year, and during the summer months the high humidity, vegetative cover, and thawing of surface and active layer water significantly limit the entrainment of dust particles. The generally low temperatures and high moisture content of the soil also tends to inhibit volatilization. The inhalation pathway will not be considered further in this risk assessment.

2.2.1.3 Surface Water Ingestion. Fresh water for drinking and domestic uses is supplied to the Barter Island installation and to the village of Kaktovik from Fresh Water Lake, located south of the installation and southwest of the village. This lake is upgradient from the installation and village and is, therefore, unaffected by any potential contamination from the radar installation. Other surface water features, particularly those potentially contaminated by operations at the installation, are not likely to be used for drinking or other domestic purposes even on an incidental basis. This is because these surface water features are not reliable, being frozen for most of the year. Ingestion of surface water will, however, be considered a potentially complete exposure pathway.

2.2.1.4 Ground Water. Permafrost limits the presence of ground water to the active layer which thaws during the summer months. The water present in the active layer is not known to be used for any purpose, therefore a ground water pathway may be eliminated from consideration in this risk assessment.

2.2.2 Estimation of Chemical Intake

The exposure assessment for the Barter Island DEW Line installation required the development of site-specific assumptions because of the unique location of the installation and potential receptors. This section of the report is focused on the exposure variables for which site-specific assumptions were made. These variables include:

- exposure frequency;
- exposure duration;
- ingestion of locally produced meat (e.g., caribou, fish, and birds);
- ingestion of locally produced vegetation (e.g., berries);
- soil ingestion rate; and
- rate of dermal contact with soil.

The exposure assumptions used in the human health risk assessment are presented in Table 2-3.

Three potential receptor groups will be evaluated for the Barter Island risk assessments: an adult assigned to a DEW Line installation (worker), an adult native of the North Slope of Alaska (native) and a native child (child). The first two receptor groups are considered to represent the reasonable maximum exposure that might occur at an installation that is not in close proximity to a native village. Because the Barter Island installation is close to the village of Kaktovik, a child will be considered as a potentially exposed individual.

The estimation of chemical intake requires the evaluation of several exposure variables: exposure point concentration; exposure frequency; exposure duration; ingestion of locally produced meat, fish, and vegetation; soil ingestion; drinking water ingestion; dermal contact with soil; inhalation; and body weight. These exposure variables are discussed in the following sections.

TABLE 2-3. EXPOSURE ASSUMPTIONS FOR ESTIMATING CHEMICAL INTAKE

PARAMETER	DEW LINE WORKER	NATIVE NORTHERN ADULT	NATIVE NORTHERN CHILD
Exposure Frequency - Soil Ingestion (days/year)	30	30	30
Exposure Frequency - Water Ingestion (days/year)	180	180	N/A
Exposure Duration (years)	10	55 ^a	6 ^a
Soil Ingestion Rate (mg/day)	50	100	200
Drinking Water Ingestion Rate (L/day)	2	2	N/A
Average Body Weight (kg)	70	70	15
Averaging Time (days)	25,550 (cancer) 3,650 (noncancer)	20,075 (cancer) 20,075 (noncancer)	2,190 (noncancer)

N/A Not applicable; drinking water pathway evaluated for adult only.

^a Exposure duration for water ingestion pathway is 55 years. For soil ingestion, exposure duration is 6 years as a child and 49 years as an adult.

2.2.2.1 Exposure Point Concentration. Based on the amount of analytical data available for the risk assessment of the Barter Island facility, and the requirement that the risk characterization be conducted individually for each of the fourteen sites, only maximum concentrations of the COCs were used for exposure point concentrations. This approach yields a conservative upper bound estimate of the average daily dose to which potential receptors would be exposed.

2.2.2.2 Exposure Frequency. The exposure frequency variable is an estimate of the amount of time a potential receptor may come in to contact with contaminated media. For the DEW Line worker, the exposure frequency estimate is based on a duty rotation of eleven months on-site, one month off-site. During the eleven months of on-site duty, it is estimated that the worker is outside for four hours per day. The remaining 20 hours is spent inside the module trains or enclosed vehicles, where exposure to contaminated media is not expected to occur. An estimated exposure frequency for the DEW Line worker is, therefore, 11 months/year x 30 days/month x 4 hours/day x 1 day/24 hours = 55 days/year. The primary environmental medium of concern is, however, contaminated soil and this estimate of exposure frequency does not account for the number of days per year that snow covers the ground and eliminates the potential for contact with contaminated soil. Six months is a conservative average estimate of the number of months per year of snow cover at all eight DEW Line installations. To be even more conservative, it is assumed that a worker's tour of duty includes all six of the months without snow cover, thus an exposure frequency of 30 days/year for the DEW Line worker is recommended.

The exposure frequency estimate for a native adult or child of the North Slope is based on an estimate of the frequency with which the individual will visit a DEW Line installation. Such visits are most likely to occur at installations sited within the vicinity of a native village. For example, at Barter Island, native adults and children use the facility as an access route to coastal areas and for recreation. In this case, a conservative estimate of exposure would be expected to be similar to that of a worker, 4 hrs/day x 30 days per month x 1 day/24 hrs x 6 months of exposed soil = 30 days per year.

The exposure frequency for water ingestion was conservatively estimated at 180 days/year that surface water would be available (i.e., not frozen).

2.2.2.3 Exposure Duration. The exposure duration variable is an estimate of the amount of time a potential receptor will remain at or near a DEW Line installation over a lifetime. For the DEW Line worker the exposure duration is an estimate of the maximum tour of duty at an installation. A conservative estimate of the duration of a tour at a particular installation is 10 years. For the potential native receptor, a conservative estimate of exposure duration is 55 years. EPA's default reasonable maximum exposure duration is 30 years, however this is based on the overall U.S. population. Because the Alaskan natives are more likely to remain in their village for an entire lifetime, 55 years was determined to be a more appropriate estimate.

2.2.2.4 Averaging Time. The averaging time represents the period of time over which exposure is averaged and is based on the assumption that intermittent exposure at a given contaminant concentration is equivalent to a continuous exposure at a lower concentration. For the DEW Line worker, the averaging time is based on the EPA default lifetime of 70 years for evaluation of carcinogens, and 10 years (equivalent to the exposure duration) for the evaluation of noncarcinogens. For the native northern adult receptors an averaging time of 70 years for carcinogens was also chosen. To evaluate exposure to noncarcinogens in soil and sediment for the native northern adult and child, an averaging time of 49 years as an adult and 6 years as a child was used (to account for 55 year total exposure). To evaluate the exposure of native northern receptors to noncarcinogens in water an averaging time of 55 years was used.

2.2.2.5 Ingestion of Locally Produced Meat, Fish and Vegetation. The food supplies of DEW Line installation workers are largely imported from outside the area. Occasionally, a worker would be expected to ingest a locally caught fish or game animal, but the frequency and magnitude of this ingestion is expected to have a negligible effect on exposure to the COCs. Food supplies for the natives of Kaktovik are also largely imported from outside the area and the reliance on hunting and fishing for subsistence is decreasing substantially as the economy moves from subsistence to wage labor (Chance 1990). Inupiat, in general, have less time to hunt and fish than in the past. Most of the hunting and fishing that is done occurs outside the village and away from the Barter Island DEW Line installation in areas unaffected by the installation. It is not likely that contamination observed at the installation has affected the mammals, birds, and vegetation that are collected for consumption. Therefore, the consumption of locally produced food is not likely to pose a significant risk of adverse health effects and will not be considered a complete exposure pathway. The ecological risk assessment, Section 3, presents a detailed assessment of risks to ecological receptors.

2.2.2.6 Soil Ingestion Rate. A conservative approach to estimating soil ingestion rate is to assume that the EPA default soil ingestion rate (100 mg/day) for adults in a residential setting is applicable to the Barter Island installation. The EPA default soil ingestion rate for children is 200 mg/day; this is the recommended value for the risk assessment.

2.2.2.7 Drinking Water Ingestion Rate. There are no circumstances at the Barter Island installation that would invalidate the EPA default adult drinking water ingestion rate of 2 L/day. Therefore, this is the recommended value for both workers and natives. However, in most, if not all, cases drinking water is imported from off-site so this may not be a route of potential exposure.

By convention (EPA 1989a), noncancer hazard and cancer risk associated with the drinking water pathway are evaluated for an adult receptor, not a child (Table 2-3). The basis for this approach is probably that adults drink more water than children. In contrast, adults are assumed to eat less soil than children, therefore, a soil ingestion rate for a child and another for an adult have been developed for use in human health risk assessments.

2.2.2.8 Dermal Contact with Soil Rate. Because of the harsh North Slope weather, potential receptors (both workers and natives) are expected to be heavily clothed and gloved. Therefore, dermal exposure to contaminated soils is considered to be negligible. In addition, the duties of installation workers that involve soil work (excavating, grading, etc.) are conducted in equipment with enclosed cabs. Thus a dermal contact rate does not appear to be necessary for the exposure assessment.

2.2.2.9 Inhalation Rate. The inhalation pathway is not complete (Section 2.2.1.2). Therefore, no estimate for this variable is necessary.

2.2.2.10 Body Weight. There are no circumstances at the Barter Island installation that would invalidate the EPA default adult body weight of 70 kg. Therefore, this is the recommended value for both workers and natives. The recommended body weight for children is the EPA default value of 15 kg.

2.2.3 Quantifying Exposure

For each complete, or potentially complete, exposure pathway at the Barter Island facility (soils ingestion, drinking water ingestion), the average daily dose (ADD) for estimating noncancer hazard, and the lifetime average daily dose (LADD) for estimating excess lifetime cancer risk were calculated. The equations used for the calculation of ADD and LADD are presented in Table 2-4.

The exposure assumptions assigned to each variable in these equations are presented in Table 2-3. The estimates of ADD and LADD for the COCs at each site are presented in the risk characterization tables in Appendix A.

TABLE 2-4. EQUATIONS USED FOR ESTIMATING POTENTIAL DOSE

EXPOSURE ROUTE	EQUATION	PARAMETER DEFINITIONS
Ingestion of Soil	<p>Native Northern Adults/Children</p> $ADD \text{ or } LADD \text{ (mg/kg/day)} = \frac{C_s * CF * EF}{AT} \sum_{i=1}^n \frac{IR_i * ED_i}{BW_i}$ <p>DEW Line workers:</p> $ADD \text{ or } LADD \text{ (mg/kg/day)} = \frac{C_s * CF * IR * EF * ED}{BW * AT}$	<p>= concentration in soil (mg/kg)</p> <p>= conversion factor (10⁻⁶ kg/mg)</p> <p>= ingestion rate (mg/day)</p> <p>= exposure frequency (days/year)</p> <p>= exposure duration (years)</p> <p>= body weight (kg)</p> <p>= averaging time (days/year x years)</p>
Ingestion of Surface Water	$ADD \text{ or } LADD \text{ (mg/kg/day)} = \frac{C_w * IR * EF * ED}{BW * AT}$	<p>= concentration in surface water (µg/L)</p> <p>= conversion factor (10⁻³ mg/µg)</p> <p>= ingestion rate (L/day)</p> <p>= exposure frequency (days/year)</p> <p>= exposure duration (years)</p> <p>= body weight (kg)</p> <p>= averaging time (days/year x years)</p>

2.3 TOXICITY ASSESSMENT

The purpose of the toxicity assessment is to weigh available evidence regarding the potential for particular contaminants to cause adverse effects in exposed individuals and to provide, where possible, an estimate of the relationship between the extent of exposure to a contaminant and the increased likelihood and/or severity of adverse effects. This is done separately for noncarcinogenic effects (2.3.1) and carcinogenic effects (2.3.2). Toxicity summaries are presented in Section 2.3.3.

Toxicity assessment for environmental contaminants is generally accomplished in two steps: hazard identification and dose-response assessment. Hazard identification is the process of determining whether exposure to an agent can cause an increase in the incidence of a particular adverse health effect (e.g., cancer, birth defects) and whether the adverse health effect is likely to occur in humans. Hazard identification involves characterizing the nature and strength of the evidence of causation. Dose-response evaluation is the process of quantitatively evaluating the toxicity information and characterizing the relationship between the dose of the contaminant administered or received and the incidence of adverse health effects in the exposed population. From this quantitative dose-response relationship, toxicity values (e.g., reference doses and slope factors) are derived that can be used to estimate the incidence or potential for adverse effects as a function of human exposure to the agent. These toxicity values are used in the risk characterization step to estimate the likelihood of adverse effects occurring in humans at different exposure levels.

2.3.1 Toxicity Assessment for Noncarcinogenic Effects

A reference dose, or RfD, is the toxicity value used most often in evaluating noncarcinogenic effects resulting from exposures at contaminated sites. Various types of RfDs are available depending on the exposure route (oral or inhalation), the critical effect (developmental or other), and the length of exposure being evaluated (chronic, subchronic, or single event). The oral RfDs used to estimate the noncancer hazard associated with exposure to soils, sediments, and surface water at the Barter Island facility are presented in Table 2-5.

A chronic RfD is defined as an estimate (with uncertainty spanning perhaps an order of magnitude or greater) of a daily exposure level for the human population, including sensitive subpopulations, that is likely to be without an appreciable risk of deleterious effects during a lifetime. Chronic RfDs are specifically developed to be protective for long-term exposure to a compound. Chronic RfDs generally should be used to evaluate the potential noncarcinogenic effects associated with exposure periods between 7 years (approximately 10 percent of a human lifetime) and a lifetime. Many chronic RfDs have been reviewed and verified by an intra-Agency RfD Workgroup and entered into the IRIS of the EPA.

TABLE 2-5. TOXICITY CRITERIA FOR NONCANCER EFFECTS OF THE CHEMICALS OF CONCERN FOR BARTER ISLAND

CHEMICAL	ORAL REFERENCE DOSE (RfD) (mg/kg-day)	TARGET ORGAN OR CRITICAL EFFECT (species) ^a	UNCERTAINTY FACTOR ^b	ORAL RfD SOURCE ^c
Aroclor 1254	0.00002	ocular exudate and styles, immunological effects (monkeys)	300	IRIS
Benzene	NA	NA	NA	NA
Beryllium	0.005	no observed effects (rat)	100	IRIS
Bis(2-ethylhexyl)phthalate	0.02	liver effects (guinea pig)	1000	IRIS
Dichlorobenzene, 1,4-	NA	NA	NA	NA
DRPH	0.08 ^d	liver effects (mice)	10,000	ECAO
GRPH	0.2 ^d	decreased body weight (rats)	1000	ECAO
Manganese (water)	0.005	CNS effects (humans)	1	IRIS
PCBs	NA	NA	NA	NA
RRPH	0.08 ^d	liver effects (mice)	10,000	ECAO
Tetrachloroethane	0.01	liver effects (mice)	1000	IRIS
Trichloroethene	NA	NA	NA	NA

- ^a A target organ is the organ apparently most sensitive to the toxicity of a chemical. A critical effect is reported when EPA has not identified a target organ for the toxicity of a given chemical.
- ^b The uncertainty factors used to develop oral reference doses are generally applied in multiples of 10 to account for shortcomings in the toxicological database. The greater the uncertainty factor, the less confidence that can be placed on that RfD. Factors of 10 are applied to account for human variability in toxic response, extrapolation from animal studies to humans, extrapolation of short-term exposures to long-term exposures, and for the extrapolation of a lowest observed adverse effect level (LOAEL) to a no observed adverse effect level (NOAEL).
- ^c Sources of oral RfD values are IRIS (Integrated Risk Information System), HEAST (Health Effects Assessment Summary Tables), or ECAO (The Environmental Criterion Assessment Office of EPA).
- ^d Oral RfD values for DRPH, GRPH, and RRPH are based on (EPA 1992d) and are considered provisional RfDs.
- NA Not available.

2.3.1.1 Concept of Threshold. For many noncarcinogenic effects, protective mechanisms are believed to exist that must be overcome before the adverse effect is manifested. For example, where a large number of cells perform the same or similar function, the cell population may have to be significantly depleted before the effect is seen. As a result, a range of exposures exists from zero to some finite value that can be tolerated by the organism with essentially no chance of expression of adverse effects. In developing a toxicity value for evaluating noncarcinogenic effects (i.e., an RfD), the approach is to identify the upper bound of this tolerance range (i.e., the maximum subthreshold level). Because variability exists in the human population, attempts are made to identify a subthreshold level protective of sensitive individuals in the population. For most chemicals, this level can only be estimated; the RfD incorporates uncertainty factors indicating the degree of extrapolation used to derive the estimated value. RfD summaries in IRIS also contain a statement expressing the overall confidence that the evaluators have in the RfD (high, medium, or low). The RfD is generally considered to have uncertainty spanning an order of magnitude or more, and therefore the RfD should not be viewed as a strict scientific demarcation between levels that are toxic and nontoxic.

2.3.2 Toxicity Assessment For Carcinogenic Effects

A slope factor and the accompanying weight-of-evidence determination are the toxicity data most commonly used to evaluate potential human carcinogenic risks. The methods EPA uses to derive these values are outlined below. Additional information can be obtained by consulting EPA's *Guidelines for Carcinogen Risk Assessment* (EPA 1986a) and IRIS Background Document #2 (IRIS 1994). The slope factors for the COCs at Barter Island are presented in Table 2-6.

2.3.2.1 Concept of Nonthreshold Effects. Carcinogenesis, unlike many noncarcinogenic health effects, is generally thought to be a phenomenon for which risk evaluation based on presumption of a threshold is inappropriate. For carcinogens, EPA assumes that a small number of molecular events can evoke changes in a single cell that can lead to uncontrolled cellular proliferation and eventually to a clinical state of disease. This hypothesized mechanism for carcinogenesis is referred to as "nonthreshold" because there is believed to be essentially no level of exposure to such a chemical that does not pose a finite probability, however small, of generating a carcinogenic response. That is, no dose is thought to be risk-free. Therefore, in evaluating cancer risks, an effect threshold cannot be estimated. For carcinogenic effects, EPA uses a two-part evaluation in which the substance first is assigned a weight-of-evidence classification, and then a slope factor is calculated.

2.3.2.2 Assigning a Weight-of-Evidence. In the first step of the evaluation, the available data are evaluated to determine the likelihood that the agent is a human carcinogen. The evidence is characterized separately for human studies and animal studies as sufficient, limited, inadequate, no data, or evidence of no effect. The characterizations of these two types of data are combined, and based on the extent to which the agent has been shown to be a carcinogen in experimental animals or humans, or both, the agent is given a provisional weight-of-evidence classification. EPA scientists then adjust the provisional classification upward or downward, based on other supporting evidence of carcinogenicity.

The EPA classification system for weight-of-evidence is shown in Table 2-7.

TABLE 2-6. TOXICITY VALUES FOR THE CARCINOGENICITY OF THE CHEMICALS OF CONCERN AT BARTER ISLAND

CHEMICAL	WEIGHT-OF-EVIDENCE (WOE)	TUMOR TYPE (species)	ORAL SLOPE FACTOR (kg-day/mg)	ORAL SLOPE FACTOR SOURCE ^a
Aroclor 1254	NA	NA	NA	NA
Benzene	A	leukemia (humans)	0.029	IRIS
Beryllium	B2	osteosarcoma (rabbit)	4.3	IRIS
Bis(2-ethylhexyl)phthalate	B2	liver carcinoma/adenoma (mouse)	0.014	IRIS
Dichlorobenzene, 1,4-	B2	liver tumors (mouse)	0.024	HEAST
DRPH	NA	NA	NA	NA
GRPH	C	liver adenoma/carcinoma (mouse)	0.0017	ECAO
Manganese (water)	NA	NA	NA	NA
PCBs	B2	trabecular adenoma/carcinoma (rat)	7.7	IRIS
RRPH	NA	NA	NA	NA
Tetrachloroethane	C-B2	not specified	0.052	ECAO
Trichloroethene	C-B2	not specified	0.011	ECAO

^a IRIS, Integrated Risk Information System; HEAST, Health Effects Assessment Summary Tables; ECAO, Environmental Criterion Assessment Office of EPA.
NA Not available.

TABLE 2-7. EPA WEIGHT-OF-EVIDENCE CLASSIFICATION SYSTEM FOR CARCINOGENICITY

GROUP	DESCRIPTION
A	Human carcinogen.
B1 or B2	Probable human carcinogen.
	B1 indicates that limited human data are available.
	B2 indicates sufficient evidence in animals and inadequate or no evidence in humans.
C	Possible human carcinogen.
D	Not classifiable as to human carcinogenicity.
E	Evidence of noncarcinogenicity for humans.

2.3.2.3 Generating a Slope Factor. In the second part of the evaluation, based on the evaluation that the chemical is a known or probable human carcinogen, a toxicity value that defines quantitatively the relationship between dose and response (i.e., the slope factor) is calculated. Slope factors are typically calculated for potential carcinogens in classes A, B1, and B2. Quantitative estimation of slope factors for the chemicals in class C is done on a case-by-case basis.

Generally, the slope factor is a plausible upper-bound estimate of the probability of a response per unit intake of a chemical over a lifetime. The slope factor is used in risk assessments to estimate an upper-bound lifetime probability of an individual developing cancer as a result of exposure to a particular level of a potential carcinogen. Slope factors should always be accompanied by the weight-of-evidence classification to indicate the strength of the evidence that the agent is a human carcinogen.

2.3.2.4 Identifying the Appropriate Data Set. In deriving slope factors, the available information about a chemical is evaluated and an appropriate data set is selected. In choosing appropriate data sets, human data of high quality are preferable to animal data. If animal data are used, the species that responds most similarly to humans (with respect to factors such as metabolism, physiology, and pharmacokinetics) is preferred. When no clear choice is possible, the most sensitive species is given the greatest emphasis. Occasionally, in situations where no single study is judged most appropriate, yet several studies collectively support the estimate, the geometric mean of estimates from all studies may be adopted as the slope. This practice ensures the inclusion of all relevant data.

2.3.2.5 Extrapolating to Lower Doses. Because risk at low exposure levels is difficult to measure directly either by animal experiments or by epidemiologic studies, the development of a slope factor generally entails applying a model to the available data set and using the model to extrapolate from the relatively high doses administered to experimental animals (or the

exposures noted in epidemiologic studies) to the lower exposure levels expected for human contact in the environment.

A number of mathematical models and procedures have been developed to extrapolate from carcinogenic responses observed at high doses to responses expected at low doses. Different extrapolation methods may provide a reasonable fit to the observed data but may lead to large differences in the projected risk at low dose.

In general, after the data are fit to the appropriate model, the upper 95th percent confidence limit of the slope of the resulting dose-response curve is calculated. This value is known as the slope factor and represents an upper 95th percent confidence limit on the probability of a response per unit intake of a chemical over a lifetime (i.e., there is only a 5 percent chance that the probability of a response could be greater than the estimated value on the basis of the experimental data and model used). In some cases, slope factors based on human dose-response data are based on the "best" estimate instead of the upper 95th percent confidence limits. Because the dose-response curve generally is linear only in the low-dose region, the slope factor estimate only holds true for low doses. Information concerning the limitations on use of slope factors can be found in IRIS.

2.3.2.6 Summary of Dose-Response Parameters. Toxicity values for carcinogenic effects can be expressed in several ways. The slope factor is generally considered to be the upper 95th percent confidence limit of the slope of the dose-response curve and is expressed as $(\text{mg/kg-day})^{-1}$. If the extrapolation model selected is the linearized multistage model, this value is also known as the q_1^* . Thus:

$$\begin{aligned}\text{Slope factor} &= \text{risk per unit dose} \\ &= \text{risk per mg/kg-day}\end{aligned}$$

Where data permit, slope factors listed in IRIS are based on absorbed doses, although many of them have been based on administered doses.

2.3.3 Summaries of the Toxicity of the Contaminants of Concern

Tables 2-5 and 2-6 present chronic cancer and noncancer health effects criteria (oral slope factors and RfDs, respectively, for the COCs). The toxicological properties of the COCs and the toxicological basis of the health effects criteria listed in Tables 2-5 and 2-6 are discussed in Appendix B.

2.4 RISK CHARACTERIZATION

In the risk characterization, the toxicity and exposure assessments are summarized and integrated into quantitative and qualitative expressions of risk. To characterize potential noncarcinogenic effects, comparisons are made between projected intakes of substances and toxicity values; to characterize potential carcinogenic effects, probabilities that an individual will develop cancer over a lifetime of exposure are estimated from projected intakes and chemical-

specific dose-response information. Major assumptions, scientific judgements, and to the extent possible, estimates of the uncertainties embodied in the assessment are also presented. In this section methods of quantifying risks are discussed and applied to individual sites on the Barter Island installation.

2.4.1 Quantifying Risks

This section describes steps for quantifying risk or hazard indices for both carcinogenic and noncarcinogenic effects to be applied to each exposure pathway analyzed. The first two subsections cover procedures for individual substances, and are followed by a subsection on procedures for quantifying risks associated with simultaneous exposures to several substances.

2.4.1.1 Risks from Individual Substances - Carcinogenic effects. For carcinogens, risks are estimated as the incremental probability of an individual developing cancer over a lifetime as a result of exposure to the potential carcinogen (i.e., incremental or excess individual lifetime cancer risk). The guidelines provided in this section are consistent with EPA (1986b). For some carcinogens, there may be sufficient information on mechanism of action that a modification of the approach outlined below is warranted. Alternative approaches may be considered in consultation with ECAO on a case-by-case basis.

The slope factor (SF) converts estimated daily intakes averaged over a lifetime of exposure directly to incremental risk of an individual developing cancer. Because relatively low intakes (compared to those experienced by test animals) are most likely from environmental exposures, it generally can be assumed that the dose-response relationship will be linear in the low-dose portion of the multistage model dose-response curve. Under this assumption, the slope factor is a constant, and risk will be directly related to intake. Thus, the linear form of the carcinogenic risk equation is usually applicable for estimating cancer risks. This linear low-dose equation is described below.

LINEAR LOW-DOSE CANCER RISK EQUATION

$$\text{Risk} = \text{LADD} \times \text{SF}$$

where:

Risk =	a unitless probability (e.g., 2×10^{-5}) of an individual developing cancer;
LADD =	lifetime average daily dose averaged over 70 years (mg/kg-day); and
SF =	slope factor, expressed in (mg/kg-day) ⁻¹

Because the slope factor is often an upper 95th percentile confidence limit of the probability of response based on experimental animal data used in the multistage model, the carcinogenic risk estimate will generally be an upper-bound estimate. This means that the "true risk" will probably not exceed the risk estimate derived through use of this model and is likely to be less than predicted.

2.4.1.2 Noncancer Hazards from Individual Substances - Noncarcinogenic Effects.

The measure used to describe the potential for noncarcinogenic toxicity to occur in an individual is not expressed as the probability of an individual suffering an adverse effect. EPA does not at the present time use a probabilistic approach to estimate the potential for noncarcinogenic health effects. Instead, the potential for noncarcinogenic effects is evaluated by comparing an exposure level over a specified time period (e.g., lifetime) with a reference dose derived for a similar exposure period. This ratio of exposure to toxicity is called a hazard quotient.

The noncancer hazard quotient assumes there is a level of exposure (i.e., RfD) below which it is unlikely for even sensitive populations to experience adverse health effects. If the exposure level (ADD) exceeds this threshold (i.e., if ADD/RfD exceeds unity), there may be concern for potential noncancer effects. As a rule, the greater the value of ADD/RfD above unity, the greater the level of concern. Ratios of ADD/RfD should not be interpreted as statistical probabilities; a ratio of 0.001 does not mean that there is a one in one thousand chance of the effect occurring. Further, it is important to emphasize that the level of concern does not increase linearly as the RfD is approached or exceeded because RfDs do not have equal accuracy or precision and are not based on the same severity of toxic effects. Thus, the slopes of the dose-response curve in excess of the RfD can range widely depending on the substance.

NONCANCER HAZARD QUOTIENT

$$\text{Noncancer Hazard Quotient} = \text{ADD/RfD}$$

where:

ADD = average daily dose (or intake);

RfD = reference dose

ADD and RfD are expressed in the same units and represent the same exposure period (i.e., chronic, subchronic, or short-term).

2.4.1.3 Aggregate Risks for Multiple Substances. Estimating risk or hazard potential by considering one chemical at a time might significantly underestimate the risks associated with simultaneous exposures to several substances. To assess the overall potential for cancer and noncancer effects posed by multiple chemicals, EPA has developed *Guidelines for the Health Risk Assessment of Chemical Mixtures* (EPA 1986b) that can also be applied to the case of simultaneous exposures to several chemicals from a variety of sources by more than one exposure pathway. Information on specific mixtures, however, is rarely available. Even if such data exist, they are often difficult to use. Monitoring for "mixtures" or modeling the movement of mixtures across space and time present significant technical problems given the likelihood that individual components will behave differently in the environment (i.e., fate and transport).

Although the calculation procedures differ for carcinogenic and noncarcinogenic effects, both sets of procedures assume dose additivity in the absence of information on specific mixtures.

Carcinogenic effects. The cancer risk equation described below is used to estimate the incremental individual lifetime cancer risk for simultaneous exposure to several carcinogens and is based on EPA's risk assessment guidelines. This equation represents an approximation of the precise equation for combining risks which accounts for the joint probabilities of the same individual developing cancer as a consequence of exposure to two or more carcinogens. The difference between the precise equation and the approximation described in the equation below is negligible for total cancer risks less than 0.1. Thus, the simple additive equation is appropriate for most risk assessments.

CANCER RISK EQUATION FOR MULTIPLE SUBSTANCES

$$\text{Risk}_T = \sum \text{Risk}_i$$

where:

Risk_T = the total cancer risk, expressed as a unitless probability; and

Risk_i = the risk estimate for the i^{th} substance.

The risk summation techniques described in the cancer risk equation on this page assume that intakes of individual substances are small. They also assume independence of action by the compounds involved (i.e., that there are no synergistic or antagonistic chemical interactions and that all chemicals produce the same effect, i.e., cancer). If these assumptions are incorrect, over- or under-estimation of the actual multiple-substance risk could result.

A separate total cancer risk for each exposure pathway is calculated by summing the substance-specific cancer risks. Resulting cancer risk estimates should be expressed using one significant figure only. The total cancer risk for each pathway should not exceed a value of 1.

There are several limitations to this approach. First, because each slope factor is an upper 95th percentile estimate of potency, and because upper 95th percentiles of probability distributions are not strictly additive, the total cancer risk estimate might become artificially more conservative as risks from a number of different carcinogens are summed. If one or two carcinogens drive the risk, however, this problem is not of concern. Second, it often will be the case that substances with different weights of evidence for human carcinogenicity are included. The cancer risk equation for multiple substances sums all carcinogens equally, giving as much weight to class B or C as to class A carcinogens. In addition, slope factors derived from animal data will be given the same weight as slope factors derived from human data. Finally, the action of two different carcinogens might not be independent.

Noncarcinogenic effects. To assess the overall potential for noncarcinogenic effects posed by more than one chemical, a hazard index (HI) approach has been developed based on EPA's *Guidelines for Health Risk Assessment of Chemical Mixtures* (EPA 1986b). This approach assumes that simultaneous subthreshold exposures to several chemicals could result in an adverse health effect. It also assumes that the magnitude of the adverse effect will be proportional to the sum of the ratios of the subthreshold exposures to acceptable exposures. The hazard index is equal to the sum of the hazard quotients. When the hazard index exceeds unity, there may be concern for potential health effects. Any single chemical with an exposure

level greater than the toxicity value (i.e., Hazard Quotient greater than unity) will cause the hazard index to exceed unity, but for multiple chemical exposures, the hazard index can also exceed unity even if no single chemical exposure exceeds its RfD. The equation used to determine noncancer hazard index is as follows:

NONCANCER HAZARD INDEX

$$\text{Hazard Index} = \text{ADD}_1/\text{RfD}_1 + \text{ADD}_2/\text{RfD}_2 + \dots + \text{ADD}_i/\text{RfD}_i$$

where:

ADD_i = average daily dose (or intake) for the i^{th} toxicant;

RfD_i = reference dose for the i^{th} toxicant; and

ADD and RfD are expressed in the same units and represent the same exposure period (i.e., chronic, subchronic, or shorter-term).

Where appropriate, a separate chronic hazard index can be calculated from the ratios of the chronic daily intake (CDI) to the chronic reference dose (RfD) for individual chemicals as described below.

CHRONIC NONCANCER HAZARD INDEX

$$\text{Chronic Hazard Index} = \text{LADD}_1/\text{RfD}_1 + \text{LADD}_2/\text{RfD}_2 + \dots + \text{LADD}_i/\text{RfD}_i$$

where:

LADD_i = lifetime average daily dose for the i^{th} toxicant in mg/kg-day,
and

RfD_i = chronic reference dose for the i^{th} toxicant in mg/kg-day.

There are several limitations to this approach. As mentioned earlier, the level of concern does not increase linearly as the reference dose is approached or exceeded because the RfDs do not have equal accuracy or precision and are not based on the same severity of effect. Moreover, hazard quotients are combined for substances with RfDs based on critical effects of varying toxicological significance. Also, it will often be the case that RfDs of varying levels of confidence that include different uncertainty adjustments and modifying factors will be combined (e.g., extrapolation from animals to humans, from LOAELs to NOAELs, from one exposure duration to another).

Another limitation with the hazard index approach is that the assumption of dose additivity is most properly applied to compounds that induce the same effect by the same mechanism of action. Consequently, application of the hazard index equation to a number of compounds that are not expected to induce the same type of effects or that do not act by the same mechanism

could overestimate the potential for effects, although such an approach is appropriate at a screening level. This possibility is generally not of concern if only one or two substances are responsible for driving the HI above unity. If the HI is greater than unity as a consequence of summing several hazard quotients of similar value, it would be appropriate to segregate the compounds by effect and by mechanism of action and to derive separate HIs for each group.

2.4.2 Site-Specific Risk Characterization

Soil and Sediment Exposures. The quantification of noncancer hazard and excess lifetime cancer risk associated with the soil ingestion pathway at Barter Island was based on analytical data from soil and sediment samples collected within the interval from ground surface to permafrost. No attempt was made to segregate surface soil samples from subsurface samples in the risk characterization.

The noncancer hazard and excess lifetime cancer risk associated with the ingestion of soil or sediment containing COCs has been estimated separately for a native northern adult, native northern child, and DEW Line worker. The noncancer hazard and the excess lifetime cancer risk associated with the ingestion of soil or sediment containing COCs has been estimated for an hypothetical native northerner based on six years of exposure as a child and 49 years of exposure as an adult. For the DEW Line worker, cancer risk has been estimated based on ten years of exposure averaged over a default lifetime of 70 years. Noncancer hazard for the DEW Line worker was based on a 10 year exposure.

Surface Water Exposures. The noncancer hazard and the excess lifetime cancer risk associated with the ingestion of surface water containing COCs has been estimated based on a native northern adult and a DEW Line worker. A native northern child receptor was not considered because, unlike exposure to soil, which is expected to be greater in a child than in an adult, exposure to drinking water is considered to be greater in the adult. Estimating hazard or risk for water ingestion based on an adult is, therefore, a more conservative approach. The exposure duration estimate for the DEW Line worker was 10 years and for the native northern adult, 55 years. Exposures were averaged over a 10 year period for DEW Line worker exposure to noncarcinogens, and 55 years for native northern adult exposure to noncarcinogens. Exposures were averaged over a 70 year period for both receptor groups to characterize the risk associated with exposure to carcinogens in surface water.

Ingestion of surface water by DEW Line workers or native northerners is considered to be only a potentially complete exposure pathway. The domestic water supply for the installation and the village of Kaktovik is Fresh Water Lake which is approximately one mile upgradient from the installation. Thus, under a current use scenario, surface water ingestion associated with any of the intermittent surface water sources at the installation is probably not complete (although, it is possible that children, for example, playing near the installation may stop for a drink of water from one of these surface water sources). Under a future use scenario, it is possible that these intermittent surface water sources might be used. Therefore, the exposure assumptions used to characterize the risk associated with surface water ingestion are based on residential use of surface water at the installation as a domestic water supply.

Table 2-8 contains a site-by-site summary of the COCs in each medium, and the noncancer hazard, and excess lifetime cancer risk associated with exposure to the soils/sediments and surface water. Table 2-8 does not include sites where no COCs were identified. Appendix A contains the spreadsheets used to calculate the noncancer hazard and excess lifetime cancer risk estimates presented in Table 2-8.

Risk Characterization of Petroleum Hydrocarbons. Petroleum hydrocarbons represent a primary source of contamination at the Barter Island installation. The laboratory analysis of soil, sediment, and surface water samples revealed the presence of DRPH, GRPH, and RRPH. In the process of characterizing the risk associated with exposure to these compounds, it was necessary to apply the provisional reference doses (RfDs) developed by EPA for petroleum hydrocarbons (EPA 1992d). These provisional RfDs provide the best available tool for characterizing the risk associated with exposure to the petroleum hydrocarbons. The RfD for JP-4 presented in EPA (1992d) was assumed to represent DRPH and RRPH, and the RfD and slope factor for unleaded gasoline was assumed to represent GRPH.

The noncancer hazard associated with exposure to DRPH, GRPH, and RRPH, was, therefore, estimated by dividing the compound- and site-specific ADD by the appropriate provisional RfD (EPA 1992d). The excess lifetime cancer risk associated with exposure to GRPH was estimated by multiplying the compound- and site-specific LADD by the slope factor for unleaded gasoline (EPA 1992d).

Although the provisional RfDs and slope factor represent the best available numerical estimate of toxicity, there is a significant amount of uncertainty associated with their use at the Barter Island installation. The RfDs and slope factor are based on studies in mice and rats by the inhalation route of exposure; whereas for this risk assessment, exposure of humans by the ingestion route of exposure is being evaluated. Furthermore, in the absence of a more thorough study to compare the DRPH, GRPH, and RRPH to known petroleum refinery streams, it is not clear how well the provisional toxicity values represent diesel and gasoline.

2.4.2.1 Old Landfill (LF01).

Soil and Sediments. No COCs were identified for the soil at the Old Landfill (Table 2-8). This does not indicate that exposure to chemicals in the soil at the site is without health risk, however, the concentrations measured were lower than the concentrations considered acceptable under Region 10 guidance (EPA 1991a) or federal ARARs.

Surface Water. The noncancer hazard associated with the ingestion of surface water at the Old Landfill by native northern adults and by DEW Line workers is 4.227, based on the maximum concentration of the COC (Tables 2-8 and A-1). The presence of manganese in surface water entirely accounts for the quantifiable noncancer hazard for this receptor/pathway combination. Other chemicals detected in the surface water at the site may be carcinogenic or produce noncancer effects, however, the lack of toxicity values (RfDs or slope factors) for these chemicals precludes the quantification of cancer risk or noncancer hazard.

TABLE 2-8. SUMMARY OF NONCANCER HAZARD AND EXCESS LIFETIME CANCER RISK FOR BARTER ISLAND

SITE	MEDIUM	CHEMICALS OF CONCERN ^a	SUMMARY OF NONCANCER HAZARD AND EXCESS LIFETIME CANCER RISK			COMMENTS
			DEW LINE WORKER	NATIVE NORTHERN ADULT	NATIVE NORTHERN ADULT/CHILD	
Old Landfill (LF01)	Soil	ND	ND ^b	ND	ND	No COCs selected for soils at Old Landfill.
	Surface Water	Manganese	HI=4.227 ^c CR=ND ^d	HI=4.227 CR=ND	HI=ND CR=ND	Manganese entirely accounts for noncancer hazard. Cancer risk not determined because no carcinogenic COCs were selected.
POL Catchment (LF03)	Soil	DRPH Tetrachloroethane	HI=0.021 CR=2e-9	ND ^e	HI=0.434 CR=5e-8	DRPH and tetrachloroethane account for noncancer hazard. Tetrachloroethane entirely accounts for cancer risk. Soil ingestion hazard and risk estimated for combined adult/child receptor.
	Surface Water	DRPH GRPH Benzene	HI=0.338 CR=1e-6	HI=0.338 CR=1e-5	HI=ND ^f CR=ND ^f	DRPH and GRPH account for noncancer hazard. GRPH and benzene account for cancer risk.
Current Landfill (LF04)	Soil	ND	ND	ND	ND	No COCs selected for soils or sediments at the Current Landfill.
	Surface Water	Manganese Trichloroethene	HI=5.072 CR=8e-7	HI=5.072 CR=6e-6	HI=ND CR=ND	Manganese entirely accounts for noncancer hazard. Trichloroethene entirely accounts for cancer risk.
Contaminated Ditch (SD08)	Soil	DRPH GRPH Beryllium	HI=0.002 CR=1e-7	ND	HI=0.036 CR=2e-6	DRPH and GRPH account for noncancer hazard. GRPH and beryllium account for cancer risk.
	Surface Water	ND	ND	ND	ND	No COCs selected for surface water at the Contaminated Ditch.

TABLE 2-8. SUMMARY OF NONCANCER HAZARD AND EXCESS LIFETIME CANCER RISK FOR BARTER ISLAND (CONTINUED)

SITE	MEDIUM	CHEMICALS OF CONCERN ^a	SUMMARY OF NONCANCER HAZARD AND EXCESS LIFETIME CANCER RISK			COMMENTS
			DEW LINE WORKER	NATIVE NORTHERN ADULT	NATIVE NORTHERN ADULT/CHILD	
Heated Storage Building (SS13)	Soil	DRPH GRPH RRPH Aroclor 1254	HI=0.012 CR=2e-7	ND	HI=0.258 CR=4e-6	All COCs identified here contribute to the noncancer hazard. GRPH and Aroclor 1254 (as PCB) account for cancer risk.
	Surface Water	DRPH Benzene Tetrachloroethane Manganese	HI=2.55 CR=2e-6	HI=2.55 CR=1e-5	HI=ND CR=ND	DRPH, tetrachloroethane, and manganese account for noncancer hazard. Benzene and tetrachloroethane account for cancer risk.
Garage (SS14)	Soil	DRPH GRPH RRPH Benzene Bis(2-ethylhexyl)phthalate	HI=0.029 CR=1e-8	ND	HI=0.602 CR=2e-7	DRPH, GRPH, RRPH and bis(2-ethylhexyl)phthalate account for noncancer hazard. GRPH, benzene and bis(2-ethylhexyl)phthalate account for cancer risk.
	Surface Water	ND	ND	ND	ND	No COCs selected for surface water at the Garage.
Weather Station Building (SS15)	Soil	DRPH GRPH	HI=0.006 CR=1e-8	ND	HI=0.134 CR=3e-7	DRPH and GRPH account for noncancer hazard. GRPH entirely accounts for cancer risk.
	Surface Water	ND	ND	ND	ND	No surface water bodies were identified at the Weather Station Building.
White Alice Facility (SS16)	Soil	PCBs	HI=0.153 CR=3e-6	ND	HI=3.155 CR=7e-5	Aroclor 1254 entirely accounts for noncancer hazard. Aroclor 1254 (as PCB) entirely accounts for cancer risk.
	Surface Water	ND	ND	ND	ND	No surface water bodies were identified at the White Alice Facility.

TABLE 2-8. SUMMARY OF NONCANCER HAZARD AND EXCESS LIFETIME CANCER RISK FOR BARTER ISLAND (CONTINUED)

SITE	MEDIUM	CHEMICALS OF CONCERN ^a	SUMMARY OF NONCANCER HAZARD AND EXCESS LIFETIME CANCER RISK			COMMENTS
			DEW LINE WORKER	NATIVE NORTHERN ADULT	NATIVE NORTHERN ADULT/CHILD	
POL Tanks (ST17)	Soil	DRPH GRPH	HI=0.001 CR=4e-9	ND	HI=0.027 CR=9e-8	DRPH and GRPH account for noncancer hazard. GRPH entirely accounts for cancer risk.
	Surface Water	ND	ND	ND	ND	No COCs selected for surface water at the POL Tanks site.
Old Dump Site (LF19)	Soil	DRPH RRPH	HI=0.005	ND	HI=0.097	DRPH and RRPH entirely account for noncancer hazard. No carcinogenic COCs were selected.
	Surface Water	ND	ND	ND	ND	No COCs selected for surface water at the Old Dump Site.
JP-4 Spill (SS21)	Soil	DRPH GRPH Benzene	HI=0.001 CR=5e-9	ND	HI=0.022 CR=1e-7	DRPH and GRPH account for noncancer hazard. GRPH and benzene account for cancer risk.
	Surface Water	ND	ND	ND	ND	No surface water bodies were identified at the JP-4 Spill site.

^a All COCs are listed together regardless of whether they contribute to the hazard index, cancer risk, or both.

^b ND, not determined

^c HI, noncancer hazard index. The HI is the sum of the hazard quotients for all of the COCs associated with a given medium, pathway, and receptor group.

^d CR, excess lifetime cancer risk. The CR is the sum of the excess lifetime cancer risks for all of the carcinogenic COCs associated with a given medium, pathway, and receptor group.

^e Children are assumed to have a soil ingestion rate greater than that for adults. Therefore, under a residential scenario, the estimates of noncancer hazard and cancer risk associated with soil ingestion are estimated for a combined adult and child receptor only. This estimate is considered a conservative upper bound on the true hazard or risk.

^f Drinking water ingestion, unlike soil ingestion, is evaluated for an adult receptor but not a child receptor because adults are assumed to have a longer exposure duration at a greater water ingestion rate. Therefore, the hazard or risk estimated will represent an upper bound, conservative estimate. For soil ingestion, the child soil ingestion rate is assumed to exceed that for adults. Therefore, a combination of the adult and child receptor groups is used to evaluate soil ingestion risk and hazard.

2.4.2.2 POL Catchment (LF03).

Soil and Sediment. The noncancer hazard associated with the ingestion of soil at the POL Catchment site by a hypothetical native northern adult/child is 0.434, and by DEW Line workers is 0.021, based on the maximum concentrations of the COCs (Tables 2-8 and A-2). The presence of DRPH accounts for more than 99 percent of the quantifiable noncancer hazard for these receptor/pathway combinations. The excess lifetime cancer risk associated with the ingestion of soil at the site by a hypothetical native northern adult/child is 5×10^{-8} , and by DEW Line workers is 2×10^{-9} , based on the maximum concentrations of the COCs (Tables 2-8 and A-3). The presence of tetrachloroethane entirely accounts for the quantifiable excess lifetime cancer risk for these receptor/pathway combinations. Other chemicals detected in the soil or sediment at the site may be carcinogenic or produce noncancer effects, however, the lack of toxicity values (RfDs or slope factors) for these chemicals precludes the quantification of cancer risk or noncancer hazard.

Surface Water. The noncancer hazard associated with the ingestion of surface water at the POL Catchment by native northern adults and DEW Line workers is 0.338, based on the maximum concentrations of the COCs (Tables 2-8 and A-4). The presence of DRPH and GRPH entirely accounts for the quantifiable noncancer hazard for these receptor/pathway combinations. The excess lifetime cancer risk associated with the ingestion of ground water at the site by native northern adults is 1×10^{-5} , and by DEW Line workers is 1×10^{-6} , based on the maximum concentrations of the COCs (Tables 2-8 and A-5). The presence of GRPH and benzene entirely accounts for the quantifiable excess lifetime cancer risk for these receptor/pathway combinations. Other chemicals detected in the surface water at the site may be carcinogenic or produce noncancer effects, however, the lack of toxicity values (RfDs or slope factors) for these chemicals precludes the quantification of cancer risk or noncancer hazard.

2.4.2.3 Current Landfill (LF04).

Soil and Sediments. No COCs were identified for the soil at the Current Landfill (Table 2-8). This does not indicate that exposure to chemicals in the soil at the site is without health risk, however, the concentrations measured were lower than the concentrations considered acceptable under Region 10 guidance (EPA 1991a) or federal ARARs.

Surface Water. The noncancer hazard associated with the ingestion of surface water at the Current Landfill by native northern adults and DEW Line workers is 5.072, based on the maximum concentration of the COC (Tables 2-8 and A-6). The presence of manganese entirely accounts for the quantifiable noncancer hazard for these receptor/pathway combinations. The excess lifetime cancer risk associated with the ingestion of surface water at the site by native northern adults is 6×10^{-6} , and by DEW Line workers is 8×10^{-7} , based on the maximum concentration of the COC (Tables 2-8 and A-7). The presence of trichloroethene entirely accounts for the quantifiable excess lifetime cancer risk for these receptor/pathway combinations. Other chemicals detected in the surface water at the site may be carcinogenic or produce noncancer effects, however, the lack of toxicity values (RfDs or slope factors) for these chemicals precludes the quantification of cancer risk or noncancer hazard.

2.4.2.4 Contaminated Ditch (SD08).

Soil and Sediment. The noncancer hazard associated with the ingestion of soil at the Contaminated Ditch by a hypothetical native northern adult/child is 0.036 and by DEW Line workers is 0.002, based on the maximum concentrations of the COCs (Tables 2-8 and A-8). The presence of DRPH accounts for more than 90 percent of the quantifiable noncancer hazard for these receptor/pathway combinations. The excess lifetime cancer risk associated with the ingestion of soil at the site by the hypothetical native northern adult/child is 2×10^{-6} , and by DEW Line workers is 1×10^{-7} , based on the maximum concentrations of the COCs (Tables 2-8 and A-9). The presence of GRPH and beryllium entirely accounts for the quantifiable excess lifetime cancer risk for these receptor/pathway combinations. Other chemicals detected in the soil or sediment at the site may be carcinogenic or produce noncancer effects, however, the lack of toxicity values (RfDs or slope factors) for these chemicals precludes the quantification of cancer risk or noncancer hazard.

Surface Water. No COCs were identified for surface water at the Contaminated Ditch (Table 2-8). This does not indicate that exposure to chemicals in surface water at the site is without health risk, however, the concentrations measured were lower than the concentrations considered acceptable under Region 10 guidance (EPA 1991a) or federal ARARs.

2.4.2.5 Old Runway Dump (LF12).

Soil and Sediments. No COCs were identified for the soil at the Old Runway Dump. This does not indicate that exposure to chemicals in the soil at the site is without health risk, however, the concentrations measured were lower than the concentrations considered acceptable under Region 10 guidance (EPA 1991a) or federal ARARs.

Surface Water. No surface water bodies were identified at the Old Runway Dump, therefore, no evaluation of noncancer hazard or excess lifetime cancer risk associated with ingestion of surface water was conducted.

2.4.2.6 Heated Storage Building (SS13).

Soil and Sediment. The noncancer hazard associated with the ingestion of soil at the Heated Storage Building site by hypothetical native northern adult/child is 0.258, and by DEW Line workers is 0.012, based on the maximum concentrations of the COCs (Tables 2-8 and A-10). The presence of DRPH, GRPH, RRPH and Aroclor 1254 entirely accounts for the quantifiable noncancer hazard for these receptor/pathway combinations. The excess lifetime cancer risk associated with the ingestion of soil at the site by the hypothetical native northern adult/child is 4×10^{-6} , and by DEW Line workers is 2×10^{-7} , based on the maximum concentrations of the COCs (Table 2-8 and A-11). The presence of PCBs accounts for more than 95 percent of the quantifiable excess lifetime cancer risk for these receptor/pathway combinations. Other chemicals detected in the soil or sediment at the site may be carcinogenic or produce noncancer effects, however, the lack of toxicity values (RfDs or slope factors) for these chemicals precludes the quantification of cancer risk or noncancer hazard.

Surface Water. The noncancer hazard associated with the ingestion of surface water at the Heated Storage Building site by native northern adults and DEW Line workers is 2.55, based on the maximum concentrations of the COCs (Tables 2-8 and A-12). The presence of DRPH and manganese accounts for more than 99 percent of the quantifiable noncancer hazard for these receptor/pathway combinations. The excess lifetime cancer risk associated with the ingestion of surface water at the site by native northern adults is 1×10^{-5} , and by DEW Line workers is 2×10^{-6} , based on the maximum concentrations of the COCs (Tables 2-8 and A-13). The presence of benzene and tetrachloroethane entirely accounts for the quantifiable excess lifetime cancer risk for these receptor/pathway combinations. Other chemicals detected in the surface water at the site may be carcinogenic or produce noncancer effects, however, the lack of toxicity values (RfDs or slope factors) for these chemicals precludes the quantification of cancer risk or noncancer hazard, or the concentrations at which these chemicals were measured were lower than the RBSLs or ARARs.

2.4.2.7 Garage (SS14).

Soil and Sediment. The noncancer hazard associated with the ingestion of soil at the Garage by a hypothetical native northern adult/child is 0.602 and by DEW Line workers is 0.029, based on the maximum concentrations of the COCs (Tables 2-8 and A-14). The presence of DRPH, GRPH, RRPB, and bis(2-ethylhexyl)phthalate accounts for the quantifiable noncancer hazard for these receptor/pathway combinations. The excess lifetime cancer risk associated with the ingestion of soil at the site by the hypothetical native northern adult/child is 2×10^{-7} , and by DEW Line workers is 1×10^{-8} , based on the maximum concentrations of the COCs (Tables 2-8 and A-15). The presence of GRPH accounts for more than 90 percent of the quantifiable excess lifetime cancer risk for these receptor/pathway combinations. Other chemicals detected in the soil or sediment at the site may be carcinogenic or produce noncancer effects, however, the lack of toxicity values (RfDs or slope factors) for these chemicals precludes the quantification of cancer risk or noncancer hazard, or the concentrations at which these chemicals were measured were lower than the risk-based screening levels or ARARs.

Surface Water. No COCs were identified for surface water at the Garage (Table 2-8). This does not indicate that exposure to chemicals in surface water at the site is without health risk, however, the concentrations measured were lower than the concentrations considered acceptable under Region 10 guidance (EPA 1991a) or federal ARARs.

2.4.2.8 Weather Station Building (SS15).

Soil and Sediment. The noncancer hazard associated with the ingestion of soil at the Weather Station Building by a hypothetical native northern adult/child is 0.134 and by DEW Line workers is 0.006, based on the maximum concentrations of the COCs (Tables 2-8 and A-16). The presence of DRPH and GRPH entirely accounts for the quantifiable noncancer hazard for these receptor/pathway combinations. The excess lifetime cancer risk associated with the ingestion of soil at the site by the hypothetical native northern adult/child is 3×10^{-7} , and by DEW Line workers is 1×10^{-8} , based on the maximum concentrations of the COCs (Tables 2-8 and A-17). The presence of GRPH entirely accounts for the quantifiable excess lifetime cancer risk for these receptor/pathway combinations. Other chemicals detected in the soil or sediment at the site may

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be carcinogenic or produce noncancer effects, however, the lack of toxicity values (RfDs or slope factors) for these chemicals precludes the quantification of cancer risk or noncancer hazard, or the concentrations at which these chemicals were measured were lower than the risk-based screening levels or ARARs.

Surface Water. No surface water bodies were identified at the Weather Station Building, therefore, no evaluation of noncancer hazard or excess lifetime cancer risk associated with the ingestion of surface water was conducted.

2.4.2.9 White Alice Facility (SS16).

Soil and Sediment. The noncancer hazard associated with the ingestion of soil at the White Alice Facility by a hypothetical native northern adult/child is 3.155, and by DEW Line workers is 0.153, based on the maximum concentration of the COC (Tables A-8 and A-18). The presence of Aroclor 1254 accounts for the quantifiable noncancer hazard for these receptor/pathway combinations.

The excess lifetime cancer risk associated with the ingestion of soil at the White Alice Facility by the hypothetical native northern adult/child is 7×10^{-5} , and by DEW Line workers is 3×10^{-6} , based on the maximum concentration of the COC (Tables 2-8 and A-19). The presence of PCBs (Aroclor 1254) entirely accounts for the quantifiable excess lifetime cancer risk for these receptor/pathway combinations.

Surface Water. No surface water bodies were identified at the White Alice Facility, therefore, no evaluation of noncancer hazard or excess lifetime cancer risk associated with ingestion of surface water was conducted.

2.4.2.10 POL Tanks (ST17).

Soil and Sediment. The noncancer hazard associated with the ingestion of soil at the POL Tanks site by a hypothetical native northern adult/child is 0.027, and by DEW Line workers is 0.001, based on the maximum concentrations of the COCs (Tables 2-8 and A-20). The presence of DRPH and GRPH entirely accounts for the quantifiable noncancer hazard for these receptor/pathway combinations. The excess lifetime cancer risk associated with the ingestion of soil at the site by the hypothetical native northern adult/child is 9×10^{-8} , and by DEW Line workers is 4×10^{-9} , based on the maximum concentration of the COC (Tables 2-8 and A-21). The presence of GRPH entirely accounts for the quantifiable excess lifetime cancer risk for these receptor/pathway combinations. Other chemicals detected in the soil or sediment at the site may be carcinogenic or produce noncancer effects, however, the lack of toxicity values (RfDs or slope factors) for these chemicals precludes the quantification of cancer risk or noncancer hazard, or the concentrations at which these chemicals were measured were lower than the risk-based screening levels or ARARs.

Surface Water. No surface water bodies were identified at the POL Tanks site, therefore, no evaluation of noncancer hazard or excess lifetime cancer risk associated with ingestion of surface water was conducted.

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workers is 5×10^{-9} , based on the maximum concentrations of the COCs (Tables 2-8 and A-24). The presence of GRPH and benzene entirely accounts for the quantifiable excess lifetime cancer risk for these receptor/pathway combinations. Other chemicals detected in the soil or sediment at the site may be carcinogenic or produce noncancer effects, however, the lack of toxicity values (RfDs or slope factors) for these chemicals precludes the quantification of cancer risk or noncancer hazard, or the concentrations at which these chemicals were measured were lower than the risk-based screening levels or ARARs.

Surface Water. No surface water bodies were identified at the JP-4 Spill, therefore, no evaluation of noncancer hazard or excess lifetime cancer risk associated with ingestion of surface water was conducted.

2.4.3 Summary and Conclusions

Eight of the fourteen sites evaluated in the remedial investigation pose noncancer hazards and excess lifetime cancer risks lower than the regulatory agency benchmarks of 1.0 for noncancer hazard quotient and 1×10^{-6} for excess lifetime cancer risk. These sites are the Old Runway Dump (LF12), Garage (SS14), Weather Station Building (SS15), POL Tanks (ST17), Fuel Tanks (ST18), Old Dump Site (LF19), Bladder Diesel Spill (SS20), and the JP-4 Spill (SS21). Of these eight, three were eliminated from further consideration in the risk assessment based on a comparison of the maximum concentrations of contaminants detected to the risk-based screening levels or ARARs or both. Thus, no COCs were selected for the Old Runway Dump (LF12), the Fuel Tanks (ST18), and the Bladder Diesel Spill (SS20) sites.

Table 2-9 presents a summary of the sites at the Barter Island installation and includes noncancer hazards and excess lifetime cancer risks for site and medium where the risks and/or hazards exceed the regulatory agency benchmark of 1.0 for noncancer hazard index and 1×10^{-6} for excess lifetime cancer risk.

The primary potential noncancer hazard at two of the six remaining sites is manganese at high concentrations in the surface water. These sites include the Old Landfill (LF01) and the Current Landfill (LF04). The noncancer hazard for manganese in surface water was calculated assuming the affected surface water would be used as a sole-source water supply for 180 days per year. Based on site-specific information, however, the manganese in surface water does not currently pose a health hazard nor is it likely to pose a hazard in the future. The surface water expressions at the installation are frozen most of the year (approximately nine months), are only intermittently filled with water during the warmer months, and are not known to be used for a water supply now or in the past. Water for domestic use at the Barter Island station and by native northerners is supplied by a fresh water lake located 1/2 to 3/4 of a mile upgradient from the station. The distance and gradient preclude the contamination of the fresh water lake by operations at the radar station. The purpose of including the surface water ingestion pathway is to evaluate a potential future exposure scenario that might include residential use of surface water sources other than Fresh Water Lake. If the Barter Island installation is retired and released for civilian use, it is possible that residential land use would occur. This situation is highly

TABLE 2-9. SUMMARY OF NONCANCER HAZARD AND EXCESS LIFETIME CANCER RISK FOR BARTER ISLAND SITES EXCEEDING REGULATORY THRESHOLDS [Hazard Index (HI) >1, Cancer Risk (CR) >1x10⁻⁶]

SITE	MEDIUM	CHEMICALS OF CONCERN ^a	SUMMARY OF NONCANCER HAZARD AND EXCESS LIFETIME CANCER RISK			COMMENTS
			DEW LINE WORKER	NATIVE NORTHERN ADULT	NATIVE NORTHERN ADULT/CHILD	
Old Landfill (LF01)	Surface Water	Manganese	HI = 4.227	HI=4.227	---	Manganese entirely accounts for noncancer hazard.
POL Catchment (LF03)	Surface Water	DRPH GRPH Benzene	---	CR=1e-5	---	GRPH and benzene account for cancer risk.
Current Landfill (LF04)	Surface Water	Manganese Trichloroethene	HI=5.072	HI=5.072 CR=6e-6	---	Manganese entirely accounts for noncancer hazard. Trichloroethene entirely accounts for cancer risk.
Contaminated Ditch (SD08)	Soil	DRPH GRPH Beryllium	---	---	CR=2e-6	GRPH and beryllium account for cancer risk.
Old Runway Dump (LF12)	Soil	None	---	---	---	Noncancer hazard index <1 and/or excess lifetime cancer risk <1x10 ⁻⁶ .
Heated Storage Building (SS13)	Soil	DRPH GRPH RRPH	---	---	CR=4e-6	GRPH and Aroclor 1254 (as PCB) account for cancer risk.
	Surface Water	Aroclor 1254/PCBs	---	---	---	
Garage (SS14)	Surface Water	DRPH GRPH Benzene Tetrachloroethane	HI=2.55 CR=2e-6	HI=2.55 CR=1e-5	---	DRPH, tetrachloroethane and manganese account for noncancer hazard. Benzene and tetrachloroethane account for cancer risk.
	Soil and Surface Water	DRPH GRPH RRPH Benzene Bis(2-ethylhexyl)phthalate Manganese	---	---	---	Noncancer hazard index <1 and/or excess lifetime cancer risk <1x10 ⁻⁶ .

TABLE 2-9. SUMMARY OF NONCANCER HAZARD AND EXCESS LIFETIME CANCER RISK FOR BARTER ISLAND SITES EXCEEDING REGULATORY THRESHOLDS [Hazard Index (HI) > 1, Cancer Risk (CR) > 1x10⁻⁶] (CONTINUED)

SITE	MEDIUM	CHEMICALS OF CONCERN ^a	SUMMARY OF NONCANCER HAZARD AND EXCESS LIFETIME CANCER RISK			COMMENTS
			DEW LINE WORKER	NATIVE NORTHERN ADULT	NATIVE NORTHERN ADULT/CHILD	
Weather Station Building (SS15)	Soil and Surface Water	DRPH GRPH	---	---	---	Noncancer hazard index <1 and/or excess lifetime cancer risk <1x10 ⁻⁶ .
White Alice Facility (SS16)	Soil	Aroclor 1254/PCBs	CR=3e-6	---	HI=3.155 CR=7e-5	Aroclor 1254 entirely accounts for noncancer hazard. Aroclor 1254 (as PCB) entirely accounts for cancer risk.
POL Tanks (ST17)	Soil and Surface Water	DRPH GRPH	---	---	---	Noncancer hazard index <1 and/or excess lifetime cancer risk <1x10 ⁻⁶ .
Fuel Tanks (ST18)	Soil and Surface Water	None	---	---	---	Noncancer hazard index <1 and/or excess lifetime cancer risk <1x10 ⁻⁶ .
Old Dump Site (LF19)	Soil	DRPH RRPH	---	---	---	Noncancer hazard index <1 and/or excess lifetime cancer risk <1x10 ⁻⁶ .
Bladder Diesel Spill (SS20)	Soil	None	---	---	---	Noncancer hazard index <1 and/or excess lifetime cancer risk <1x10 ⁻⁶ .
JP-4 Spill (SS21)	Soil and Surface Water	DRPH GRPH Benzene	---	---	---	Noncancer hazard index <1 and/or excess lifetime cancer risk <1x10 ⁻⁶ .

^a All COCs are listed together regardless of whether they contribute to the hazard index, cancer risk, or both.

unlikely given the long term plans to maintain the Barter Island installation as a long-range radar installation. In addition, it is highly unlikely the village of Kaktovik would move from its current location.

Surface water at two sites with excess lifetime cancer risks that slightly exceed regulatory agency benchmarks include the POL Catchment (LF03), 1×10^{-5} ; Current Landfill (LF04), 1×10^{-6} ; and Heated Storage (SS13), 1×10^{-5} . Again these risks are based on the surface water at these sites being used as a sole-source drinking water supply. Currently the surface water at these sites does not pose a carcinogenic risk.

The Contaminated Ditch (SD08) and Heated Storage (SS13) have cancer risks for soil ingestion that slightly exceed regulatory benchmarks. For soil ingestion at the Contaminated Ditch (SD08) the risk is 2×10^{-6} , and at the Heated Storage (SS13) the risk is 4×10^{-6} .

Except for the White Alice Facility, the noncancer hazard does not exceed one for the soil ingestion pathway at any site at the Barter Island installation. According to EPA, remedial action is generally not warranted at sites where the noncancer hazard is less than one (EPA 1991b). Therefore, on the basis of the estimated noncancer hazard, remediation of the soil at any site at the installation besides the White Alice Facility is not necessarily warranted.

The maximum excess lifetime cancer risk estimated for any site at the Barter Island installation is 7×10^{-5} and is associated with the potential exposure to PCBs through the soil ingestion pathway at the White Alice Facility (SS16). According to EPA, remedial action is generally not warranted at sites where the excess lifetime cancer risk is less than 1×10^{-4} (EPA 1991b). Therefore, on the basis of carcinogenic risk alone, remediation of the soil at any site at Barter Island is not necessarily warranted.

The excess lifetime cancer risk and hazard index for each of the sites was based on the maximum concentration of COCs detected at each site. This is a very conservative approach and had the average concentrations been used in the risk evaluation, few COCs would have been identified and risks and hazards would be below regulatory thresholds. The risk evaluation also assumed a future residential scenario where surface water at the sites would be used as a source of drinking water. Again, this is a very conservative approach as the streams and ponds at the sites are not likely to be used as a source of drinking water.

In conclusion, under current uses the COCs at the Barter Island sites pose only a minimal, if any, potential threat to human health. Based on the human health risk assessment, remedial actions/cleanup are not warranted at any of the 14 Barter Island sites.

2.5 RISK CHARACTERIZATION UNCERTAINTY

Several sources of uncertainty affect the estimates of excess lifetime cancer risk and noncancer hazard as presented in this risk assessment. The sources are generally associated with:

- Sampling and analysis of soil, sediment and surface water;

- Assigning the source of contamination;
- Exposure assumptions, including estimates of exposure point concentrations;
- Evaluation of the toxicity of the COCs; and
- Methods and assumptions used to characterize the cancer risk and noncancer hazard.

Uncertainties associated with sampling and analysis include the inherent variability (standard error) in the analysis, representativeness of the samples, sampling errors, and heterogeneity of the sample matrix. The quality assurance/quality control program used in conducting the sampling and analysis serves to reduce errors, but it can not eliminate all errors associated with sampling and analysis.

Simplifying assumptions were made about the environmental fate and transport of the site contamination, specifically, no contaminant loss or transformation has or will occur. Thus, the data chosen to represent exposure point concentrations in the sample-by-sample risk calculations is an additional source of potential error.

The estimation of exposure requires many assumptions to describe potential exposure situations. There are uncertainties regarding the likelihood of exposure, frequency of contact with contaminated media, the concentration of contaminants at exposure points, and the time period of exposure. These tend to simplify and approximate actual site conditions. In general, these assumptions are intended to be conservative and yield an overestimate of the true risk or hazard.

The toxicological database is also a source of uncertainty. The EPA has outlined some of the sources of uncertainty in EPA (1986a,b, and 1989a). These sources include extrapolation from high to low doses and from animals to humans; species, gender, age, and strain differences in uptake, metabolism, organ distribution, and target site susceptibility; and human population variability with respect to diet, environment, activity patterns, and cultural factors.

In the risk characterization, the assumption was made that the total risk of developing cancer from exposure to site contaminants is the sum of the risk attributed to each individual contaminant. Likewise, the potential for the development of noncancer adverse effects is the sum of the hazard quotients (previously defined) estimated for exposure to each individual contaminant. This approach does not account for the possibility that chemicals act synergistically or antagonistically.

In addition to the more general sources of uncertainty associated with risk assessment methodology, there are site-specific sources of uncertainty. Primarily, these sources are associated with the lifestyle of the native northerners (inhabitants of Kaktovik), the time spent on the 14 sites that were investigated during the RI, and with specific exposure assumptions (soil ingestion rate, exposure frequency, and exposure duration).

Inhabitants of Kaktovik are known to use the installation as an access route to the coast and occasionally for recreation (riding motorized vehicles). No studies have been conducted to measure the time these potential receptors spend on contaminated sites at the installation. Some of the sites with levels of contamination that exceed regulatory benchmarks are not likely to be accessed by this group (e.g., the White Alice facility, the Heated Storage Building, or the Garage). Therefore, the assumptions made regarding exposure frequency probably result in an overestimate of the true noncancer hazard and cancer risk.

Similarly, no studies have been conducted to measure the soil ingestion rate of potential receptors on the contaminated sites. Potentially, soil ingestion by the inhabitants of Kaktovik may be greater than the default rate of 100 mg/day for adults and 200 mg/day for children. Given the rugged, partially subsistence, lifestyle of this group, it is possible that they incidentally ingest soil at a higher rate than receptors of a similar age in the continental United States. The estimate of soil ingestion rate used in this risk assessment may over- or underestimate the true rate.

The maximum exposure duration assumed for native northerners, 55 years, is probably fairly accurate. The RME estimate for inhabitants of the continental United States is 30 years, however, native northerners are more likely to remain in their villages or nearby for a lifetime. Although, the exposure duration of 55 years is an estimate, it is not expected to significantly over- or underestimate hazard or risk.

3.0 ECOLOGICAL RISK ASSESSMENT

The objective of the ERA is to estimate potential impacts to aquatic and terrestrial plants and animals at the Barter Island DEW Line installation. This document assesses potential ecological risks at the Barter Island installation based on sampling and analyses conducted during the RI of the 14 sites located at the installation. The RI was completed during the summer of 1993 in conjunction with RIs at seven other radar installations.

Guidance documents used during preparation of this assessment include:

- Handbook to Support the Installation Restoration Program Statements of Work (Air Force 1991); and,
- Framework for Ecological Risk Assessment (EPA 1992a).

The approach used to assess potential ecological impacts is conceptually similar to that for human health risks; potentially exposed populations (receptors) are identified, and then information on exposure and toxicity are combined to derive estimates of risk. The ecological assessment focuses, however, on potential impacts on a population of organisms rather than on individual organisms (except in the case of endangered species where individuals are considered). Because ecosystems are composed of a variety of species, ecological assessments evaluate potential impacts to numerous species.

Ideally, ERAs should evaluate potential risks to communities and ecosystems, as well as to individual populations. Because of the large number of species and communities present in natural systems such ecosystem-wide assessments are very complex, however, appropriate assessment methodologies have not yet been developed. In addition, dose-response data on community or ecosystem responses are generally lacking. Therefore evaluations of potential impacts to communities or ecosystems are qualitative.

The degree to which potential ecological impacts can be characterized is highly dependent upon the data available to support such estimates. Such data include: information regarding contaminant release, transport, and fate; characteristics of potential receptor populations; and adequate supporting toxicity data for the COCs.

This ERA is intended to be at a screening level, rather than a full scale investigation of the state of the ecosystem. No specific studies of the biota were undertaken. The assessment is based on media sampling (i.e., surface water and soil/sediment samples). It is divided into six sections:

- Section 3.1 - Selection of Site Contaminants;
- Section 3.2 - Exposure Assessment;
- Section 3.3 - Ecological Toxicity Assessment;
- Section 3.4 - Risk Characterization for Ecological Receptors;
- Section 3.5 - Ecological Risk Assessment Uncertainty Analysis; and
- Section 3.6 - Summary of Ecological Risk.

3.1 SELECTION OF SITE CONTAMINANTS

A stressor in the environment is a chemical, physical or biological action that can cause a negative impact on an ecosystem (EPA 1992a). Only chemical stressors are evaluated as part of this ERA, and are identified as COCs. A review of the site data indicate that the chemical stressors are primarily petroleum products and solvents.

Only the sites that are thought to consist of useable habitat for potential receptors are evaluated for COCs in the ERA. Four of the 14 sites shown in Table 1-1 were eliminated from further evaluation because they are not likely to be used by ecological receptors. These sites include Weather Station Building (SS15), White Alice Facility (SS16), POL Tanks (ST17), and JP-4 Spill (SS21). The rationale for elimination and a discussion of habitat suitability are presented in Section 3.2.4. The other 10 sites, Old Landfill (LF01), POL Catchment (LF03), Current Landfill (LF04), Contaminated Ditch (SD08), Old Runway Dump (LF12), Heated Storage (SS13), Garage (SS14), Fuel Tanks (ST18), Old Dump Site (LF19), and Bladder Diesel Spill (SS20), appear to be useable habitat for representative species. Figures E-1 through E-16 in Appendix E illustrate habitat suitability at each of the sites.

COCs are chemicals detected above background concentrations and action levels, and at a high frequency. Chemicals present onsite at concentrations in excess of background concentrations and action levels (Ambient Water Quality Criteria for surface water; surface and ground water cleanup levels: AS 46.03.070; AS 46.04.020; AS 46.09.020; 18 AAC 70.020(b); and 18 AAC 75.140; and the ADEC determination of cleanup levels for petroleum contaminated soils, EPA sediment quality criteria) were evaluated for frequency of detection in onsite media. If a chemical was detected at a frequency of less than five percent, it was not considered representative of actual site conditions, and was eliminated from quantitative evaluation. If a chemical was detected only once it was eliminated from further evaluation, even if this violates the five percent screening criteria (e.g., 1 detect in 18 samples was screened out of the risk assessment). In certain cases infrequently detected chemicals were not eliminated as COCs; e.g., those with high potential for causing adverse ecological effects or those detected at a concentration significantly above the action level. If no action levels were available, the maximum detected concentration of the chemical was compared to a toxicity value derived from chronic exposure tests available in the literature. If the concentration was above this level, the compound was considered a COC. Tables 3-1 and 3-2 present the data used in the screening process for surface water and soils/sediment. These tables do not include data for sites that were eliminated from the ecological evaluation based on lack of suitable habitat.

In summary, the decisions for selecting COCs were made using the following logic:

- Is the chemical detected above the maximum detected background concentration?
 - No:** No longer considered a COC.
 - Yes:** Is the chemical detected above the action level or toxicity value?

TABLE 3-1. SUMMARY OF CHEMICALS OF CONCERN: SURFACE WATER

CHEMICALS OF CONCERN: BARTER ISLAND INSTALLATION SURFACE WATER						
CHEMICAL	RANGE OF DETECTED CONCENTRATION (µg/L)	BACKGROUND CONCENTRATION (µg/L)	ACTION LEVEL (µg/L)	FREQUENCY OF DETECTION	AVERAGE CONCENTRATION FOR COC (µg/L)	COC
ORGANICS						
DRPH	196-5,760	<200	---	8/28	719	YES
GRPH	6.9-367	<20.0	---	2/24	44	NO
Benzene	2.7-6.9	<1	106 ^a	2/24	NA	NO
Toluene	56	<1	350 ^a	1/24	NA	NO
Ethylbenzene	3.1-19	<1	640 ^a	3/24	NA	NO
Xylenes (Total)	3.8-38.4	<2	66 ^b	5/24	NA	NO
Chloromethane	4.2	<1	2,700-5,500 ^c	1/18	.59	NO
n-Butylbenzene	2.4	<1	---	1/14	.64	NO
sec-Butylbenzene	1.0	<1	---	1/14	.54	NO
1,2-Dichloroethane	9.1	1.3B-3.2B	20,000 ^d	1/18	NA	NO
cis-1,2-Dichloroethene	1.5	<1	70 ^e	1/14	NA	NO
Dichlorofluoromethane	3.7	<1	---	1/14	.69	NO
Isopropylbenzene	2.9	<1	---	1/14	.67	NO
1,2,4-Trimethylbenzene	1.4-19	<1	77.2 ^f	2/14	1.89	NO
1,3,5-Trimethylbenzene	1.3-13	<1	77.2 ^f	2/14	1.45	NO
Naphthalene	1.6-35	<1	620 ^e	3/15	NA	NO
n-Propylbenzene	3.6	<1	---	1/14	.72	NO

NA Not applicable.
 --- Not available.
 a Federal Ambient Water Quality Criteria, Fresh Acute Criteria (divided by an uncertainty conversion factor of 50).
 b 96 hour LC50 for Rainbow Trout divided by an uncertainty conversion factor of 50.
 c Dawson et. al. 1977.
 d Drinking water health advisory for reference exposure - human health.
 e Federal Ambient Water Quality Criteria, Fresh chronic criteria.
 f Geiger 1986.

TABLE 3-1. SUMMARY OF CHEMICALS OF CONCERN: SURFACE WATER (CONTINUED)

CHEMICALS OF CONCERN: BARTER ISLAND INSTALLATION SURFACE WATER					
CHEMICAL	RANGE OF DETECTED CONCENTRATION (µg/L)	BACKGROUND CONCENTRATION (µg/L)	ACTION LEVEL (µg/L)	FREQUENCY OF DETECTION	AVERAGE CONCENTRATION FOR COC (µg/L)
Tetrachloroethene	12	<1	840 ^e	1/20	NA
p-Isopropyltoluene	1.1-2.1	<1	65 ^g	2/14	.7
Trichloroethene	6-36	<1	21,900 ^e	2/20	NA
INORGANICS ^j					
Aluminum	160-1,400	<100-350	87 ^e	5/10	318
Barium	74-150	<50-93	1,000 ^h	9/10	90
Calcium	5,000-190,000	4,100-88,000	---	10/10	104,820
Iron	1,600-21,000	<100-2,800	1,000 ^e	10/10	8,580
Magnesium	26,000-78,000	<5,000-54,000	---	10/10	43,900
Manganese	70-1,800	<50-510	200 ⁱ	10/10	574
Potassium	5,100-110,000	<5,000	See text	8/10	18,040
Sodium	5,000-440,000	8,200-450,000	---	10/10	NA
					NO

NA Not applicable.
 --- Not available.
 e Federal Ambient Water Quality Criteria, Fresh Chronic Criteria.
 g LeBlanc 1980.
 h Based on Arsenic (III); no criteria for Arsenic (total).
 i Maximum contaminant level (MCL).
 j Total metals.

TABLE 3-2. SUMMARY OF CHEMICALS OF CONCERN: SOILS AND SEDIMENTS

CHEMICALS OF CONCERN: BARTER ISLAND INSTALLATION SEDIMENT AND SOIL						
CHEMICAL	RANGE OF CONCENTRATION (mg/kg)	BACKGROUND CONCENTRATION (mg/kg)	ACTION LEVEL (mg/kg) ^a	FREQUENCY OF DETECTION	AVERAGE CONCENTRATION FOR COC (mg/kg)	COC
ORGANICS						
DRPH	4.76-28,600	9.55-1,150	500 ^b	39/91	1159.2	YES
GRPH	0.429-700	<0.4-<9	100 ^b	30/88	34.4	YES
RRPH (Approx.)	180-27,000	<480	2,000 ^b	10/59	968	YES
Benzene	0.72-1.4	<0.02-<0.3	0.5 ^b /0.05 ^c	3/86	NA	NO
Toluene	0.026-2.3	<0.02-<0.3	0.786	13/86	0.093	NO
Ethylbenzene	0.02-11	<0.02-<0.3	4.36 ^c	19/86	0.13	YES
Xylenes (Total)	0.21-47	<0.04-<0.6	1.21	25/86	1.09	YES
n-Butylbenzene	0.635-4.22	<0.500	4.36 ^d	4/30	0.41	NO
sec-Butylbenzene	1.24-1.62	<0.500	4.36 ^d	3/30	0.36	NO
tert-Butylbenzene	.256	<0.500	---	1/30	NA	NO
cis-1,2-Dichloroethene	0.069	<0.500	---	1/30	NA	NO
1,2,4-Trichlorobenzene	0.046	<0.500	---	1/30	NA	NO
1,2,4-Trimethylbenzene	0.041-14.7	<0.500	---	5/30	0.98	YES
1,3,5-Trimethylbenzene	0.048-9.32	<0.500	---	7/30	0.93	YES
1,4-Dichlorobenzene	0.044	<0.500	---	1/48	NA	NO
Isopropylbenzene	0.409-0.681	<0.500	4.36 ^d	3/30	0.08	NO
p-Isopropyltoluene	1.69-2.47	<0.500	4.36 ^d	3/30	0.23	NO

NA

Not applicable.

Not available.

NOAA 1991, Sediment ER-L (Effects Range-Low).

ADEC, Interim Guidance for non-UST contaminated soil cleanup levels, 17 July 1991.

EPA Sediment Quality Criteria.

EPA Sediment Quality Criteria for Ethylbenzene (see text).

55 FR 30798-Proposed Rule RCRA Corrective Action for SWMVs 40 CFR [Section 264.521(m)(2)(i-iv)] Health-Based Criteria for Carcinogens.

TABLE 3-2. SUMMARY OF CHEMICALS OF CONCERN: SOILS AND SEDIMENTS (CONTINUED)

CHEMICALS OF CONCERN: BARTER ISLAND INSTALLATION SEDIMENT AND SOIL						
CHEMICAL	RANGE OF CONCENTRATION (mg/kg)	BACKGROUND CONCENTRATION (mg/kg)	ACTION LEVEL (mg/kg) ^a	FREQUENCY OF DETECTION	AVERAGE CONCENTRATION FOR COC (mg/kg)	COC
n-Propylbenzene	0.918-1.17	<0.500	4.36 ^d	3/30	0.097	NO
Methylene Chloride	0.225	<0.300	0.43 ^c	1/48	NA	NO
Tetrachloroethene	0.023-5.42	<0.300	---	2/62	NA	NO
Isophorone	1.49	<0.23	---	1/18	0.39	NO
Fluoranthene	1.7-2.28	<0.23	6.2 ^c	2/18	0.46	NO
Naphthalene	0.105-46	<0.500	0.34	5/30	2.26	YES
Phenanthrene	1.96-4.79	<0.23	1.8 ^c	2/18	0.62	NO
Pyrene	1.26	<0.23	13.1 ^c	1/18	0.87	NO
2-Methylnaphthalene	3.44-14.5	<0.23	.065	3/18	1.55	YES
Bis(2-ethylhexyl)phthalate	4.6	<0.23	50 ^e	1/18	NA	NO
INORGANICS						
Aluminum	1,600-7,500	1,500-25,000	---	15/16	NA	NO
Barium	13-120	27-390	---	15/16	NA	NO
Beryllium	3.2	<2.6-6.4	---	1/16	NA	NO
Cadmium	2.7-2.8	<3.0 - <36	5	2/16	NA	NO
Calcium	2,200-33,000	360-59,000	---	16/16	NA	NO
Chromium	3-53	<4.3-47	80	16/16	NA	NO
Iron	4,700-17,000	5,400-35,000	---	16/16	NA	NO
Lead	6.1-231	<5.1-22	35	8/16	26.6	YES
Magnesium	990-17,000	360-7,400	---	16/16	3,528	NO

NA Not applicable.

c Not available.

d EPA Sediment Quality Criteria.

e EPA Sediment Quality Criteria for Ethylbenzene (see text).

40 CFR 307.98-Proposed Rule RCRA Corrective Action for SWMUs, 40 CFR [Section 201.521(m)(2)(i-iv)] Health-Based Criteria for Carcinogens.

TABLE 3-2. SUMMARY OF CHEMICALS OF CONCERN: SOILS AND SEDIMENTS (CONTINUED)

CHEMICALS OF CONCERN: BARTER ISLAND INSTALLATION SEDIMENT AND SOIL						
CHEMICAL	RANGE OF CONCENTRATION (mg/kg)	BACKGROUND CONCENTRATION (mg/kg)	ACTION LEVEL (mg/kg) ^a	FREQUENCY OF DETECTION	AVERAGE CONCENTRATION FOR COC (mg/kg)	COC
Manganese	40-380	25-290	---	16/16	106	NO
Nickel	4.3-16	4.2-46	30	16/16	NA	NO
Potassium	300-610	<300-2200	---	9/16	NA	NO
Sodium	38-870	<160-680	---	16/16	132	NO
Vanadium	4.9-21	6.3-59	---	15/16	NA	NO
Zinc	12-500	9.2-95	120	16/16	76	YES
PCBs (Arochlor 1254)	0.112-2.72	<0.02-<0.1	10	5/26	0.42	YES

NA Not applicable.
 I Not available.
 c EPA Sediment Quality Criteria.
 d EPA Sediment Quality Criteria for Ethylbenzene (see text).
 e 55 FR 30798-Proposed Rule RCRA Corrective Action for SWMUs, 40 CFR [Section 264.521 (m) (2) (i-iv)] Health-Based Criteria for Carcinogens.

- No:** No longer considered a COC.
Yes: Does the chemical have a frequency of detection greater than five percent?
- No:** No longer considered a COC.
Yes: Is the chemical detected more than once?
- No:** No longer considered a COC.
(Unless chemical has high potential for ecological effects or the one detection was at concentrations significantly above the action level.)
Yes: Chemical is classified as a COC.
(Unless chemical is an essential nutrient.)

All data for COCs were averaged according to media (arithmetic mean). In the case of non-detects averages were calculated for organic compounds and metals using one-half of the quantitation limits. Total metal concentrations were used in determining COCs in surface water. This is a conservative approach because dissolved metal concentrations are generally significantly less than total metal concentrations. Section 3.1.1 describes surface water COCs. Section 3.1.2 describes soil and sediment COCs.

3.1.1 Surface Water

Analytical results from the 10 sites were compiled and evaluated to determine the COCs. Surface water samples were collected and analyzed for contaminants likely to be present at the specific sites. Not all samples were analyzed for a "full suite" of parameters, but instead were analyzed for some combination of the following: DRPH, GRPH, RRPB, BTEX volatile organic compounds (VOCs), semivolatile organic compounds, PCBs, pesticides, and metals. Complete analytical results for all sampling conducted at the installation are presented in Appendix D. The following sections present the evaluation of the surface water data, and Table 3-1 summarizes the screening results.

3.1.1.1 Petroleum Hydrocarbons. Thirty surface water samples were collected from the ten areas identified above and selectively analyzed for a combination of DRPH, GRPH, and RRPB. A discussion of these petroleum hydrocarbon mixtures and their toxicity is presented in Section 3.3.1.

DRPH were detected in 8 of 28 surface water samples, ranging from 196 to 5,760 $\mu\text{g/L}$. The background concentration is $<200 \mu\text{g/L}$. Because DRPH were detected in some samples at concentrations in excess of the action level and background concentrations, DRPH are retained for further analysis in the ERA. The exposure concentration used is 719 $\mu\text{g/L}$, the average concentration of DRPH in surface water.

GRPH were detected in 2 of 24 surface water samples, ranging from 6.9 to 367 $\mu\text{g/L}$. The background concentration is $<20.0 \mu\text{g/L}$. Because GRPH were detected in only two surface

water samples that contained significantly higher levels of DRPH, and are not considered independently of DRPH (see discussion of petroleum toxicity, Section 3.3.1), GRPH are not considered a COC.

RRPH were not detected in any surface water samples collected from the 10 sites.

3.1.1.2 Benzene, Toluene, Ethylbenzene and Xylenes.

Benzene was detected in 2 of 24 surface water samples at levels of 2.7 and 6.9 µg/L. The background concentration is <1 µg/L. Onsite concentrations are well below the action level of 106 µg/L (ambient water quality criteria divided by an uncertainty factor of 50), so benzene is not considered a COC.

Toluene was detected in 1 of 24 surface water samples at a concentration of 56 µg/L. The background concentration for toluene in surface water is <1 µg/L. Onsite concentrations do not exceed the action level of 350 µg/L, so it is not considered a COC.

Ethylbenzene was detected in 3 of 24 surface water samples at concentrations ranging from 3.1 to 19 µg/L. The background concentration of ethylbenzene in surface water is <1 µg/L. Onsite concentrations of ethylbenzene are well below the action level of 640 µg/L, so ethylbenzene is not considered a COC.

Xylene was detected in 5 of 24 samples. Xylene concentrations ranged from 3.8 to 38.4 µg/L. The background concentration of xylene is <2 µg/L. Onsite concentrations are well below the action level of 66 µg/L, so xylene is not considered a COC.

3.1.1.3 Other Organic Compounds. Fourteen additional organic compounds were detected in surface water samples collected from Barter Island. The compounds detected were: chloromethane; n-butylbenzene; sec-butylbenzene; 1,2-dichloroethane; cis-1,2-dichloroethene; dichlorofluoromethane; isopropylbenzene; 1,2,4-trimethylbenzene; 1,3,5-trimethylbenzene; naphthalene; n-propylbenzene; tetrachloroethene; p-isopropyltoluene; and trichloroethene. This section presents the evaluation of these compounds as COCs in surface water for the ERA.

Chloromethane was detected in 1 of 18 surface water samples analyzed for VOCs, at a concentration of 4.2 µg/L. The background concentration of chloromethane is <1 µg/L. Because no action levels are available for this compound, and it was detected at a frequency greater than five percent, the maximum detected concentration is compared to toxicity values reported in the literature. This compound is relatively non-toxic to aquatic organisms. Toxicity values (LC₅₀s) were 270,000 µg/L for *Menidia berylina* (inland silverside) and 550,000 µg/L for *Lepomis macrochirus* (Blue gill) (Dawson et al. 1977). Applying a conservative uncertainty factor of 100 (Suter 1991) yields action levels ranging from 2,700 to 5,500 µg/L. The detected concentration is well below this toxicity-based action level range, so chloromethane is not considered a COC.

n-Butylbenzene was detected in 1 of 14 surface water samples at a concentration of 2.4 µg/L. The background concentration for this compound is <1 µg/L. There are no action levels for n-butylbenzene. This compound was detected in only one sample, so it is not considered a COC.

sec-Butylbenzene was detected in 1 of 14 surface water samples at a concentration of 1.0 µg/L. The background concentrations for this compound is <1 µg/L. There are no action levels for sec-butylbenzene. This compound was detected in only one sample, so it is not considered a COC.

1,2-Dichloroethane was detected in 1 of 18 surface water samples collected at the Barter Island facility at a concentration of 9.1 µg/L. The background concentrations of 1,2-dichloroethane in surface water are 1.3 to 3.2 µg/L, however at least some of the blanks for the background samples also had detections. The action level of this compound is 20,000 µg/L. 1,2-dichloroethane was detected at a level substantially less than the action level, so this compound is not considered a COC.

cis-1,2-Dichloroethene was detected in 1 of 14 surface water samples. It was detected at concentrations of 1.4 and 1.5 µg/L in replicate samples obtained from the Contaminated Ditch (SD08). The background concentration of this compound is <1 µg/L. The action level is 70 µg/L. Because cis-1,2-dichloroethene was detected below action levels, it is not considered a COC.

Dichlorofluoromethane was detected in 1 of 14 samples at 3.7 µg/L. The background concentration of this compound is <1 µg/L. There are no action levels. This chemical was detected in only one sample, so it is not considered a COC.

Isopropylbenzene was detected in 1 of 14 surface water samples at a concentration of 2.9 µg/L. The background concentration is <1 µg/L. There are no action levels for this compound. This compound was detected in only one sample, so it is not considered a COC.

1,2,4-Trimethylbenzene was detected in 2 of 14 surface water samples at concentrations of 1.4 and 19 µg/L. The background concentration is <1 µg/L, and no action levels are established. It was detected at a frequency greater than five percent, so the maximum detected concentration is compared to toxicity values reported in the literature. There is not a great deal of information regarding the aquatic toxicity of 1,2,4-trimethylbenzene. One study, however, derived a 96 hr LC₅₀ of 7,720 µg/L for *Pimephales promelas* (fathead minnow) (Geiger 1986). Applying a conservative uncertainty factor of 100 (Suter 1991) yields an action level of 77.2 µg/L. The maximum detected concentration is below this toxicity-based action level, so 1,2,4-trimethylbenzene is not considered a COC.

1,3,5-Trimethylbenzene was detected in 2 of 14 surface water samples at concentrations of 1.3 and 13 µg/L. 1,3,5-Trimethylbenzene was not detected in the background samples. For the purposes of this ERA, the two isomers of trimethylbenzene are considered similar. Thus, the action level for this compound is 77.2 µg/L. The maximum detected concentration is below the action level, so 1,3,5-trimethylbenzene is not considered a COC.

Naphthalene was detected in 3 of 15 surface water samples at concentrations ranging from 1.6 to 35 µg/L. The background concentration is <1 µg/L. The action level established for this compound is 620 µg/L. This compound was detected at levels lower than the action level, and it is not considered a COC.

n-Propylbenzene was detected in 1 of 14 surface water samples at 3.6 µg/L. There are no action levels established for n-propylbenzene. The background concentration is <1 µg/L. This compound was detected in only one sample, so it is not considered a COC.

Tetrachloroethene was detected in 1 of 20 samples at 12 µg/L. This compound was not detected in background samples (<1 µg/L). The action level for this compound is 840 µg/L. It was detected in only one sample at concentrations below the action level, so it is not considered a COC.

p-Isopropyltoluene was detected in 2 of 14 surface water samples at concentrations of 1.1 and 2.1 µg/L. The background concentration is <1 µg/L. No action level is available for this compound. It was detected at a frequency greater than five percent, so the maximum detected concentration is compared to toxicity values reported in the literature. There is not a great deal of information regarding the aquatic toxicity of p-isopropyltoluene. One study derived a 48 hr LC₅₀ of 6,500 µg/L for *Daphnia magna* (LeBlanc 1980). Applying a conservative uncertainty factor of 100 (Suter 1991) yields an action level of 65 µg/L. Because the maximum detected concentration is below this toxicity-based action level, p-isopropyltoluene is not considered a COC.

Trichloroethene was detected in 2 of 20 samples at 6 and 36 µg/L. It was not detected in background samples (<1 µg/L). The action level for this compound is 21,900 µg/L. This compound was detected below action levels, so it is not considered a COC.

3.1.1.4 Metals. Eight inorganic analytes were detected in surface water samples collected from Barter Island. The metals detected were: aluminum, barium, calcium, iron, magnesium, manganese, potassium and sodium. This section presents the evaluation of these metals as COCs for the ERA. Analytes not detected in surface water samples were: antimony, beryllium, cadmium, chromium, lead, molybdenum, nickel, selenium, silver, thallium, vanadium and zinc.

Aluminum was detected in 5 of 10 surface water samples. Concentrations ranged from 160 to 1,400 µg/L. Background concentrations ranged from <100 to 350 µg/L. The action level for aluminum is 87 µg/L. Aluminum is present in surface water at concentrations in excess of the action level, so it is retained as a COC. The exposure concentration evaluated in this ERA is the average concentration of 318 µg/L.

Barium was detected in 9 of 10 surface water samples. Concentrations ranged from 74 to 150 µg/L. Background concentrations of barium ranged from <50 to 93 µg/L. The action level is 1,000 µg/L. Although barium concentrations exceed background, this chemical was not retained as a COC because it does not exceed the action level.

Calcium was detected in all 10 surface water samples. Concentrations ranged from 5,000 to 190,000 µg/L. Background concentrations ranged from 4,100 to 88,000 µg/L. There is no action level for calcium, and toxicity information is limited for ecological receptors. This chemical was not retained as a COC because it is ubiquitous in the environment and not expected to present

a risk to ecological receptors. In addition, this mineral is an essential nutrient and is well regulated by plants and animals.

Iron was detected in all 10 surface water samples. Concentrations ranged from 1,600 to 21,000 µg/L. Background concentrations ranged from <100 to 2,800 µg/L. Iron exceeds the background concentration in surface water and the 1,000 µg/L action level, so this metal was retained as a COC. The exposure concentration evaluated in this ERA is the average concentration of 8,580 µg/L.

Magnesium was detected in all 10 surface water samples. Concentrations ranged from 26,000 to 78,000 µg/L. Background concentrations ranged from <5,000 to 54,000 µg/L. There are no action levels for magnesium. Magnesium is not retained as a COC in the ERA because it is ubiquitous in the environment and an essential nutrient which is regulated by plants and animals.

Manganese was detected in all 10 surface water samples. Concentrations ranged from 70 to 1,800 µg/L. Background concentrations ranged from <50 to 510 µg/L. The action level for manganese is 200 µg/L. Because manganese was detected at concentrations in excess of the background and action levels, it is retained as a COC. The exposure concentration evaluated in this ERA is the average concentration of 574 µg/L.

Potassium was detected in 8 of 10 surface water samples. Concentrations ranged from 5,100 to 110,000 µg/L, which all exceeded the background concentration of <5,000 µg/L. There is no action level for potassium, so the maximum detected concentration was compared to toxicity values in the literature. As potassium is an essential nutrient, regulated by plants and animals, toxicity information was sparse. Concentrations of potassium at 391,000 µg/L were found to inhibit enzymes in freshwater algae (Matson et al. 1972). Potassium was toxic to the freshwater amphipod (*Gammarus lacustris*) at concentrations of 53,200 µg/L (De March 1988). There is some discrepancy between these toxicity values, so a review of the analytical data was conducted. The maximum detected concentration of potassium was found in a waterway that is connected to the Beaufort Sea (LF01-SW04). This sample location lies in a waterway that drains to the ocean approximately 300 feet away. The concentration of potassium in seawater is 390,000 µg/L [U.S. Geological Survey (USGS) 1985]. Because this sample location is hydrologically connected to the ocean, it is likely that these elevated levels of potassium are attributable to the influence of saltwater. The average concentration of potassium onsite is 18,040 µg/L (this includes the maximum detected concentration). This concentration is within the normal range of potassium concentrations in surface waters (USGS 1985). As a result of this review, potassium was not included as a COC.

Sodium was detected in all 10 surface water samples. Concentrations ranged from 5,000 to 440,000 µg/L, exceeding the background concentrations of 8,200 to 450,000 µg/L. There is no action level for sodium. Sodium is ubiquitous in the environment and is not expected to present a threat to ecological receptors. Thus, sodium is not considered a COC for the ERA.

3.1.2 Soils and Sediments

The following sections present the evaluation of soil and sediment data for the 10 sites. Table 3-2 summarizes the screening results for soils and/or sediment.

3.1.2.1 Petroleum Hydrocarbons. Ninety-two soil and/or sediment samples were collected from the ten sites and selectively analyzed for a combination of DRPH and GRPH. Fifty-nine soil and/or sediment samples were collected and analyzed for RRPB. A discussion of these petroleum hydrocarbon mixtures and their toxicity is presented in Section 3.3.1.

DRPH were detected in 39 of 91 soil and/or sediment samples ranging from 4.76 to 28,600 mg/kg; the background concentrations were 9.55 to 1,150 mg/kg. The action level for DRPH in soils/sediments is 500 mg/kg. Because 22 percent of onsite samples exceeded the action level, DRPH are retained as a COC. The exposure concentration used in the risk assessment is the average concentration of 1,159.2 mg/kg.

GRPH were detected in 30 of 88 soil and/or sediment samples ranging from 0.429 to 700 mg/kg. The background concentrations ranged from <0.4 to <9 mg/kg. The action level for GRPH is 100 mg/kg. GRPH were detected at concentrations above action levels, so they are considered a COC.

RRPB were detected in 10 of 59 samples ranging from 180 to 27,000 mg/kg. The background concentration for RRPB is <480 mg/kg; the action level is 2,000 mg/kg. RRPB were detected at concentrations above action levels, so they are considered a COC.

3.1.2.2 Benzene, Toluene, Ethylbenzene and Xylenes.

Benzene was detected in 3 of 86 soil and/or sediment samples at concentrations of 0.72 to 1.4 mg/kg. Although the action level of benzene is 0.5 mg/kg, and background concentration is <0.3 mg/kg, benzene is not considered a COC because of the low frequency of detection (three percent).

Toluene was detected in 13 of 86 soil and/or sediment samples ranging from 0.026 to 2.3 mg/kg. The background concentration is <0.3 mg/kg. Only one detected concentration exceeded the action level (0.786 mg/kg). The detected toluene above the action level is at the garage in a gravel pad not expected to be frequented by ecological receptors. Therefore, toluene is not considered to be a threat to ecological receptors and is not considered a COC.

Ethylbenzene was detected in 19 of 86 soil and/or sediment samples at concentrations ranging from 0.02 to 11 mg/kg. The background concentration of ethylbenzene is <0.3 mg/kg for soil and sediment; the action level is 4.36 mg/kg. Onsite concentrations exceed action levels, so ethylbenzene is considered a COC. The exposure concentration used in this ERA is the average concentration of 0.13 mg/kg.

Xylene was detected in 25 of 86 samples. Xylene concentrations ranged from 0.21 to 47 mg/kg. The background concentration of xylene is <0.6 mg/kg. The action level is 1.21 mg/kg. Xylene

is considered a COC as onsite concentrations are above action levels. The exposure concentration used in this ERA is the average concentration of 1.09 mg/kg.

3.1.2.3 Volatile Organic Compounds. Thirteen VOCs were detected in soil and/or sediment samples collected from Barter Island. The compounds detected were: n-butylbenzene; sec-butylbenzene; tert-butylbenzene; cis-1,2-dichloroethene; 1,2,4-trichlorobenzene; 1,2,4-trimethylbenzene; 1,3,5-trimethylbenzene; 1,4-dichlorobenzene; isopropylbenzene; p-isopropyltoluene; n-propylbenzene; tetrachloroethene; and methylene chloride. This section presents the evaluation of these compounds as COCs for the ERA.

n-Butylbenzene was detected in 4 of 30 soil and/or sediment samples analyzed for VOCs. The range of concentrations detected was 0.635 to 4.22 mg/kg. The background concentration of n-butylbenzene is <0.5 mg/kg. There are no action levels for this compound. n-Butylbenzene is an alkyl-substituted benzene, and the action level of 4.36 mg/kg for a similar compound, ethylbenzene, was used. Onsite concentrations do not exceed the action level, so it is not considered a COC.

sec-Butylbenzene was detected in 3 of 30 soil and/or sediment samples analyzed for VOCs. The range of concentrations detected was 1.24 to 1.62 mg/kg. The background concentration of sec-butylbenzene is <0.5 mg/kg. There are no action levels for this compound, so the action level of 4.36 mg/kg for a similar compound, ethylbenzene, was used. Onsite concentrations do not exceed the action level, so it is not considered a COC.

tert-Butylbenzene was detected in 1 of 30 soil and/or sediment samples analyzed for VOCs at a concentration of 0.256 mg/kg. The background concentration of tert-butylbenzene is <0.5 mg/kg. There are no action levels for this compound. Since tert-butylbenzene is an alkyl-substituted benzene, the action level of 4.36 mg/kg for ethylbenzene was used for this compound. The chemical was only detected once and onsite concentrations do not exceed the action level, so it is not considered a COC.

cis 1,2-Dichloroethene was detected in 1 of 30 soil and/or sediment samples collected at the Barter Island facility, at concentrations of 0.069 mg/kg. The background concentration of cis 1,2-dichloroethene in soil and sediment is <0.5 mg/kg. There are no action levels for cis 1,2-dichloroethene. The compound was only detected in one sample (<5% detection frequency), so it is not considered a COC.

1,2,4-Trichlorobenzene was detected in 1 of 30 soil and/or sediment samples, at a concentration of 0.046 mg/kg. The background concentration of this compound is <0.5 mg/kg. There is no action level for this compound. 1,2,4-Trichlorobenzene was only detected in one sample, so it is not considered a COC.

1,2,4-Trimethylbenzene was detected in 5 of 30 samples at concentrations of 0.041 to 14.7 mg/kg. The background concentration of this compound is <0.5 mg/kg. There is no action level for this compound. This compound is considered a COC with an average concentration of 0.98 mg/kg.

1,3,5-Trimethylbenzene was detected in 7 of 30 samples at concentrations of 0.048 and 9.32 mg/kg. The background concentration of this compound is <0.5 mg/kg. There is no action level for this compound. This compound is considered a COC with an average concentration of 0.93 mg/kg.

1,4-Dichlorobenzene was detected in 1 of 48 samples at a concentration of 0.044 mg/kg. The background concentration of this compound is <0.5 mg/kg. There is no action level for this compound. This chemical had a low frequency of detection (two percent) and was only detected in one sample, so it is not considered a COC.

Isopropylbenzene was detected in 3 of 30 soil and/or sediment samples at concentrations between 0.409 and 0.681 mg/kg. The background concentration is <0.5 mg/kg. There are no action levels for this compound. Isopropylbenzene is an alkyl-substituted benzene, and the action level of 4.36 mg/kg for a similar compound, ethylbenzene, was used. This compound was detected below the action level, so it is not considered a COC.

p-Isopropyltoluene was detected in 3 of 30 soil and/or sediment samples at concentrations between 1.69 and 2.47 mg/kg. The background concentration is <0.5 mg/kg. There are no action levels for this compound. The action level of 4.36 mg/kg for a similar compound, ethylbenzene, was used. Because this compound was detected below the action level, it is not considered a COC.

n-Propylbenzene was detected in 3 of 30 soil and/or sediment samples at concentrations between 0.918 and 1.17 mg/kg. The background concentration is <0.5 mg/kg. There are no action levels for this compound, so the action level of 4.36 mg/kg for a similar compound, ethylbenzene, was used. Because this compound was detected below the action level, it is not considered a COC.

Methylene chloride was detected in 1 of 48 soil and/or sediment samples at a concentration of 0.255 mg/kg. The background concentration is <0.3 mg/kg. The action level for this compound is 90 mg/kg. This compound was detected below the action level, so it is not considered a COC.

Tetrachloroethene was detected in 2 of 62 soil and/or sediment samples at concentrations of 0.023 and 5.42 mg/kg. The background concentration is <0.3 mg/kg. There is no action level for this compound. It had at a low frequency of detection (three percent), so it is not considered a COC.

3.1.2.4 Semivolatile Organic Compounds. Seven semivolatile compounds were detected in samples from the Barter Island installations: isophorone; fluoranthene; naphthalene; phenanthrene; pyrene; 2-methylnaphthalene; and bis(2-ethylhexyl)phthalate. This section presents the evaluation of these compounds as COCs for the ERA.

Isophorone was detected in 1 of 18 soil and/or sediment samples at a concentration of 1.49 mg/kg. The background concentration is < 0.23 mg/kg. There is no action level for this compound. It was only detected in one sample, so it is not considered a COC.

Fluoranthene was detected in 2 of 18 soil and/or sediment samples at concentrations of 1.7 and 2.28 mg/kg. The background concentrations is < 0.23 mg/kg; the action level for fluoranthene is 6.2 mg/kg. This compound was detected at concentrations below the action level, so it is not considered a COC.

Naphthalene was detected in 5 of 30 soil and/or sediment samples at concentrations of 0.105 to 46 mg/kg. The background concentration is <0.5 mg/kg. The action level for this compound is 0.34 mg/kg. It was retained as a COC, and the exposure concentration evaluated in this ERA is the average concentration of 2.26 mg/kg.

Phenanthrene was detected in 2 of 18 soil and/or sediment samples at concentrations of 1.96 and 4.79 mg/kg. The background concentrations ranged from <0.23 to <3.5 mg/kg. The action level for this compound is 1.8 mg/kg. Four additional soil/sediment samples in the area of the samples where phenanthrene was detected were all non-detect (<0.21 to <2.1 mg/kg). Therefore, the average concentration of phenanthrene in the immediate area is below action level, and installation wide the average concentration is significantly below action level. Patton and Dieter (1980) found no mortality or signs of toxicity when mallards were fed a diet that contained 4,000 mg/kg PAH (as naphthalenes and phenanthrene) over a period of seven months. The action level for this compound is 1.8 mg/kg (EPA Sediment Quality Criteria). The first sample location (4.79 mg/kg) is in close proximity to structures and is unlikely to be used on a regular basis for foraging or nesting. The other sample location, while less contaminated, may offer more suitable (and less regularly disturbed) habitat for local fauna, but the toxicity standard presented is so much higher than the sample concentration, that phenanthrene is not expected to present a risk to ecological receptors and is not considered a COC.

Pyrene was detected in 1 of 18 soil and/or sediment samples at a concentration of 1.26 mg/kg. The background concentration is < 0.23 mg/kg. The action level for pyrene is 13.1 mg/kg. This compound was detected at a concentration below the action level, so it is not considered a COC.

2-Methylnaphthalene was detected in 3 of 18 samples at concentrations ranging from 3.44 to 14.5 mg/kg. The background concentration of this compound is <0.23 mg/kg. The action level for this compound is 0.065 mg/kg. 2-Methylnaphthalene is considered a COC, and the exposure concentration used in this ERA is the average concentration of 1.55 mg/kg.

Bis(2-ethylhexyl)phthalate was detected in 1 of 18 samples at a concentration of 4.6 mg/kg. The background concentration of this compound is <0.23 mg/kg. The action level for this compound is 50 mg/kg. This chemical was detected below the action level, so it is not considered a COC.

3.1.2.5 Metals. Seventeen inorganic analytes were detected in soil and/or sediment samples collected from Barter Island. The metals detected were: aluminum; barium; beryllium; cadmium; calcium; chromium; iron; lead; magnesium; manganese; nickel; potassium; sodium; vanadium; and zinc. This section presents the evaluation of these metals as COCs for the ERA.

Aluminum was detected in 15 of 16 soil and/or sediment samples. Concentrations ranged from 1,600 to 7,500 mg/kg. Background concentrations ranged from 1,500 to 25,000 mg/kg. There

is no action level for aluminum. Onsite concentrations did not exceed background concentrations, so aluminum was not retained as a COC.

Barium was detected in 15 of 16 soil and/or sediment samples at concentrations between 13 and 120 mg/kg. The background concentrations of barium ranged from 27 to 390 mg/kg. There is no action level for barium. Onsite concentrations did not exceed background concentrations, so barium was not retained as a COC.

Beryllium was detected in 1 of 16 soil and/or sediment samples at a concentration of 3.2 mg/kg. The maximum background concentration of barium is 6.4 mg/kg. There is no action level for beryllium. Beryllium was not retained as a COC because onsite concentrations did not exceed background concentrations.

Cadmium was detected in 2 of 16 soil and/or sediment samples at a concentrations of 2.7 and 2.8 mg/kg. Cadmium was not detected in background samples (<3.0 - <36 mg/kg). The action level for cadmium is 5 mg/kg. Cadmium was not retained as a COC because the detected concentrations were below the action level.

Calcium was detected in all 16 soil and/or sediment samples. Concentrations ranged from 2,200 to 33,000 mg/kg. Background concentrations ranged from 360 to 59,000 mg/kg. There is no action level for calcium. Onsite concentrations did not exceed background concentrations, so this chemical was not retained as a COC.

Chromium was detected in all 16 soil and/or sediment samples. Concentrations ranged from 3 to 53 mg/kg. The maximum background concentration is 47 mg/kg. The action level for chromium is 87 mg/kg. The detected concentrations did not exceed the action level, so this metal is not retained as a COC.

Iron was detected in all 16 soil and/or sediment samples. Concentrations ranged from 4,700 to 17,000 mg/kg. The background concentrations ranged from 5,400 to 35,000 mg/kg. There is no action level for iron. This metal was not retained as a COC because onsite concentrations did not exceed background concentrations.

Lead was detected in 8 of 16 soil and/or sediment samples. Concentrations ranged from 6.1 to 231 mg/kg. The maximum background concentration for lead is 22 mg/kg. There is no action level for lead. Lead is retained as a COC; the exposure concentration evaluated in this ERA is the average concentration of 26.6 mg/kg.

Magnesium was detected in all 16 soil and/or sediment samples. Concentrations ranged from 990 to 17,000 mg/kg. The background concentrations for magnesium ranged from 360 to 7,400 mg/kg. There is no action level for magnesium. Because magnesium is an essential nutrient and is ubiquitous in the environment it was not retained as a COC.

Manganese was detected in all 16 soil and/or sediment samples. Concentrations ranged from 40 to 380 mg/kg. The background concentrations for manganese ranged from 25 to 290 mg/kg.

There are no action levels for manganese. This trace mineral is not retained as a COC because it was only detected above background in a single sample.

Nickel was detected in all 16 soil and/or sediment samples ranging in concentration from 4.3 to 16 mg/kg. The background concentrations ranged from 4.2 to 46 mg/kg. The action level for nickel is 30 mg/kg. This metal was not retained as a COC because onsite concentrations did not exceed the action level.

Potassium was detected in 9 of 16 soil and/or sediment samples. Concentrations ranged from 300 to 610 mg/kg. The background concentrations ranged from <300 to 2,200 mg/kg. There is no action level for potassium. This metal was not retained as a COC because onsite concentrations were below background concentrations.

Sodium was detected in all 16 soil and/or sediment samples. Concentrations detected ranged from 38 to 870 mg/kg, which slightly exceeded the maximum background concentration of 680 mg/kg. There is no action level for sodium. This metal was not retained as a COC because sodium was detected only once above background and the sample did not significantly exceed the background concentration. In addition, sodium is ubiquitous in the environment and is not expected to pose a threat to ecological receptors.

Vanadium was detected in 15 of 16 soil and/or sediment samples ranging in concentration from 4.9 to 21 mg/kg. The background concentrations ranged from 6.3 to 59 mg/kg. There is no action level for vanadium. This metal was not retained as a COC because onsite concentrations were below background concentrations.

Zinc was detected in all 16 soil and/or sediment samples at concentrations from 12 to 500 mg/kg. The background concentrations for zinc ranged from 9.2 to 95 mg/kg. The action level for zinc is 120 mg/kg. Zinc was detected in excess of the action level, so zinc was retained as a COC. The exposure concentration evaluated in this ERA is the average concentration of 76 mg/kg.

3.1.2.6 Polychlorinated Biphenyls. Twenty-seven soil and/or sediment samples were collected and analyzed for PCBs. PCBs were detected in 5 samples at levels ranging from 0.112 to 2.72 mg/kg. The background concentrations of PCBs in soils ranged from <0.02 to <0.1 mg/kg. The action level for PCBs is 10 mg/kg. Onsite concentrations of PCBs did not exceed action levels, however because PCBs can bioaccumulate in the environment they are considered a COC. The exposure concentration is the average concentration of 0.42 mg/kg. (Note: These PCB concentrations do not include PCB detections at sites that were eliminated from further evaluation because of the lack of suitable habitat.)

3.2 ECOLOGICAL EXPOSURE ASSESSMENT

The vegetation of the Arctic Coastal Plain and the ecosystems it characterizes have developed primarily as a result of the low relief and harsh environment. The growing season is short, typically extending from June through mid-September. Winters are long, cold, dry, and dark. Air

temperatures that average below freezing for most of the year result in a permafrost layer that begins near the surface and reaches to depths as great as 610 meters. Seasonal thawing results in an active layer between ground surface and 3.7 meters below the surface (Hart Crowser 1987).

The impervious permafrost layer prevents percolation and infiltration of water below the active layer, and the generally flat terrain provides poor drainage. As a result, the ecosystems of the Arctic Coastal Plain are often defined not only by their plant associations but also by the degree of water found in and on them. Hart Crowser (1987) describes five major ecosystems for the classification of tundra and Arctic Coastal Plain communities:

- **Marine zones:** these include lagoons, estuaries, barrier islands, strands and beaches. The abundance of vegetation along the marine coastal zone is inversely related to the amount of beach scouring by waves and ice. Mainland beaches support a variety of vegetation, including sedges, grasses, and forbs.
- **Wet sedge meadows:** an association of meadows, ponds and lakes also known as "wet tundra". This system, with its associated wetlands, is dominant in the area extending west from the Colville River to the Chukchi Sea (including the Point Lonely, Point Barrow, Wainwright, Point Lay, and Cape Lisburne installations). Differences in vegetation within this ecosystem are related to moisture and microrelief.
- **Tussock tundra:** or "moist tundra" consisting primarily of areas dominated by tussock-forming cottongrass. This system covers significant portions of the Arctic Coastal Plain.
- **Riverine systems and floodplains:** including riparian shrubland on recent and old alluvium. Being better drained than surrounding lands, the riparian environment supports a distinctive "shrub thicket" vegetation.
- **Alpine tundra:** including rocky upland areas of sparse, mat-forming or fell-field vegetation.

The species associated with each ecosystem at the Barter Island DEW Line installation are potential receptors.

3.2.1 Potential Receptors

Both the potential receptors and the representative species selected will be characteristic of the Barter Island installation as well as the seven installations to be assessed in the future.

The Barter Island installation is located along the northern boundary of the Arctic Coastal Plain. Hart Crowser (1987) and Woodward-Clyde (1993) have listed the species likely to occur along the coastal plain based on site-specific studies and a review of the literature. The marine zone, wet sedge meadows, tussock tundra, and riverine/riparian are the primary ecosystems found at the Barter Island installation. Alpine tundra is minimal at the site and is not evaluated further.

3.2.1.1 Plants. Plants commonly associated with the marine zone are sedges, grasses and forbs. *Carex subspathacea* and *C. aquatilis* are dominant plants in the coastal wetlands.

The wet sedge meadow (also known as "wet tundra") is characterized by a variety of sedges and grasses. Typical species include: cottongrass, *Eriophorum* spp.; tundra grass, *Dupontia fischeri*; and mosses, *Sphagnum* spp. Marsh marigold, *Caltha palustris*; and horsetail, *Equisetum* spp. may be found in wetter areas (Hart Crowser 1987).

The tussock tundra (or moist tundra) is drier than the wet sedge meadow/wet tundra association. Tussock-forming cottongrass is the dominant plant species. Grasses, sedges, dwarf shrubs, mosses, and lichens are scattered throughout the tussock complex. These species include: willows, *Salix* spp.; Labrador tea, *Ledum palustre*; blueberry and lingonberry, *Vaccinium* spp.; and lousewort, *Pedicularis* spp (NPRA Task Force 1978; Bergman et al. 1977).

Riverine/riparian systems are composed of a diversity of habitat types and species. The dominant plants here are shrubs with a scattered understory of grasses, herbs, and lower growing shrubs. Larkspur, *Delphinium brachycentrum*; cinquefoil, *Potentilla* spp.; bearberry, *Arctostaphylos* spp.; and wormwood, *Artemisia arctica* are common species (NPRA Task Force 1978; Bergman et al. 1977).

3.2.1.2 Aquatic Organisms. Sixty-six species of fish inhabiting marine, estuarine, and freshwater systems have been identified in the Arctic region (Hart Crowser 1987). Marine species habitating the nearshore and offshore waters include: boreal smelt, *Osmerus eperlanus*; Pacific herring, *Clupea harengus*; Arctic cod, *Boreogadus saida*; and fourhorn sculpin, *Myoxocephalus quadricornis*. Anadromous species using Arctic rivers for spawning include the Arctic cisco, *Coregonus autumnalis*; Arctic char, *Salvelinus alpinus*; and occasional pink and chum salmon, *Onchorhynchus* spp. Lack of overwintering habitat is a significant limiting condition for both anadromous and freshwater fish of the Arctic region. The principal freshwater fish found in the region are grayling, *Thymallus arcticus*; lake trout, *Salvelinus namaycush*; burbot, *Lota lota*; and nine-spine stickleback, *Pungitius pungitius* (Hart Crowser 1987).

Invertebrates that may be present in the waters and wet habitats of the Arctic Coastal Plain are well represented by the crustaceans (i.e., copepods, isopods, amphipods, and decapods).

3.2.1.3 Birds. There are approximately 180 species of birds seasonally associated with the habitats of the Arctic Coastal Plain. Of these, 36 are considered to be regular migratory breeders along the Beaufort Sea coast and 17 are intermittent breeders (Koranda and Evans 1975).

Waterfowl, seabirds, and shorebirds are numerous and the dominant components of the coastal plain avifauna. Bird use of the coastal plain is highly seasonal and associated with typical avian breeding and migration cycles. Shoreline habitats are used significantly in association with molting, pre-migratory staging, and post breeding movement. These habitats are considered critical by the U.S. Fish and Wildlife Service (USFWS). Principal species include: glaucous gull, *Larus hyperboreus*; red phalarope, *Phalaropus fulicaria*; dunlin, *Calidris alpina*; loons, *Gavia* spp.; sandpipers, *Calidris* spp.; eiders, *Somateria* spp.; and geese, *Branta* spp. and *Chen* sp. Among

the migratory passerine species using the coastal habitats are the Savannah sparrow, *Passerculus sandwichensis*; common and hoary redpolls, *Carduelis* spp.; snow bunting, *Plectrophenax nivalis*; and Lapland longspur, *Calcarius lapponicus* (Woodward-Clyde 1993).

3.2.1.4 Mammals. The mammalian fauna of the Arctic Coastal Plain and adjacent waters is relatively simple compared to faunas at lower latitudes. A review of species lists indicates a total of 38 species that commonly occur in the Arctic; 11 of these are marine mammals (Hart Crowser 1987). A sampling of the terrestrial mammals geographically associated with the DEW Line stations, including Barter Island, consists of: brown lemming, *Lemmus trimucronatus*; masked shrew, *Sorex cinereus*; Arctic fox, *Alopex lagopus*; red fox, *Vulpes vulpes*; weasels, *Mustela* spp.; tundra vole, *Microtus oeconomus*; caribou, *Rangifer tarandus*; and grizzly bear, *Ursus arctos* (Hart Crowser 1987; Woodward-Clyde 1993).

Marine mammals of the Arctic coast include six species of whales; polar bear, *Ursus maritimus*; five species of seals; and walrus, *Odobenus rosmarus*. The most common of these are: beluga, *Delphinapterus leucas*; bowhead whale, *Balaena mysticetus*; polar bear; gray whale, *Eschrichtius robustus*; ringed seal, *Phoca hispida*; bearded seal, *Erignathus barbatus*; and walrus (Hensel et al. 1984).

3.2.1.5 Endangered and Threatened Species. Species of the Arctic Coastal Plain and nearby waters that are protected by federal and state designations include: bowhead whale; fin whale, *Balaenoptera physalus*; sei whale, *Balaenoptera borealis*; hump-backed whale, *Megaptera novaeangliae*; and gray whale. Avian species include the Arctic peregrine falcon, *Falco peregrinus tundrius*; spectacled eider, *Somateria fischeri*; and Steller's eider, *Polysticta stelleri*. Two plant species that potentially occur on, or near, the sites are a mustard, *Thlaspi arcticum*, and round-leaf willow, *Salix ovalifolia* var. *glacialis* (Hart Crowser 1987).

3.2.2 Representative Species

It is impractical to evaluate all of these receptors individually because of the great diversity of plants and animals at a given site. Thus, for ERAs, a set of "representative species" is selected for further evaluation. The representative species are selected based primarily on the species' likelihood of exposure based on their preferred habitat and feeding habits. The abundance of a species, relative to the areal extent of the sites, is also considered. The representative species encompass a range of ecological niches to best characterize the ecosystems being examined. In addition, species are selected, in part, as a result of the availability of toxicity, exposure, and life history information. Species that may be sensitive to environmental impacts, such as endangered or threatened species, are also evaluated.

For the DEW Line stations, a number of groups of receptors are evaluated including plants, aquatic invertebrates, fish, birds and mammals. Potential risks to representative species are estimated by evaluating sampling data for the relevant exposure media (i.e., soil, sediments and surface water). For the birds and mammals selected, exposures are estimated by evaluating their potential dietary intakes of COCs.

The similarity of ecosystems at each of the station sites allows the use of the same set of representative species for all installations. It may be possible that a representative species may inhabit the general area of an installation, but not occur specifically on the installation property. When and if this situation occurs, it will be noted. Table 3-3 presents the representative species for the DEW Line installation sites (including Barter Island) and Table 3-4 lists the threatened and endangered species considered in the ERA. The USFWS was consulted in the selection of the threatened and endangered species evaluated.

3.2.2.1 Representative Plants. Plants selected as representative species are: sedges, *Carex* spp.; willows, *Salix* spp.; and cottongrass, *Eriophorum* spp. These species are selected because they are abundant on all the sites, are important links in the trophic structure of the ecosystems of the Arctic, and represent a major percentage of the primary production along the coastal plain. The blueberry, huckleberry, and lingonberry, *Vaccinium* spp., are evaluated because of their roles as forage plants and as subsistence species.

3.2.2.2 Representative Aquatic Invertebrates and Fish. The invertebrates selected as representative species are *Daphnia* spp. The fish species chosen are the Arctic char and the nine-spined stickleback. *Daphnia* spp. are abundant and represent a portion of the diet of the selected fish species (Johnson and Burns 1984; Wootton 1976), and toxicity information is readily available for them. The Arctic char is a common anadromous species (exposed to both fresh and saltwater) and is a valuable recreation and subsistence resource (Johnson and Burns 1984).

The nine-spined stickleback is a freshwater species that also uses brackish habitats, nests in aquatic vegetation, and is prey for other fish and bird species (Wootton 1976).

3.2.2.3 Representative Birds. The avian species selected as representative are: Lapland longspur; brant, *Branta bernicla*; glaucous gull; and pectoral sandpiper, *Calidris melanotos*. The Lapland longspur is a passerine belonging to a terrestrial feeding guild (including sandpipers, turnstones, and phalaropes) (Custer and Pitelka 1978). The longspur's diet of insects and seeds (Custer and Pitelka 1978) makes it an important link in the Arctic trophic web. The brant nests and molts among the numerous ponds in the tussock tundra and grazes on sedges and cottongrass (Palmer 1976). It is considered to be an important subsistence resource. The glaucous gull is a predatory scavenger that feeds on small mammals, young birds, carrion and garbage, and breeds along the Arctic Coastal Plain (Farrand 1983). The pectoral sandpiper is an abundant shorebird that is primarily insectivorous and breeds on the Arctic Coastal Plain.

3.2.2.4 Representative Mammals. The representative species of mammals selected are the brown lemming, Arctic fox, and the barren-ground caribou. The brown lemming is the predominant small mammal at all sites. The lemming consumes more vegetation than expected for an animal its size, due to its low assimilation efficiency, the low nutrient value of winter forage, and the high metabolic demands of the Arctic environment (Chappell 1980). The Arctic fox is a representative species because it is ubiquitous along the coastal plain and its carnivorous diet (mostly lemmings) places it near the top of the trophic structure in the Arctic. Eberhardt et al. (1982) note that in fall and winter, and to a lesser extent in summer, the Arctic fox frequently uses areas near development. This tendency may expose the fox to potential pathways of

TABLE 3-3. REPRESENTATIVE SPECIES AT THE DEW LINE INSTALLATION SITES

COMMON NAME	GENUS AND SPECIES
Sedge	<i>Carex</i> spp.
Cottongrass	<i>Eriophorum</i> spp.
Willow	<i>Salix</i> spp.
Berries	<i>Vaccinium</i> spp.
Water fleas	<i>Daphnia</i> spp.
Nine-spined stickleback	<i>Pungitius pungitius</i>
Arctic char	<i>Salvelinus alpinus</i>
Lapland longspur	<i>Calcarius lapponicus</i>
Brant	<i>Branta bernicla</i>
Glaucous gull	<i>Larus hyperboreus</i>
Pectoral sandpiper	<i>Calidris melanotos</i>
Brown lemming	<i>Lemmus trimucronatus</i>
Arctic fox	<i>Alopex lagopus</i>
Barren-ground caribou	<i>Rangifer tarandus</i>

contamination. Additionally the fox, a relatively common furbearer, can be an important subsistence resource. The caribou is selected as a representative species because it uses areas on, or near, a number of the sites during migration and for calving and summering grounds. In addition, the caribou is a significant subsistence resource (USFW 1982; Cuccarese et al. 1984; Hensel et al. 1984).

3.2.2.5 Threatened and Endangered Species. The threatened and endangered species that potentially occur at the DEW Line installations are the arctic peregrine falcon, spectacled eider, and Steller's eider. When the ecological risk assessment process for the DEW Line installations began, the arctic subspecies of peregrine falcon (*Falco peregrinus tundrius*) was listed as threatened and, as such, it is included in the Barter Island ERA. In the interim, the peregrine's arctic subspecies has been delisted and will not be considered in future ERAs. The spectacled eider is federally listed as threatened, and Steller's eider is a candidate for listing as threatened. The U.S. Fish and Wildlife Service indicated that Steller's eider was likely to be listed as threatened sometime in 1995 (Ambrose 1994 pers. comm.), but as of April 1995 a federal moratorium on additions to the threatened and endangered lists was in effect.

TABLE 3-4. THREATENED AND ENDANGERED SPECIES TO BE CONSIDERED IN THE ECOLOGICAL RISK ASSESSMENT

COMMON NAME	GENUS AND SPECIES
Peregrine falcon ^a	<i>Falco peregrinus</i>
Spectacled eider ^a	<i>Somateria fischeri</i>
Steller's eider ^b	<i>Polysticta stelleri</i>

^a threatened status

^b candidate for threatened status

3.2.3 Exposure Pathways

This section discusses potential exposure pathways for ecological receptors. In addition, methods used to quantify exposures to selected species of plants, aquatic organisms, birds, and mammals are presented. Quantitative estimates of exposure will be used with toxicity reference values (TRVs) derived in Section 3.3 to estimate risks in the risk characterization section (Section 3.4).

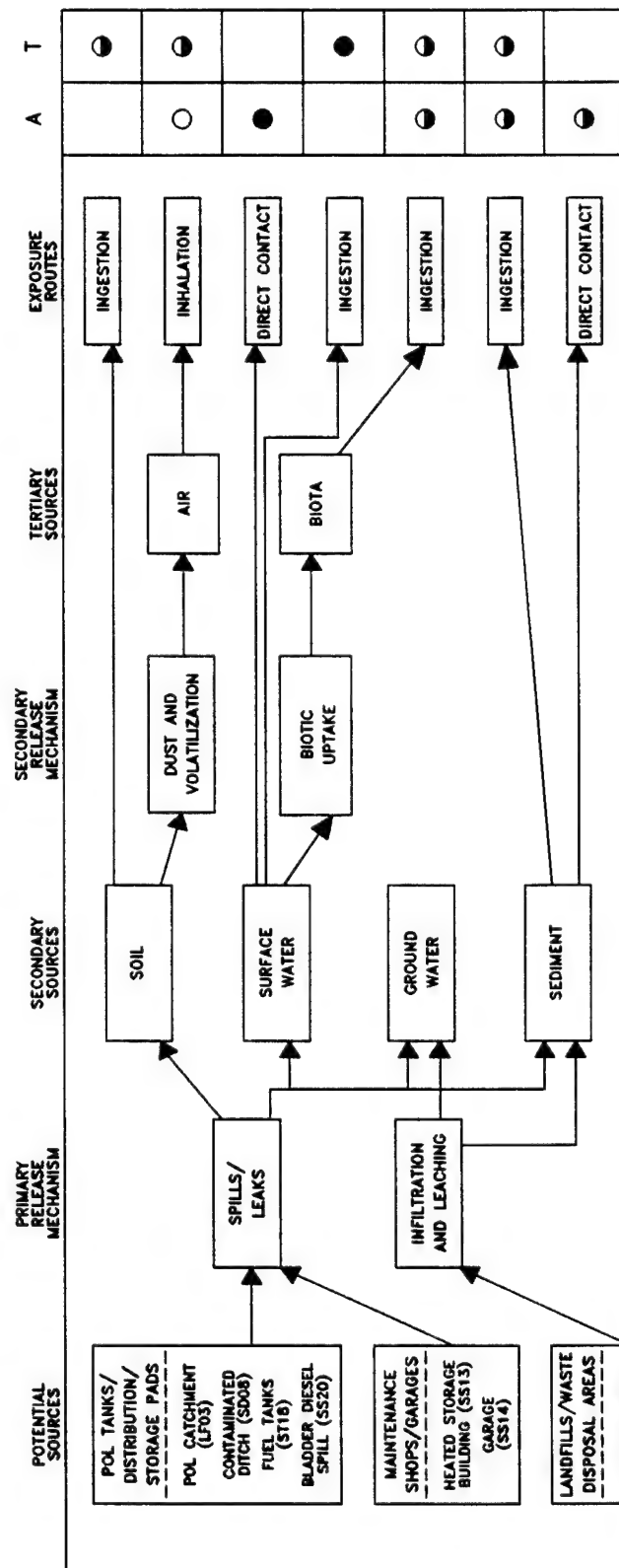
Ecological receptors can be exposed to COCs through abiotic and biotic media. Potential exposure pathways for terrestrial and aquatic organisms are summarized in Figure 3-1. The following sections describe the potential exposure routes and a determination of pathways evaluated in the risk assessment.

Potential risks to plants from exposure to COCs in soil and water will be addressed. The most significant route of exposure for plants is direct contact with soil at the site.

Aquatic organisms such as fish and invertebrates are primarily exposed through direct contact with surface water. Surface water is in direct contact with dermal surfaces as well as gills and other respiratory structures. Fish and invertebrates also may be exposed to COCs through ingestion of plant and animal items in the diet, and incidental ingestion of sediments while foraging. Direct contact is the primary exposure route, however, and these secondary routes will not be evaluated for aquatic organisms.

Wildlife, such as birds and mammals, may be exposed to COCs through a variety of pathways including ingestion of surface water used for drinking, ingestion of plant and animal diet items, and incidental ingestion of surface soils and sediments while foraging. Wildlife is not expected to be exposed to COCs via inhalation because the surface soils are well vegetated and moist during the growing season and frozen and/or snow covered the remainder of the year. Therefore this pathway is not evaluated in the ERA.

DRAWING No. BTR-FL02



BARTER ISLAND RADAR INSTALLATION

USAF 611th CES

FIGURE NO. 3-1

ECOLOGICAL RISK ASSESSMENT POTENTIAL EXPOSURE PATHWAYS

- A AQUATIC RECEPTORS
- T TERRESTRIAL RECEPTORS
- COMPLETE EXPOSURE PATHWAY
- POTENTIALLY COMPLETE PATHWAY
- INCOMPLETE EXPOSURE PATHWAY

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Insufficient information is available for the representative species to quantify exposures from dermal contact with soil or sediments, therefore these pathways are not quantitatively evaluated. Because soils and sediments represent potential pathways, total exposures for the representative species could be underestimated. This represents one of the uncertainties in this risk assessment that are discussed in Section 3.5.

3.2.4 Habitat Suitability for Representative Species

In order to assess the representative species' degree of exposure to the COCs, the habitat suitability of each of the 14 separate sites was evaluated. The habitat suitability evaluation considered each representative aquatic, bird, and mammal species.

Not all habitats in the sites are suitable for the representative species. This habitat suitability evaluation avoids unnecessarily quantifying a species' exposure to COCs in unsuitable habitats.

Habitat suitability for the representative species is based on availability of food, water, cover, resting places, and breeding or nesting sites. Life history information and the habitat preferences and needs of the representative species were used to determine the habitat suitability of the sites. The preferences and needs give a general indication of areas where a species is likely to occur. In situations where there was uncertainty about the likelihood of a species' occurrence, a conservative approach was used, and it was assumed that the species use the area.

The sites at the Barter Island installation were evaluated using site plans developed for Barter Island and photographs taken during the remedial investigation. No site-specific studies were done at the Barter Island installation to assist in the habitat evaluation. In some cases, chemical sampling locations are limited to a small portion of the area of concern, while other areas are sampled along potential chemical migration pathways. The extent of the area sampled was used as a general guide for the bounds of a particular area. Professional judgement was used to make assumptions and decisions regarding the suitability of a particular area for each species. In particular, note that the habitat at the JP-4 Spill is classified as unsuitable under current conditions. One sample location, S03 (see Figure 2-17), is located in an area that may be viewed as potentially suitable habitat (disturbed tundra). The remaining samples at the JP-4 Spill site are characterized by a gravel pad and classified as unsuitable habitat. Overall the site is deemed to be unsuitable habitat, however, it is evaluated as potentially suitable habitat in a future scenario in Section 3.4.5, Potential Future Risks.

A summary of the habitat suitability evaluation is presented in Table 3-5. Figures E-1 through E-16 in Appendix E illustrate habitat suitability at each of the sites. The criteria and rationale used to characterize habitat suitability for each of the representative species follow.

3.2.4.1 Aquatic Organisms.

Arctic char. The Arctic char is an anadromous species. Exposure of the char to COCs is only possible in streams that flow from the Barter Island installation to the Beaufort Sea. Sites associated with such streams are the Old Landfill, Current Landfill, Contaminated Ditch, and Old Dump Site.

TABLE 3-5. HABITAT POTENTIAL FOR REPRESENTATIVE SPECIES AT THE BARTER ISLAND DEW LINE INSTALLATION

SITE	LF01	LF03	LF04	SD08	LF12	SS13	SS14	SS15	SS16	ST17	ST18	LF19	SS20	SS21
SPECIES														
Arctic char	YES	NO	YES	YES	NO	NO	NO	NO	NO	NO	NO	YES	NO	NO
Nine-spined stickleback	YES	NO	YES	YES	NO	YES	YES	NO	NO	NO	YES	YES	YES	NO
<i>Daphnia</i> spp.	YES	YES	YES	YES	NO	YES	YES	NO	NO	NO	YES	YES	YES	NO
Lapland longspur	YES	YES	YES	YES	NO	YES	YES	NO	NO	NO	YES	YES	YES	
Brant	YES	YES	NO	NO	NO	YES	YES	NO	NO	NO	NO	YES	YES	NO
Glaucous gull	YES	YES	YES	YES	YES	YES	YES	NO	NO	NO	YES	YES	YES	NO
Pectoral sandpiper	YES	YES	YES	YES	YES	YES	YES	NO	NO	NO	YES	YES	YES	NO
Peregrine falcon	YES	YES	YES	YES	YES	YES	YES	NO	NO	NO	YES	YES	YES	NO
Spectacled eider	YES	YES	NO	NO	YES	YES	YES	NO	NO	NO	NO	YES	YES	NO
Steller's eider	YES	YES	NO	NO	YES	YES	YES	NO	NO	NO	NO	YES	YES	NO
Brown lemming	YES	NO	YES	YES	NO	YES	NO	NO	NO	NO	YES	YES	YES	NO
Arctic fox	YES	NO	YES	YES	YES	YES	NO	NO	NO	NO	YES	YES	YES	NO
Caribou	NO	NO	NO	NO	NO	NO	NO	NO	NO	NO	NO	NO	NO	NO

LF01 = Old Landfill
 LF03 = POL Catchment
 LF04 = Current Landfill
 SD08 = Contaminated Ditch
 LF12 = Old Runway Dump
 SS13 = Heated Storage
 SS14 = Garage

SS15 = Weather Station Building
 SS16 = White Alice Facility
 ST17 = POL Tanks
 ST18 = Fuel Tanks
 LF19 = Old Dump Site
 SS20 = Bladder Diesel Spill
 SS21 = JP-4 Spill

Areas that do not have water access to the sea, and do not provide habitat for the Arctic char are the POL Catchment, Old Runway Dump, Heated Storage, Garage, Weather Station Building, White Alice Facility, POL Tanks, Fuel Tanks, Bladder Diesel Spill, and JP-4 Spill.

Nine-spined stickleback. The nine-spined stickleback is a freshwater fish species, although it will tolerate some degree of brackish water. It needs ponds, lakes, or streams with vegetation for breeding. Such environment is present at the Old Landfill, Current Landfill, Contaminated Ditch, Heated Storage, Garage, Fuel Tanks, Old Dump Site, and Bladder Diesel Spill.

Areas that do not have sufficient water to provide habitat for nine-spined sticklebacks are the POL Catchment, Old Runway Dump, Weather Station Building, White Alice Facility, POL Tanks, and JP-4 Spill.

***Daphnia* spp.** Aquatic invertebrates, like the crustacean, *Daphnia* spp., may be found throughout the Barter Island installation wherever there is sufficient standing water. These sites include the Old Landfill, POL Catchment, Current Landfill, Contaminated Ditch, Heated Storage, Garage, Fuel Tanks, Old Dump Site, and Bladder Diesel Spill.

Areas that probably do not support viable *Daphnia* populations are the Old Runway Dump, Weather Station Building, White Alice Facility, POL Tanks, and JP-4 Spill.

3.2.4.2 Birds.

Lapland longspur. The Lapland longspur is a migratory passerine that shares a feeding guild with shorebirds (sandpipers and phalaropes) (Custer and Pitelka 1978). Its Alaskan diet consists mainly of seeds in May and August, and primarily insects when they are abundant in June and July. Competition between the longspur and the shorebirds is mitigated by the longspurs' preference for a more upland, or drier habitat than the shorebirds.

Areas that potentially meet the habitat needs of the Lapland longspur include the Old Landfill, POL Catchment, Current Landfill, Contaminated Ditch, Heated Storage, Garage, Fuel Tanks, Old Dump Site, and Bladder Diesel Spill.

The longspur is not likely to use the habitats of the Old Runway Dump, Weather Station Building, White Alice Facility, POL Tanks, or JP-4 Spill. These areas are mostly gravel and lack vegetation to provide cover, forage or nest sites.

Brant. The brant is a migratory waterfowl species that nests in cottongrass/sedge communities on islands or peninsulas in large shallow ponds along the coastal plain (Cross 1993). It is mainly herbivorous, with small amounts of crustaceans, insects, and mollusks in its diet (Palmer 1976). Breeding pairs of brant are territorial and can become dense in suitable habitat (Hansen and Nelson 1957 in Palmer 1976).

The areas that contain potential brant habitat include the Old Landfill, POL Catchment, Heated Storage, Garage, Old Dump Site, and Bladder Diesel Spill.

The brant is not likely to be found at the Current Landfill, Contaminated Ditch, Old Runway Dump, Weather Station Building, White Alice Facility, POL Tanks, Fuel Tanks, or JP-4 Spill sites. The primary considerations for excluding these areas is the lack of ponds or lakes and the scarcity of vegetation for foraging, cover, or nest sites.

Glaucous gull. The glaucous gull is a migratory species that is both predatory and a scavenger. Its diet is varied, consisting of fish, birds, insects, crustaceans, mollusks, and garbage. The glaucous gull is opportunistic and will feed when and wherever prey, carrion or refuse is available.

The areas that have suitable habitat for the glaucous gull include the Old Landfill, POL Catchment, Current Landfill, Contaminated Ditch, Old Runway Dump, Heated Storage, Garage, Fuel Tanks, Old Dump Site, and Bladder Diesel Spill.

Areas that are not expected to provide suitable habitat for glaucous gulls include the Weather Station Building, White Alice Facility, POL Tanks, and JP-4 Spill. These sites are mostly gravel and are not likely to have any of the prey species or other habitat needs for the gull.

Pectoral sandpiper. The pectoral sandpiper is a migratory shorebird that feeds along the wet margins of lakes, ponds, and streams (Scott 1983). Suitable habitat on the coastal plain tundra is abundant, as are the insects it feeds on. Other diet items include mollusks, crustaceans, worms and plant debris (Pitelka 1959).

Areas with suitable pectoral sandpiper habitat are the Old Landfill, POL Catchment, Current Landfill, Contaminated Ditch, Old Runway Dump, Heated Storage, Garage, Fuel Tanks, Old Dump Site, and Bladder Diesel Spill.

The pectoral sandpiper is not likely to find suitable habitat at the Weather Station Building, White Alice Facility, POL Tanks, or JP-4 Spill. As with the other species, these habitats are unsuitable for the sandpiper because of the lack of foraging areas (mostly gravel) and inappropriate areas for breeding (lack of cover).

Threatened and endangered species. The Barter Island installation was evaluated for habitat suitability for threatened and endangered species using site maps, photographs, and information from the literature. Professional judgement was used to integrate the evidence and make an informed decision regarding suitability of the habitat for threatened and endangered species. In addition, the Air Force had surveys conducted for spectacled and Steller's eiders at most DEW Line stations on the North Slope and concluded there was a low probability that either species was currently nesting or raising broods in the vicinity of the Barter Island installation; although historical records do indicate that spectacled and Steller's eiders have been present in the vicinity (Alaska Biological Research 1994). In the interest of a complete evaluation of all sensitive species, spectacled and Steller's eiders are included in this ERA.

Peregrine falcon. The peregrine falcon is a migratory raptor that is listed as threatened under the Federal Endangered Species Act (as this assessment was being prepared, the arctic subspecies of the peregrine was delisted). The falcon preys predominately on small to medium-

sized birds. Peregrine falcon habitat can be described as riparian cliffs, bluffs, and tall inaccessible structures where it nests and wherever its prey may be found (Farrand 1983). It has been reported that the falcon nests on the Beaufort Sea bluffs (1991 and 1992) along the north shore of the installation (Woodward-Clyde 1993). A survey conducted in 1994 (Alaska Biological Research) reports peregrine falcons at the Barter Island installation. No other site-specific studies are available to evaluate the Barter Island installation for suitability for threatened and endangered species.

The potential peregrine falcon habitats at the Barter Island installation are the Old Landfill, POL Catchment, Current Landfill, Contaminated Ditch, Old Runway Dump, Heated Storage, Garage, Fuel Tanks, Old Dump Site, and Bladder Diesel Spill.

Areas that are unlikely to provide habitat for the peregrines' prey are the Weather Station Building, White Alice Facility, POL Tanks, and JP-4 Spill. In the absence of prey, it is unlikely that the peregrine will use the terrestrial habitat in these areas. It is possible that the towers in these areas are used for resting or nesting, but these uses will not result in potential pathways of exposure to contaminants.

Spectacled eider. The spectacled eider is a migratory waterfowl species that is listed as threatened under the Federal Endangered Species Act. The spectacled eider's habitat is primarily marine outside of the breeding season. This species is being evaluated for the Barter Island installation because it may nest on islets in tundra ponds and lakes, as well as on the coastal plain around the DEW Line site (Palmer 1976).

The sites that contain potential habitat for the spectacled eider are the Old Landfill, POL Catchment, Old Runway Dump, Heated Storage, Garage, Old Dump Site, and Bladder Diesel Spill.

Areas that are unlikely to provide habitat for the spectacled eider are the Current Landfill, Contaminated Ditch, Weather Station Building, White Alice Facility, POL Tanks, Fuel Tanks, and JP-4 Spill. These areas do not provide adequate expanses of water or sufficient vegetation for foraging or nesting cover to accommodate the spectacled eider.

Steller's eider. Steller's eider is also a migratory waterfowl species. It is currently listed as a candidate for threatened status by the USFWS, and is likely to be listed as threatened within one year (Ambrose 1994 pers. comm.). The majority of the world population of Steller's eider breeds along the Arctic coast of Siberia and the only other location where they are described as "regular" breeders is near Barrow, Alaska (USFWS 1992). Steller's eider is being evaluated as a protected species for Barter Island because it may be a casual visitor or breeder. Its habitat preferences and habits are similar to those of the spectacled eider.

The sites that contain habitat suitable for Steller's eider are the Old Landfill, POL Catchment, Old Runway Dump, Heated Storage, Garage, Old Dump Site, and Bladder Diesel Spill.

Steller's eider is not likely to find suitable habitat at the Current Landfill, Contaminated Ditch, Weather Station Building, White Alice Facility, POL Tanks, Fuel Tanks, or JP-4 Spill. These areas

do not provide adequate expanses of water or sufficient vegetation to provide suitable habitat for this species.

3.2.4.3 Mammals.

Brown lemming. The brown lemmings' home range may average 0.5 hectare (ha), 1.23 acres, (Nowak 1991). Some of the sites are large enough to encompass the lemmings' entire home range. The lemming needs thick vegetation to support its foraging demands (Chappell 1980) and to use as cover from predators. Additionally, the soil must be well-drained and have enough cohesion to keep tunnels and burrows from collapsing .

The sites that have suitable habitat for brown lemmings are the Old Landfill (LF01), Current Landfill (LF04), Contaminated Ditch (SD08), Heated Garage (SS13), Fuel Tanks (ST18), Old Dump Site (LF19), and Bladder Diesel Spill (SS20). The Heated Storage, Fuel Tanks, and Old Dump Site areas are questionable habitats, but sample locations linked to these areas are in the drainage pathways from these sites. The drainage pathways and the adjacent land provides good lemming habitat.

Areas lacking suitable lemming habitat include the POL Catchment (LF03), Old Runway Dump (LF12), Garage (SS14), Weather Station Building (SS15), White Alice Facility (SS16), POL Tanks (ST17), and Bladder Diesel Spill (SS21). These areas lack dense vegetation and are principally gravel mounds that do not meet the food, cover and burrowing requirements of the lemming.

Arctic fox. The Arctic foxes' home range can be very extensive. Depending on the abundance of prey (primarily lemmings), the fox may range over 20 km² (Eberhardt et al. 1982). Combined with the foxes' preference for sandy-loam soils (unlike the gravel pads) for denning (Chesemore 1967) and the likelihood that the fox will den away from areas of human interference, this limits the use of the sites by the Arctic fox.

Areas that provide potentially suitable habitat for the Arctic fox include the Old Landfill, Current Landfill, Contaminated Ditch, Old Runway Dump, Heated Storage, Fuel Tanks, Old Dump Site, and Bladder Diesel Spill.

The sites that are not likely to be suitable habitat for the Arctic fox are the POL Catchment, Garage, Weather Station Building, White Alice Facility, POL Tanks, and JP-4 Spill. These sites are not likely to have fox prey populations or the foxes' preferred soils for denning.

Caribou. The major migration route and staging areas for barren-ground caribou are to the south of Barter Island. Caribou, a representative species for all the DEW Line installations, will not be evaluated in the risk assessment for the Barter Island installation because they are uncommon, rarely making the crossing to the island. Those that do cross are likely to avoid the developed area around the installation (Cameron et al. 1992). Caribou are an important subsistence resource, and any caribou venturing near the DEW Line site, and the village of Kaktovik, would probably be harvested by the local people.

3.2.4.4 Summary of Habitat Characterization and Screening. Based on the evaluations of habitat suitability, none of the representative species are expected to use habitat at four of the sites: the Weather Station Building, White Alice Facility, POL Tanks, or JP-4 Spill. The exposure pathways are not complete for the representative species at these sites, so they are screened from the ERA. The potential future ecological risks associated with the sites, should they become suitable habitat in the future, is presented in Section 3.4.5. The remaining 10 sites are assumed to have suitable habitat for at least some of the representative species.

The ERA is being conducted for the entire Barter Island installation, but only a small portion of the facility consists of sites. If the ERA considered the entire installation to be of concern, potential exposures may be overestimated. However, overestimated exposures may be mitigated by taking into account the spatial extent of the sites in relation to the size of the entire installation, in addition to screening via the habitat evaluation. The areal extent of the Barter Island DEW Line installation is estimated to be 140 ha. The sites that have habitat suitable for the representative species are estimated to total less than 3.5 ha., and those areas that have been determined to be unsuitable for all the representative species total less than 0.2 ha. In general, based on professional judgement and onsite observations, but not on site-specific surveys, the installation provides habitat less suitable than nearby areas because of the numerous roads, gravel pads, and overall development. The spatial extent of the sites was considered when estimating the onsite dietary intake (IS) in Section 3.2.7.2.

Although some of the sites may provide suitable habitat for the representative species, development and activities at the installation may lessen the overall attractiveness of these sites. Human presence and activity will deter many species from using the habitats. The region surrounding the installation meets the needs of the representative species, and it is likely that the species will show a preference for the undeveloped areas over the installation lands. On the other hand, there may be some sites that are more attractive to wildlife. Examples are the landfill, where refuse may attract carrion eaters and towers that may provide perches for peregrine falcons. Without detailed study and research, habitat selection behavior is not quantifiable.

3.2.5 Exposure Assessment for Representative Species of Plants

The harsh environment of the Arctic Coastal Plain imposes many restrictions on plant life. The presence of permafrost limits infiltration and percolation of water, so the water table is often at or above the surface. The vast majority of plant species are perennial, with much of their biomass (50 to 98 percent) underground (Raven et al. 1986). The potential pathways of contamination for plants are through the soil/sediment and surface water.

Carex spp., *Salix* spp., and *Eriophorum* spp. all store food reserves in rhizomes. Mycorrhizal fungi play an important role in the transport and delivery of nutrients to the rhizomes and the roots of these species. This underground system is most likely a development in response to the harsh aboveground arctic environment. As a result, surface water contaminated with lighter than water chemicals (i.e., petroleum and its derivatives) does not present a greatly increased hazard to the below ground portion of plants. This has been shown experimentally by exposing arctic coast vegetation to petroleum products (Walker et al. 1978). The experiments showed that sedges, willows and cottongrass plants were not adversely affected by low to moderate amounts

of petroleum in wet environments. Thus, soil will be considered the primary pathway of potential contamination for plants. The chemical concentration used in the risk characterization (Section 3.4) is the average concentration of the COC in the soil/sediment.

3.2.6 Exposure Assessment for Representative Aquatic Organisms

Organisms that dwell in an aquatic environment are exposed to chemicals contained in the water column. For this reason, the exposure assessment considers the concentrations in surface water to be the exposure concentrations to aquatic organisms. The risk assessment compares the average concentration of the COCs found in surface waters to toxicity data for the representative aquatic species to determine the risk.

3.2.7 Exposure Estimates for Representative Bird and Mammal Species

Exposure estimates for the representative species of mammals and birds (expressed as a unit of chemical ingested per unit of body weight) are based on their total exposure to COCs from diet, soils and surface water using the following equation:

$$DI = [(FI \times CD) + (WI \times CW) + (SI \times CS)] \times IS / BW$$

where:

DI	=	daily intake (kg/day).
FI	=	food intake rate (kg/day); rates are derived in the life history tables (3-6 through 3-17). Diets that consist of multiple food groups are proportioned according to the diet composition information in the life history tables.
CD	=	chemical concentration in diet (mg/kg); based on concentrations for each group of food items.
WI	=	water intake rate (L/day); rates are derived in the life history tables.
CW	=	chemical concentration in water (mg/L).
SI	=	soil/sediment intake rate (kg/day); based on a percentage of dietary intake.
CS	=	chemical concentration in soil/sediment (mg/kg).
IS	=	percent of dietary intake at sites on the site (%).
BW	=	body weight (kg)

In the case of species that have partial herbivorous dietary intakes, the CD value is multiplied by the proportion of vegetation in their diet. Those species and their respective proportions are: Lapland longspur, 0.25; brant, 0.90; glaucous gull, 0.10; pectoral sandpiper, 0.10; spectacled eider, 0.25; and Steller's eider, 0.25 (see the Life History tables, 3-7 to 3-13, for references regarding the proportion of vegetation in the species' diets). The estimated exposure calculations for bird and mammal receptors are presented in Appendix C.

3.2.7.1 Potential Bioaccumulation of COC in Representative Species.

Bioaccumulation factors are considered for carnivorous species that may ingest contaminants in prey.

The potential risk from secondary exposure to the COCs is difficult to determine because of the complexity of the trophic web. Inputs to the exposure estimate equation include concentrations

of contaminants in water and soil, ingestion rates for water, food and soil, range of the species and onsite areas, and body weight. The food ingested, in the case of higher level consumers, may be composed of elements from different levels of the trophic web. For example, a contaminant may be taken up by a plant, which is eaten by a lemming, which is then eaten by an Arctic fox. The amount of contaminant to which the fox is exposed, without supporting empirical data at each trophic level, is not readily quantified. Based on this and the lack of data to assist in quantifying bioaccumulation, the risk assessment does not account for bioaccumulation in the animal portion of the trophic web.

For illustrative purposes, bioconcentration factors (BCF) were calculated (Veith et al. 1979) for the organic COCs and presented in Table 3-6.

The metals that are of concern at Barter Island (aluminum, iron, lead, manganese, and zinc) are evaluated for plant uptake quantitatively, as discussed in Section 3.2.5. The bioaccumulation of metals in the animal portion of the trophic web is not amenable to quantification without sample concentrations at each level of consumer. This does not present as large a problem as it initially seems because the bioaccumulative properties of the metals of concern can be addressed qualitatively.

Aluminum is not likely to bioaccumulate because, taken orally, it is poorly absorbed through the gastrointestinal tract. Most ingested aluminum leaves an animal's body in the feces, and the little that does enter the bloodstream is excreted in urine [Agency for Toxic Substances and Disease Registry (ATSDR) 1990a].

Information about the bioaccumulation of iron is not available in the literature, but because iron is an essential nutrient, it is likely that the metabolic processes that make use of iron will prevent undue bioaccumulation.

Lead tends to accumulate in bone (Talmadge and Watson 1991), so ingestion of animal tissue would not contribute greatly to increased lead concentrations. Food chain biomagnification of lead is uncommon in terrestrial communities (Eisler 1988). Kraus (1989) showed that in environments that were high in lead, the concentration of lead in insects and the tissues of insectivorous birds was low. Thus, lead is not likely to bioaccumulate to a degree that could contribute to risk at Barter Island.

Manganese and zinc are considered essential nutrients (ATSDR 1990b; Eisler 1993). It is not likely that the concentrations of these metals found at the Barter Island installation could bioaccumulate because animal systems are conditioned to regulate these minerals for metabolic use.

The exposure estimates for organic chemicals do not include potential bioaccumulation of chemicals in the animal portion of the trophic web. The estimates are at least one order of magnitude below the TRVs. It is unlikely that the organic chemicals will bioaccumulate (based on the concentrations reported in the soil and water), such that the exposure estimates would exceed, or even approach, the TRVs.

TABLE 3-6. BIOCONCENTRATION FACTORS FOR SELECTED ORGANIC COMPOUNDS IN WATER

CHEMICAL	Log K _{ow}	BCF
1,2-Dichloroethane	1.45	7.5
Phenol	1.48	7.9
Methylene chloride	1.51	8.3
Benzene	2.13	24.5
Toluene	2.69	65.2
Ethylbenzene	3.15	146
2-methylnaphthalene	3.87	514
Naphthalene	3.36	211
Trimethylbenzene	3.78	439
Xylene	3.16	149
Pyrene	5.18	5,091
DRPH	5.30	6,238
Hexachlorobenzene	5.5	8,913
Endrin	5.61	10,804
Benzo(a)anthracene	5.61	10,804
Chrysene	5.61	10,804
DDT	5.98	20,644
Benzo(a)pyrene	6.25	33,113
Benzo(g,h,i)perylene	6.51	52,191
PCB (Aroclor 1254)	6.94	110,764

Note: BCF calculated from Log K_{ow} according to the following equation:

$$\text{Log BCF} = 0.76 \text{ Log } K_{ow} - 0.23 \quad (\text{Veith et al. 1979 in Spacie and Hamelink 1985})$$

K_{ow} = octanol/water partition coefficient

The BCF for PCBs is high ($>110,000$), which indicates that bioaccumulation may be significant. The TRV ratios (comparison of the COC exposure of a species to the Toxicity Reference Value, presented in Table 3-30) for PCBs, however, are three to seven orders of magnitude less than one. Thus development of adverse effects in ecological receptors is not likely and the bioaccumulation of PCBs will probably not be sufficient to produce a TRV ratio greater than one.

3.2.7.2 Estimation of Percent Ingested Onsite. The habitat suitability evaluation determines the sites a given species might use. The evaluation provides species-specific "exposure areas". This information, combined with home range data and the extent of the sites, can be used to estimate the percent of dietary intake that a species gets from the sites. This estimate is referred to as the "percent of dietary intake at sites on site" (IS) value in the exposure estimate equation. The IS value is the total extent of the sites (3.5 ha) as a percentage of the reported home range size. When home range information for a species is not available, population density values have been converted to estimate areal extent of the species. The representative species are most likely to be present at Barter Island during the breeding season. During the breeding season, most species become territorial. These territories are represented in the density of the population and may be used as substitute values for home range areas. This presents an added degree of uncertainty (see Section 3.5.3). If the home range (or converted population density value) is less than 3.5 ha, the maximum value for IS will be 1.0. The IS values for the representative bird and mammal species are:

Birds. Lapland longspur. IS = 0.5; Derksen et al. (1981) report a breeding density of $38.6/\text{km}^2$. This corresponds to about 1 bird/2.6 ha. Potentially, the longspur could meet all its dietary demands within the sites only if they are all contiguous. Nevertheless, an IS value of 0.5 is used because the longspur prefers drier upland habitat over the wetter areas where the majority of the contaminant pathways occur.

Brant. IS = 0.18; density of breeding pairs reported by Derksen et al. (1981) is $5.0/\text{km}^2$. At this density of 1 brant/20 ha, the total extent of the sites is about 18 percent of the area a brant might use.

Glaucous gull. IS = 0.30; the density for the glaucous gull is reported by Derksen et al. (1981) as $0.8/\text{km}^2$. This density, about 1 gull/125 ha, yields a value of three percent (about 35 times the area). This may be misleading because the gull is attracted to refuse and areas where scavenging opportunities are high. The landfills and potential for mishandling of refuse may attract the glaucous gull. To compensate for this, the IS has been adjusted by a factor of 10, from 0.03 to 0.30.

Pectoral sandpiper. IS = 0.78; the density of the pectoral sandpiper along the Arctic Coastal Plain is reported by Derksen et al. (1981) as $22.4/\text{km}^2$. This density equates to one sandpiper/4.5 ha, and an IS value of 0.78.

Peregrine falcon. IS = 0.10; the peregrine falcon forages over very large areas. At level flight speeds up to 60 mph (Craig 1994 pers. comm.), the falcon can cover great distances in search of prey. Likely prey at the Barter Island installation includes shorebirds. The low levels of

bioaccumulation of the COC will limit the exposure the peregrine receives in its diet. The falcon's relatively low potential for exposure, however, is balanced by the potential attraction of the towers and structures as nest sites or resting perches (Craig 1994 pers. comm.), so an IS of 0.10 has been assigned. This value may overestimate the exposure to the peregrine, but a conservative approach is warranted in the evaluation of a threatened species.

Spectacled eider. IS = 0.01; Derksen et al. (1981) report an average population density of 0.32/km² for the spectacled eider. The resulting density of 1 bird/312.5 ha in 1981 may currently be too high, considering the decline in the species' population. The spectacled eider is a threatened species, so the overestimation of exposure is appropriate in evaluation of a sensitive species.

Steller's eider. IS = 0.01; Steller's eider is rare worldwide. Its occurrence at the Barter Island installation is uncommon, although some birds may breed nearby. As noted earlier, Steller's eider breeds primarily in Siberia, with Barrow, Alaska being the only other location where the bird is described as a "regular" breeder (USFWS 1992). The breeding density near Barrow is reported as 0.22/km². Using this value for Barter Island is likely to result in an overestimation of exposure, but this is justified because of the species' special population status.

Mammals. Brown lemming. IS = 0.5; the lemming's home range is reported as 0.5 ha (Nowak 1991). It is possible that several lemmings may consume all their dietary needs within the bounds of a site. However, the lemming is not likely to use the wetter sites (which constitute well over 50 percent of the total extent of the sites), which have the majority of the contaminant pathways. The sites are mostly gravel pads that have been constructed for development purposes, support little or no vegetation, and offer a poor matrix for the lemming to use for burrowing. For these reasons the IS used for the brown lemming is 0.5 rather than 1.0.

Arctic fox. IS = 0.01; the home range of the fox is extremely variable. Eberhardt et al. (1982) report a home range of 3.7 to 20.8 km² for juvenile and adult Arctic foxes, respectively. Even the lower end of this range is much greater than the extent of the sites.

3.2.7.3 Exposure Assessment for Representative Species of Birds. In this section the methods for quantifying exposures to the selected representative species of birds are presented. The threatened and endangered bird species, peregrine falcon, spectacled eider, and Steller's eider are included in this section.

In order to estimate exposures of the representative species of birds, life history information was compiled for the selected species. This information includes: occurrence at the DEW Line sites, habitat, average body weight, estimated food intake rate, estimated WI rate, diet composition, and population density. This life history information was used to estimate the representative species' potential exposure to the COC.

Plant uptake of contaminants has been quantified for use in the exposure estimations for herbivores (bird and mammal species). Herbivores are potentially exposed to contamination directly from ingestion of soil and water intake as well as through their diet. The dietary component (CD in the exposure estimate equation) is calculated by multiplying the contaminant's

soil concentration by the BCF, B_v . B_v is defined as the ratio of the concentration in aboveground parts of a plant (mg of compound/kg of dry plant) to the concentration in soil (mg of compound/kg of dry soil). The B_v can be used to predict the level of a potential contaminant taken up by a plant, and this information can then be used to assess the potential transport of the contaminant in the trophic web.

The uptake of metals by plants is quantified using the B_v values in Baes et al. (1984). The approach for organic chemicals is basically the same, except that the B_v s for organic chemicals are derived through a regression equation (Travis and Arms 1988). The equation is:

$$\log B_v = 1.588 - 0.578(\log K_{ow})$$

where:

$$B_v = \text{the BCF (unitless); and}$$

$$K_{ow} = \text{the octanol-water partition coefficient of the chemical (unitless);}$$

In order to calculate the potential uptake of DRPH by plants, the K_{ow} of diesel fuel was estimated. The estimation of the K_{ow} was conducted using equation 2-3 in Lyman et al. (1982):

$$\log S = -0.922 \log K_{ow} + 4.184$$

where:

$$S = \text{solubility (mg/L)}$$

$$K_{ow} = \text{octanol/water partition coefficient (unitless)}$$

This equation estimates the solubility of an organic chemical in water. However, it may also be manipulated arithmetically to calculate the $\log K_{ow}$ based on the known solubility:

$$\log K_{ow} = \frac{\log S - 4.184}{-0.922}$$

The solubility of diesel fuel (0.2 mg/L) (Custance et al. 1992) was used to calculate the $\log K_{ow}$ of diesel fuel. The $\log K_{ow}$ is calculated to be 5.29.

Life history information for the Lapland longspur, brant, glaucous gull, and pectoral sandpiper is presented in Tables 3-7 through 3-10. Life history information for the peregrine falcon, spectacled eider, and Steller's eider is presented in Tables 3-11 through 3-13.

Information is not available on the daily food intake rate (grams/day) and water intake rate (liters/day) for the representative bird species in the Arctic habitat. Therefore this information was estimated using regression equations associated with body weight. The severity of the Arctic climate may impose higher metabolic demands on animals. As a result, the food and water intake rates should be considered to be estimates only and their uncertainty should be kept in mind. The food intake rate was estimated using Nagy's (1987) equations:

Passerine birds:

$$FI \text{ (g/day dry matter)} = 0.398 \times (\text{body weight in grams})^{0.850}$$

TABLE 3-7. LIFE HISTORY INFORMATION FOR THE LAPLAND LONGSPUR, *Calcarius lapponicus*

PARAMETER	VALUE	NOTES	REFERENCE
OCCURRENCE AT DEW LINE SITES	seasonal breeder at all eight sites	dominant breeding passerine	USAF 1993a
HABITAT	breeds on Arctic coastal tundra		Scott 1983
BODY WEIGHT	27.3 g (0.027 kg)	mean of 68 specimens	Dunning 1984
FOOD INTAKE RATE	6.6 g/day dry matter	$FI=0.398(BWg)^{0.850}$	Nagy 1987
WATER INTAKE RATE	0.005 liters/day	$WI=0.059(BWkg)^{0.67}$	Calder and Braun 1983
DIET COMPOSITION	during breeding (June and July): insects (crane flies); pre- and post-breeding (May and August): seeds (grasses); average 25% vegetation in diet	passerine member of insectivorous foraging guild which includes shorebirds	Custer and Pitelka 1978
POPULATION DENSITY	38.6/km ²	varies with changing predation pressures	Derksen et al. 1981

TABLE 3-8. LIFE HISTORY INFORMATION FOR THE BRANT, *Branta bernicla*

PARAMETER	VALUE	NOTES	REFERENCE
OCCURRENCE AT DEW LINE SITES	seasonal, breeds at or near all eight sites	breeding, migratory sp., subsistence sp.	USAF 1993a
HABITAT	breeds on Arctic Coastal Plain	prefers low, barren, wet, coastal terrain	Palmer 1976
BODY WEIGHT	1,305 g (1.305 kg)	mean of 791 specimens	Dunning 1984
FOOD INTAKE RATE	69.2 g/day dry matter	$FI=0.648(BWg)^{0.651}$	Nagy 1987
WATER INTAKE RATE	0.07 liters/day	$WI=0.059(BWkg)^{0.67}$	Calder and Braun 1983
DIET COMPOSITION	sedges, grasses; average 90% vegetation in diet	some insects during breeding (June and July)	Palmer 1976
POPULATION DENSITY	5.0/km ²	average from 3 coastal sites	Derksen et al. 1981

TABLE 3-9. LIFE HISTORY INFORMATION FOR THE GLAUCOUS GULL, *Larus hyperboreus*

PARAMETER	VALUE	NOTES	REFERENCE
OCCURRENCE AT DEW LINE SITES	seasonal breeder and migrant at all eight sites		Woodward-Clyde 1993
HABITAT	coastal tundra, lakes, ponds, and marine environment	breeds on Arctic coast	Farrand 1983
BODY WEIGHT	1,445 g (1.445 kg)	mean of 65 specimens	Dunning 1984
FOOD INTAKE RATE	73.9 g/day dry matter	$FI=0.648(BWg)^{0.651}$	Nagy 1987
WATER INTAKE RATE	0.08 liters/day	$WI=0.059(BWkg)^{0.67}$	Calder and Braun 1983
DIET COMPOSITION	small fish, birds, insects, crustaceans, mollusks, garbage; average 10% of vegetation in diet	predatory scavenger	Martin et al. 1961
POPULATION DENSITY	0.8/km ²	average from 3 coastal sites	Derksen et al. 1981

TABLE 3-10. LIFE HISTORY INFORMATION FOR THE PECTORAL SANDPIPER, *Calidris melanotos*

PARAMETER	VALUE	NOTES	REFERENCE
OCCURRENCE AT DEW LINE SITES	seasonal breeder at all eight sites	abundant on Arctic Coastal Plain	Woodward-Clyde 1993
HABITAT	grassy margins of wet meadows, marshes, riparian areas, ponds	nests hidden on well-drained grassy sites	Scott 1983; Martin et al. 1961
BODY WEIGHT	79 g (0.079 kg)	mean of 35 specimens	Dunning 1984
FOOD INTAKE RATE	11.1 g/day dry matter	$FI=0.648(BWg)^{0.651}$	Nagy 1987
WATER INTAKE RATE	0.01 liters/day	$WI=0.059(BWkg)^{0.67}$	Calder and Braun 1983
DIET COMPOSITION	insects, mollusks, crustaceans, worms, vegetable debris; average 10% vegetation in diet	crane flies are major diet component	Martin et al. 1961; Pitelka 1959
POPULATION DENSITY	22.4/km ²	average from 3 coastal sites	Derksen et al. 1981

TABLE 3-11. LIFE HISTORY INFORMATION FOR THE PEREGRINE FALCON, *Falco peregrinus*

PARAMETER	VALUE	NOTES	REFERENCE
OCCURRENCE AT DEW LINE SITES	potentially at all sites; breeding pair at Barter Island installation in 1991, 92, 93	will use man-made structures for nesting, resting and foraging	Woodward-Clyde 1993
HABITAT	nests on cliffs, banks, and tall structures near coasts, lakes, rivers, marshes	migrates to southern hemisphere	Farrand 1983
BODY WEIGHT	820 g (0.82 kg)	mean of 31 specimens	Dunning 1984
FOOD INTAKE RATE	51.1 g/day dry matter	$FI=0.648(BWg)^{0.651}$	Nagy 1987
WATER INTAKE RATE	0.05 liters/day	$WI=0.059(BWkg)^{0.67}$	Calder and Braun 1983
DIET COMPOSITION	carnivorous predator of small-medium sized birds	most prey taken in air	Farrand 1983
POPULATION DENSITY	2 birds per 51.8 km ² (average)	value is for a breeding pair, may be higher out of breeding season	U.S. Army Corps of Engineers (COE) 1991

TABLE 3-12. LIFE HISTORY INFORMATION FOR THE SPECTACLED EIDER, *Somateria fischeri*

PARAMETER	VALUE	NOTES	REFERENCE
OCCURRENCE AT DEW LINE SITES	potential seasonal breeder at all sites	winter whereabouts unknown	Woodward-Clyde 1993
HABITAT	marine when not breeding, nests on coastal tundra	nests on islets in tundra ponds and lakes, as well as ashore	Palmer 1976
BODY WEIGHT	1,375 g (1.375 kg)	mean of 32 specimens	Dunning 1984
FOOD INTAKE RATE	71.6 g/day dry matter	$FI=0.648(BWg)^{0.651}$	Nagy 1987
WATER INTAKE RATE	0.07 liters/day	$WI=0.059(BWkg)^{0.67}$	Calder and Braun 1983
DIET COMPOSITION	75% insects, mollusks, crustaceans; average of 25% plant matter in diet	mostly insects when they are abundant; June and July	Kistchinski and Flint 1974
POPULATION DENSITY	0.32/km ²	average of three coastal sites	Derksen et al. 1981

TABLE 3-13. LIFE HISTORY INFORMATION FOR STELLER'S EIDER, *Polysticta stelleri*

PARAMETER	VALUE	NOTES	REFERENCE
OCCURRENCE AT DEW LINE SITES	small potential as seasonal breeder at all sites	known to nest near Barrow	Woodward-Clyde 1993
HABITAT	coastal tundra, near lakes and ponds		Scott 1983
BODY WEIGHT	805 g (0.805 kg)	mean of 90 specimens	Dunning 1984
FOOD INTAKE RATE	50.5 g/day dry matter	$FI=0.648(BWg)^{0.651}$	Nagy 1987
WATER INTAKE RATE	0.05 liters/day	$WI=0.059(BWkg)^{0.67}$	Calder and Braun 1983
DIET COMPOSITION	crustaceans, mollusks; average of 25% plant matter in diet	primarily insects when they are abundant; June and July	Palmer 1976
POPULATION DENSITY	0.22/km ²	average at three coastal sites	Derksen et al. 1981

All other birds:

$$FI \text{ (g/day dry matter)} = 0.648 \times (\text{body weight in grams})^{0.651}$$

The water intake rate was estimated using the regression equation developed by Calder and Braun (1983):

All birds:

$$WI \text{ (liters/day)} = 0.059 \times (\text{body weight in kilograms})^{0.67}$$

As animals forage they may incidentally ingest soil and sediment particles. The average concentration of contaminants in soil/sediment can be multiplied by the amount of soil/sediment ingested to estimate the potential uptake of contaminants by this route. Soil intake rates have been reported for just a few wildlife species (Beyer et al. 1991). The soil ingestion rates for the representative species are extrapolated from Beyer et al. (1991) by using similar species with reported values. The rates are reported as percentages of total diet. Table 3-14 lists the representative bird species, the species used as surrogates, and the estimated percentages of soil ingested in quantifying exposure to contaminants. Species that forage directly in the soil or sediment, such as the sandpiper or goose, show relatively high percentages of soil in their diet. The Lapland longspur and peregrine falcon do not have appropriate surrogate species with soil ingestion data. The foraging habits of the longspur and falcon minimize their soil contact and their estimates of soil ingestion reflect this. The glaucous gull ingests stones and sand as a mechanical addition to its diet (Belopol'skii 1961).

TABLE 3-14. SOIL INGESTION BY REPRESENTATIVE BIRD SPECIES

REPRESENTATIVE SPECIES	SURROGATE SPECIES	ESTIMATED % OF SOIL IN DIET
Lapland longspur	no suitable surrogate	<2.0
Brant	Canada goose	8.2
Glaucous gull ^a	Siberian glaucous gull	7.6
Pectoral sandpiper	4 sandpiper species	13.1 (average)
Peregrine falcon	no suitable surrogate	<2.0
Spectacled eider	Canada goose	8.2
Steller's eider	Canada goose	8.2

Source: Beyer et al. 1991

^a information from Belopol'skii 1961.

3.2.7.4 Exposure Assessment for Representative Species of Mammals. This section assesses exposure to contaminants for the selected representative species of mammals (brown lemming and Arctic fox; barren-ground caribou are not include in the exposure assessment as noted previously). Tables 3-15 (brown lemming) and 3-16 (Arctic fox) present life history data that were used to evaluate the representative species and their potential pathways of exposure to contaminants. Table 3-17 (barren-ground caribou) is included as information on a representative species at the DEW Line installations, although caribou are not expected to occur at the Barter Island installation. Home range and/or population density has been listed for the representative mammal species, depending on appropriateness and availability.

Information on the daily food intake rates of the Arctic fox is not available. The rate has been estimated using regression equations associated with average body weights (Nagy 1987). The food intake rates for the fox estimated using the following equation developed by Nagy (1987) for mammals in general.

Arctic fox:

$$FI \text{ (g/day dry matter)} = 0.235 \times (\text{body weight in grams})^{0.822}$$

Because of very low assimilation efficiencies, the low nutrient content of winter forage, and the high metabolic demands in Arctic habitats (Chappell 1980), the equation for food intake rate significantly underestimates the rate for the brown lemming. A more appropriate rate for the brown lemming is reported by Chappell (1980).

TABLE 3-15. LIFE HISTORY INFORMATION FOR THE BROWN LEMMING, *Lemmus trimucronatus*

PARAMETER	VALUE	NOTES	REFERENCE
OCCURRENCE AT DEW LINE SITES	resident at all eight sites	dominant small mammal	USAF 1993a
HABITAT	tundra and alpine meadows	nests aboveground in winter, below in summer	Burt and Grossenheider 1976
BODY WEIGHT	55 g (0.055 kg)		Chappell 1980
FOOD INTAKE RATE	24-45 g/day dry matter	has low assimilation efficiencies (31-36%), variation also related to seasons	Chappell 1980
WATER INTAKE RATE	0.007 liters/day	$WI=0.099(BWkg)^{0.90}$	Calder and Braun 1983
DIET COMPOSITION	sedges, grasses, lichens, roots, leaves, bark, berries		Nowak 1991
HOME RANGE SIZE (AVG)	0.5 ha (females) 1.0 ha (males)	0.5 ha used in assessment	Nowak 1991
POPULATION DENSITY	0 to 325/ha	populations have large fluctuations on a 3-5 year cycle; currently populations are low	Nowak 1991; Snyder-Conn 1994

TABLE 3-16. LIFE HISTORY INFORMATION FOR THE ARCTIC FOX, *Alopex lagopus*

PARAMETER	VALUE	NOTES	REFERENCE
OCCURRENCE AT DEW LINE SITES	resident at all eight sites	ubiquitous	USAF 1993a
HABITAT	tundra and coastal plain	dens in sandy mounds >1 m high	Chesemore 1967
BODY WEIGHT	4950 g (4.95 kg)		Burt and Grossenheider 1976
FOOD INTAKE RATE	256 g/day dry matter	FI= $0.235(BWg)^{0.822}$	Nagy 1987
WATER INTAKE RATE	0.42 liters/day	WI= $0.099(BWkg)^{0.90}$	Calder and Braun 1983
DIET COMPOSITION	brown lemming (summer), nesting birds, carrion, seal pups, non-food items	brown lemming in >85% of all scats, n=224	Chesemore 1967; Nowak 1991
HOME RANGE SIZE (AVG)	20.8 km ² adult 3.7 km ² juvenile (<1 yr)	adult range used in assessment	Eberhardt et al. 1982

TABLE 3-17. LIFE HISTORY INFORMATION FOR THE BARREN-GROUND CARIBOU,
Rangifer tarandus

PARAMETER	VALUE	NOTES	REFERENCE
OCCURRENCE AT DEW LINE SITES	seasonal, at or near all eight sites during migrations	some sites used for calving	USAF 1993a
HABITAT	tundra in summer, open coniferous forest in winter	varies much related to migration	Burt and Grossenheider 1976
BODY WEIGHT	95.5 kg	mean for adults, male and female	Nowak 1991
FOOD INTAKE RATE	2.9 kg/day dry matter	FI= $0.0687(BW)^{0.822}$	Nagy 1987
WATER INTAKE RATE	6.0 liters/day	WI= $0.099(BW)^{0.90}$	Calder and Braun 1983
DIET COMPOSITION	willows, sedges, cottongrass, lichens	selection based on plant phenology	Skogland 1980; White and Trudell 1980
POPULATION DENSITY	1.41 km ² 0.31 km ² 4.53 km ²	undisturbed calving area within 1 km of road within 5-6 km of road (this value to be used at Barter Island)	Cameron et al. 1992

The rates for water intake of the representative mammals were estimated using the regression equation generated by Calder and Braun (1983) because of the unavailability of species-specific information in the literature. The equation is:

$$WI \text{ (liters/day)} = 0.099 \times (\text{body weight in kilograms})^{0.90}$$

Incidental soil intake was evaluated for mammals in the same manner as for birds (Section 3.2.7.3 and Table 3-14). Table 3-18 shows the percent of soil ingested for the representative mammal species.

3.3 ECOLOGICAL TOXICITY ASSESSMENT

This section presents toxicity information for each COC in surface water and soils/sediments. COCs in surface water (Section 3.1.1) are DRPH, aluminum, iron, and manganese. COCs in soils/sediment (Section 3.1.2) are DRPH, GRPH, RRPB, ethylbenzene, xylenes, trimethylbenzene, naphthalene, 2-methylnaphthalene, lead, zinc, and PCBs. Sections 3.3.1 through 3.3.12 discuss the toxicity of all COCs to the receptor groups. Section 3.3.13, presents the derivation of TRVs used for this ERA.

3.3.1 Petroleum Hydrocarbons

Section 3.1 presented the COCs for sites at the Barter Island installation. DRPH were identified as a COC in surface water, and DRPH, GRPH, and RRPB were identified as COCs in soils and/or sediments. This section is a discussion of the chemical differences between DRPH, GRPH, and RRPB, and the toxicity of these three petroleum mixtures.

Crude petroleum contains thousands of different chemical compounds. Gasoline and diesel oil are refined petroleum products. The composition of gasoline and diesel fuel depends not only on the origin of the crude oil from which the gasoline is derived, but also the process technique and the blending scheme (Von Burg 1993). Once gasoline or diesel fuel is released to the environment, weathering and volatilization further alter its composition.

TABLE 3-18. SOIL INGESTION FOR REPRESENTATIVE MAMMAL SPECIES

REPRESENTATIVE SPECIES	SURROGATE SPECIES	ESTIMATED % OF SOIL IN DIET
Brown lemming	prairie dog	2.7
Arctic fox	fox (species unreported)	2.8
Caribou	elk	< 2.0

Source: Beyer et al. 1991

GRPH contain hydrocarbon molecules with 4 to 10 carbon atoms and include the following chemical classes: paraffins (straight-chained alkenes), olefins (straight-chained alkenes), naphthenes (cycloalkanes and alkenes), and aromatic hydrocarbons (alkylbenzenes and polynuclear) (Von Burg 1993). As many as 140 compounds have been identified as constituents of gasoline. The methods used by the analytical laboratory detected GRPH with 6 to 9 carbon atoms. Diesel fuel is a complex, variable mixture of the same classes of compounds containing 6 to 21 carbon atoms. The analytical method detected DRPH with 10 to 24 carbon atoms. As many as 45 compounds have been identified as constituents of diesel fuel (Von Burg 1993). The concentration of RRPH was calculated by subtracting the GRPH and DRPH concentrations from the total petroleum hydrocarbon concentration (TPH). Therefore the RRPH could include many different types of chemicals, although the majority of molecules would include 24 carbon atoms or more.

Table 3-19 presents the chemical classes and weight percent for GRPH and DRPH. Generally, gasoline contains more aromatic compounds and simple chained alkanes, whereas diesel fuel is characterized by cycloparaffins or cycloalkanes. Both gasoline and diesel fuel will be affected by the environment. Weathering will change the chemical composition of petroleum and, specifically, concentrations of aromatic compounds such as benzene will decrease as a result of volatilization.

Available toxicity test data have been derived from pure, fresh product and therefore, the applicability to the weathered product encountered at Barter Island is questionable. Gasoline is the most studied of the petroleum products, however most data are based on inhalation studies. Gasoline was classified by EPA (1992b) as a Group C (possible human) carcinogen whereas diesel oil was classified as Group D (not classifiable as to human carcinogenicity). Presumably, this classification of gasoline is due to benzene which, under the conditions of environmental exposure, would volatilize more rapidly than any other constituent. The gasoline and diesel petroleum hydrocarbon data from surface water and sediment samples collected at Barter Island indicate that benzene was not detected frequently at concentrations above either background or action levels. Physical-chemical data from the literature indicates that TPH in soil would reflect all constituents with eventual loss of aromatic (e.g., BTEX) components first, lighter alkanes second, lighter polycyclic aromatic hydrocarbons (PAHs) third, followed by naphthalenes. For an old spill, TPH measurements may reflect predominantly trace amounts of high molecular-weight PAHs or higher molecular-weight and branched alkanes [Massachusetts Department of Environmental Protection (MDEP) 1993].

For the purposes of ranking the toxicity of GRPH, DRPH, and RRPH, it is assumed that BTEX and lighter-weight alkanes were significantly weathered from exposure to the arctic environment, and that toxicity is more dependent upon noncarcinogenic endpoints associated with alkanes and cycloalkanes. MDEP (1993) reviewed noncarcinogenic toxicological endpoints and determined that diesel oil was an order of magnitude more toxic than gasoline. Other sources indicate that the toxicity of alkanes and cycloalkanes is similar (Armstrong Laboratory 1994; Sax and Lewis 1989). Based on the MDEP review and the chemical data reported for the Barter Island surface water and sediment samples, DRPH are evaluated and used to represent ecological risks from petroleum hydrocarbon contamination conservatively.

TABLE 3-19. CHEMICAL CLASSES OF GRPH AND DRPH

CHEMICAL CLASS	WEIGHT PERCENT (%) ^a
GRPH ^b	
Normal paraffins (n-alkanes)	19.3-38.4 (28.8)
Isoparaffins (isoalkanes)	11.5-50.3 (30.9)
Naphthenes (cycloparaffins or cycloalkanes)	1.0-2.8 (1.9)
Aromatics (benzene, toluene, pyrene)	9.7-54.7 (32.2)
DRPH ^c	
Normal paraffins (n-alkanes)	5.6
Isoparaffins (isoalkanes)	11.1
Naphthenes (cycloparaffins or cycloalkanes)	46.3
Aromatics (benzene, toluene, pyrene)	33.3
Nitrogen, sulfur and oxygen compounds	3.7

^a Average shown in parentheses.

^b Heath et al. 1993

^c Weeks et al. 1988

As discussed above, diesel is comprised of a complex mixture of paraffins (straight-chained alkanes), olefins (straight-chained alkenes), naphthenes (cycloalkanes and alkenes), and aromatic (alkylbenzenes, and polynuclear) petroleum hydrocarbons containing 6 to 21 carbon atoms. Hydrocarbons containing 8 to 18 carbon atoms predominate (Von Burg 1993). There are six grades of diesel fuel (Diesel Oil No. 1, Diesel Oil No. 2, Diesel Oil No. 4, Fuel Oil No. 1, Fuel Oil No. 2 and Home Heating Oil) (Von Burg 1993). The specific components of diesel are expected to change from source to source, so the toxicity of diesel fuels is expected to be variable.

3.3.1.1 Plants. Petroleum released to the aquatic environment is expected to be toxic to aquatic plants. Toxicity tests have shown that the water soluble components of petroleum are toxic to an algal species (*Chlorella vulgaris*) (Kauss and Hutchinson 1975). However, in this specific study, the toxicity was short term. The algal community recovered after a "lag phase".

It was theorized (Kauss and Hutchinson 1975) that this trend was due to the loss of highly volatile fractions from the testing chamber over time. Exposure to water extracts of No. 2 Fuel Oil depressed algal biomass in communities and resulted in blue-green algal dominance and decreased diatom occurrence (Bott and Rogenmuser 1978).

3.3.1.2 Aquatic Organisms. Moles et al. (1979) tested the acute toxicity of Prudhoe Bay crude oil to several Alaskan freshwater and anadromous fish. Salmonids were the most sensitive species tested, and demonstrated median tolerance limits (the concentration at which one half the organisms survive in 96 hours: LC_{50}) ranging from 2.7 to 4.4 mg/L. The three-spined stickleback was more tolerant, with an LC_{50} of 10.4 mg/L. Klein and Jenkins (1983) studied the toxicity of the water soluble fraction of jet fuel to fish. Growth of fry was retarded by 1.5 mg/L of the water soluble fraction of JP-8 (jet fuel with de-icer). In a study conducted by Hedtke and Puglis (1982), the method of introducing the oil to the test chamber was an important variable driving toxicity. Emulsified oils were substantially more toxic than either floating oils or the water soluble fraction. The 96 hour LC_{50} for fathead minnows (*Pimephales promelas*) exposed to the emulsion of No. 2 jet fuel was 38.6 mg/L (dose used to calculate TRV).

Aquatic organisms other than fish may also be exposed to diesel fuel in the environment. Studies have shown that freshwater Arctic zooplankton may be more sensitive to oil pollution than any other Arctic freshwater organisms (O'Brien 1978). Geiger and Buikema (1981) estimated an LC_{20} of No. 2 Fuel oil to *Daphnia pulex* of 5.6 mg/L (dose used to calculate TRV). Wong et al. (1981) determined that concentrations of crude oil as low as 1 ppm affect the reproduction of *Daphnia pulex*.

3.3.1.3 Birds. Petroleum hydrocarbons in the environment may affect bird reproduction. External application of Number 2 fuel oil to mallard (*Anas platyrhynchos*) and eider (*Somateria mollissima*) eggs significantly increased embryo mortality (Albers 1977; Szaro and Albers 1977). Mallard eggs were treated with 1, 5, 10, 20 and 50 μ l of fuel oil. Ingestion of crude oil by mallard ducks at a concentration of five percent in the diet resulted in depressed growth (Szaro et al. 1978). Hartung (1964) demonstrated a decrease in weight gain in mallard ducks during the first 10 days after receiving 6,000 mg/kg No. 2 fuel oil (dose used to calculate TRV). However, after 34 days, there was no difference between treatment groups and the controls.

3.3.1.4 Mammals. The available literature does not present a great deal of information regarding the toxicity of diesel fuel to mammals. The toxicity of diesel fuel to mammals can be represented by the toxicity of the compound to rats. Diesel fuel is slightly toxic to rats based on an acute oral LD_{50} of 7,380 mg/kg (Beck et al. 1982) (dose used to calculate TRV). A dermal LD_{50} in rabbits was reported as >4,290 mg (Beck et al. 1982).

3.3.2 Ethylbenzene

Ethylbenzene is a COC in soil and/or sediment at the Barter Island installation. It is a VOC, and most toxicity information in the literature relates to its inhalation.

3.3.2.1 Plants. In a cell multiplication inhibition test using *Microcystis aeruginosa* (algae) the toxicity threshold of ethylbenzene was 33,000 μ g/L; The toxicity threshold of *Scenedesmus*

quadricauda (green algae) to ethylbenzene was $> 160,000 \mu\text{g/L}$ (Verschuere 1983). Galassi et al. (1988 in AQUIRE 1994) reported an EC_{50} (growth) of $4,600 \mu\text{g/L}$ for *Selenastrum capricornutum* (green algae).

3.3.2.2 Aquatic Organisms. Ethylbenzene was a COC in soil and/or sediment but not in water. As a result, no quantitative presentation of water exposure was conducted in this ERA.

3.3.2.3 Birds. There is no information in the literature regarding the toxicity of ethylbenzene to birds.

3.3.2.4 Mammals. One study regarding the toxicity of ethylbenzene administered orally to rats reported an LD_{50} $5,460 \text{ mg/kg}$ (Budavari 1989) (dose used to calculate TRV).

3.3.3 Xylenes

Xylene is a COC in soil and/or sediment at the Barter Island installation. It is a VOC, and most toxicity information in the literature relates to the inhalation of xylene.

3.3.3.1 Plants. In a study of the green algae, *Selenastrum capricornutum*, xylene decreased growth at concentrations of $72,000 \mu\text{g/L}$ (Gaur 1988 in AQUIRE).

3.3.3.2 Aquatic Organisms. Xylene is not a COC in surface waters, so the toxicity to aquatic organisms is not presented.

3.3.3.3 Birds. When mallard eggs were immersed in xylene (1 percent & 10 percent) for 30 seconds, there were no significant effects at concentrations of 10 percent on embryonic weight & length when compared to controls [Hoffman and Eastin 1981 in Hazardous Substance Data Bank (HSDB) 1994]. Japanese quail (*Coturnix japonica*) fed xylene demonstrated no sign of toxicity up to 5,000 ppm (USFWS 1986a). The LC_{50} was $>20,000 \text{ ppm}$ (USFWS 1986a). Hill and Camardese (1986) report a maximum dietary exposure level for Japanese quail of 625 mg/kg total xylenes (dose used to calculate TRV).

3.3.3.4 Mammals. Ingestion of xylene in mammals may cause prenatal mortality, growth inhibition, and malformations, primarily cleft palate. The LD_{50} for ingestion of xylene (rat) was reported as $4,300 \text{ mg/kg}$ (Clayton and Clayton 1981) (dose used to calculate TRV).

3.3.4 Trimethylbenzene

Trimethylbenzene is a COC in soils/sediments, but not in surface water. Trimethylbenzene exists in three isomeric forms, 1,2,3-TMB, 1,2,4-TMB, and 1,3,5-TMB.

3.3.4.1 Plants. Specific toxicity information for plants and trimethylbenzene is not available.

3.3.4.2 Aquatic Organisms. Trimethylbenzene was not found to be a COC in water, so it is not quantitatively evaluated in the risk assessment for aquatic organisms.

3.3.4.3 Birds. No toxicity information relating birds to trimethylbenzene was available.

3.3.4.4 Mammals. No toxicity information relating mammals to trimethylbenzene was available.

3.3.5 Naphthalene

Naphthalene was determined to be a COC in soils/sediments, but not in surface water. Naphthalene belongs to the group of chemical compounds known as PAHs. PAHs are hydrogen and carbon atoms combined to form two or more fused benzene rings (Eisler 1987). The structure of naphthalene is a two-ring, unsubstituted molecule (C₁₀H₈). Specific toxicity data for naphthalene is generally lacking, so the following discussion may include toxicity information for other PAH compounds.

3.3.5.1 Plants. Specific toxicity information for plants exposed to naphthalene or other PAHs is not available. Some general trends have been observed by researchers (EPA 1980; Lee and Grant 1981; Wang and Meresz 1982; Edwards 1983; Sims and Overcash 1983 cited in Eisler 1987). PAHs may be absorbed from soil through plant roots and can be translocated to other parts of the plant. The factors that appear to govern plant uptake include soil concentration, water solubility, and soil type.

3.3.5.2 Aquatic Organisms. Naphthalene was not found to be a COC in water so it is not quantitatively evaluated in terms of exposure of aquatic organisms.

3.3.5.3 Birds. There is limited information regarding the toxicity of PAHs to birds. In a study conducted by Patton and Dieter (1980), mallards fed 4,000 mg PAHs/kg for 7 months demonstrated increased liver weight and blood flow to the liver. The PAH mixture tested contained naphthalenes, naphthenes, and phenanthrene.

3.3.5.4 Mammals. Some PAHs are animal carcinogens. However, unsubstituted PAHs with fewer than four rings (as is naphthalene) have not been shown to induce tumorigenic activity (Eisler 1987). In addition, although unsubstituted PAHs are highly lipid soluble, they do not accumulate in mammalian tissue because of ready metabolism by animals (EPA 1980; Lee and Grant 1981). Specific studies regarding the toxicity of naphthalene to mammals are limited. 1,780 mg/kg body weight of naphthalene caused acute oral toxicity in rats (Sims and Overcash 1983 in Eisler 1987). The Hazardous Substance Data Bank (HSDB 1994) reports a NOAEL dose of 50 mg/kg naphthalene (dose used to calculate TRV) for a laboratory rat.

3.3.6 2-Methylnaphthalene

2-Methylnaphthalene was determined to be a COC in soils/sediments, but not in surface water. 2-Methylnaphthalene is a PAH, and is closely related to naphthalene. 2-Methylnaphthalene is a two-ring, substituted molecule (C₁₁H₁₀). Specific toxicity data for 2-methylnaphthalene is generally lacking so the toxicity data for naphthalene was considered.

3.3.7 Polychlorinated Biphenyls

PCBs are COCs in the soil and/or sediment at the Barter Island installation. PCBs are a group of aromatic hydrocarbons halogenated with chlorine atoms. The toxicity of PCBs is influenced primarily by the octanol/water partition coefficient (K_{ow}), and the position and number of substituted chlorine atoms. Generally, the PCBs that have the highest K_{ow} and a high number of chlorine atoms in adjacent positions are the most toxic (Eisler 1986). PCBs cause reproductive failure, tumors, liver disorders, and birth defects. They are bioaccumulative and may magnify in the food web (Eisler 1986).

3.3.7.1 Plants. In a study conducted by Klekowski (1982 in Eisler 1986), a significant increase in mutations was noted in plants (*Matteuccia struthiopteris*) growing on sediments that contained concentrations of 26 mg/kg PCBs compared to the same species of plants collected from a control area.

3.3.7.2 Aquatic Organisms. PCBs were not detected in surface waters, so they are not quantitatively evaluated for exposure of aquatic organisms.

3.3.7.3 Birds. Birds that ingest PCBs in the diet may exhibit abnormal growth patterns, reproduction, metabolism or behavior (Eisler 1986). PCBs may accumulate in birds. Concentrations of PCBs in liver of birds (mg/kg fresh weight) were correlated to the feeding niche of the bird. Birds that fed on fish had the highest liver concentrations (900 mg/kg fresh weight), species that fed on small birds and mammals had the second highest levels (50 mg/kg fresh weight), and birds whose diet is composed of invertebrates had levels of 0.65 mg/kg fresh weight. Herbivorous species demonstrated the lowest accumulation of 0.2 mg/kg fresh weight [National Academy of Science (NAS) 1979 in Eisler 1986]. The toxicity of PCBs to different species of birds was investigated by Heath et al. (1972 in Eisler 1986). The LD_{50} s for birds fed PCBs (Aroclor 1254) for 5 days, then untreated for 3 days were: Northern bobwhite (*Colinus virginianus*) 604 mg/kg diet; Mallard (*Anas platyrhynchos*) 2,699 mg/kg diet (dose used to calculate TRV); Ring-necked pheasant (*Phasianus colchicus*) 1,091 mg/kg diet; Japanese quail (*Coturnix coturnix japonica*) 2,898 mg/kg diet.

3.3.7.4 Mammals. The toxicity of PCBs to mammals varies widely between species. Mink (*Mustela vison*) are the most sensitive mammals to PCB exposure, with an LD_{50} of 6.7 mg/kg diet (as Aroclor 1254) (Ringer 1983 in Eisler 1986). The cottontail rabbit's (*Sylvilagus floridanus*) LD_{50} for Aroclor 1254 was >10 mg/kg diet (Zepp and Kirkpatrick 1976 in Eisler 1986). The raccoon (*Procyon lotor*) displayed an LD_{50} of >50 mg/kg diet when exposed to Aroclor 1254 (Montz et al. 1982 in Eisler 1986). Rats were more sensitive: LD_{50} of >75 mg/kg diet for Aroclor 1254 (Hudson et al. 1984 in Eisler 1986) (dose used to calculate TRV).

3.3.8 Aluminum

Aluminum was found to be a COC in surface water, but not in soils/sediments. Aluminum is an ubiquitous naturally occurring element. It is amphoteric, with solubility lowest at a pH of 5.5 and increasing as pH deviates from 5.5 in either direction (EPA 1988). Attempts to relate toxicity to pH have yielded diverse results. Some researchers reported a direct correlation between

increased aluminum toxicity and increased pH (Freeman and Everhardt 1971; Hunter et al. 1980 in EPA 1988), although other researchers have reported decreased toxicity (Call 1984; Boyd 1979; Kimball unpublished in EPA 1988).

3.3.8.1 Plants. According to the EPA (1988), single-celled plants are more sensitive to aluminum exposure than other plants. Concentrations of 810 $\mu\text{g/L}$ inhibited the growth of the diatom *Cyclotella meneghiniana* (Rao and Subramanian 1982 in EPA 1988). The green alga, *Selenastrum capricornutum* (Call 1984 in EPA 1988) was affected by concentrations of aluminum of 460 $\mu\text{g/L}$. The Eurasian watermilfoil demonstrated decreased root weight when exposed to aluminum concentrations of 2,500 $\mu\text{g/L}$. Concentrations of aluminum of 8,000 $\mu\text{g/L}$ in culture solutions resulted in toxic threshold concentrations in rice shoots and soybean leaves of 20 mg/kg and 30 mg/kg, respectively (Wallace and Romney 1977 in Gough et al. 1979). Most of the aluminum remained in the roots of the plant.

3.3.8.2 Aquatic Organisms. Aluminum is toxic to carp (*Cyprinus carpio*) at 4,000 $\mu\text{g/L}$ (48-hr LC_{50}) (Muramoto 1981 in EPA 1988). A 96-hr LC_{50} of 3,680 $\mu\text{g/L}$ (dose used to calculate TRV) was reported for the brook trout (*Salvelinus fontinalis*) (Decker and Menendez 1974 in EPA 1988). TRVs for the Arctic char and nine-spined stickleback are based on this value. In a chronic toxicity test with *Daphnia magna*, a reduction in survival (29 percent) was shown at to occur at concentrations of 1,020 $\mu\text{g/L}$ (Kimball unpublished in EPA 1988). The organisms that survived this test did not exhibit any adverse effect on reproduction.

3.3.8.3 Birds. In a toxicity study conducted using red-winged blackbirds as the test organism, aluminum concentrations >111 mg/kg body weight were toxic (LD_{50}) (Schafer et al. 1983). In a study conducted by Cakir et al. (1978 in NAS 1980) a no-effect concentration of 486 ppm aluminum was reported for both turkeys and chicks. Studies reported in COE (1991) show a dose of 60.8 mg/kg-bw/day to be a LD_{50} (dose used to calculate TRV).

3.3.8.4 Mammals. No toxicity study using wild mammalian species was located in the literature. Calves and sheep fed 1,200 and 1,000 ppm (respectively) of aluminum in the diet showed no adverse effects (Valdivia et al. 1978; Bailey 1977 in NAS 1980). Studies reported in ATSDR (1990a) indicate an acute systemic NOAEL for rats exposed to aluminum of 108 mg/kg-bw/day (dose used to calculate TRV).

3.3.9 Iron

Iron is an essential trace element required by both plants and animals. It is a COC in surface water at the Barter Island installation.

3.3.9.1 Plants. In a study conducted by Foy et al. (1978 in EPA 1985), concentrations of 100 to 500 ppm soluble iron in soil were toxic to rice.

3.3.9.2 Aquatic Organisms. Iron may be a threat in aquatic environments as precipitates can destroy habitat, coat gills, and inhibit oxygen uptake. The EPA uses 1,000 $\mu\text{g/L}$ as the chronic ambient water quality criterion protective of aquatic life (dose used to calculate TRV). In a study conducted by Warnick and Bell (1969 in EPA 1976) mayflies, stoneflies, and caddisflies

were affected by iron concentrations of 320 µg/L (96-hr LC₅₀). Doudoroff and Katz (1953 in EPA 1976) found iron concentrations of 1,000 to 2,000 µg/L toxic to *Esox lucius* (northern pike) and trout (unknown species).

3.3.9.3 Birds. There are few studies available that address the toxicity of iron to species of wild birds. There were no adverse effects produced in turkeys at concentrations of 440 ppm (Woerpel and Balloun 1964 in NAS 1980). NAS (1980) recommends that the maximum tolerable level of dietary iron of 1,000 ppm be used for poultry. The 1,000 ppm dose converts to 125.0 mg/kg for a maximum tolerable dietary level for a chicken (dose used to calculate TRV).

3.3.9.4 Mammals. At high concentrations, iron is toxic to livestock and interferes with phosphorus metabolism (NAS 1974 in EPA 1976). Cattle fed 477 µg/g iron demonstrated a slight decrease in weight gain; concentrations of 1,677 µg/g of iron produced a significant decline in growth rate (EPA 1985). COE (1991) reports an acute hepatic LOAEL dose of iron for the rat to be 750 mg/kg (dose used to calculate TRV).

3.3.10 Lead

Lead was found to be a COC in soils/sediments, but not in surface water.

3.3.10.1 Plants. Lead inhibits plant growth, reduces photosynthesis and reduces mitosis and water absorption (Eisler 1988). Concentrations of 500 mg/kg in soils were found to result in reduced pollen germination in several weed species, but the same study found that 46 mg/kg lead concentrations in soil did not have adverse effects on pollen germination (COE 1991).

3.3.10.2 Aquatic Organisms. Lead was not found to be a COC in surface water and, as a result, aquatic organisms are not evaluated quantitatively for lead exposure in the ERA.

3.3.10.3 Birds. The bulk of the toxicity information in the literature regarding avian exposure to lead concerns waterfowl that have ingested spent lead shot and died. These results are reported as body burdens of lead. There is, however, limited dose-response information available for some species. Mautino and Bell (1987) reported neurological effects in mallard ducks that had ingested and absorbed lead shot for a total intake of 423.8 mg/kg body weight. In young American kestrels (1 day old) ingesting 125 and 625 mg/kg body weight, growth and hematocrit values were significantly depressed (Hoffman et al. 1985). Based on a review of several studies, 12.0 mg/kg is the highest chronic NOAEL lead dose for many species of birds (dose used to calculate TRV).

3.3.10.4 Mammals. Lead may affect the survival, growth, development and metabolism of animal species. Rats are affected by 5 to 108 mg/kg body weight (acute oral); dogs by 0.32 mg/kg body weight daily (chronic oral); and horses by chronic dietary concentrations of 1.7 mg/kg (Eisler 1988). Eisler (1988) reports an oral dose of 12.0 mg/kg-bw/day to be a LD₅₀ for laboratory rats (dose used to calculate TRV).

3.3.11 Manganese

Manganese was determined to be a COC in surface water, but not in soils/sediments. Manganese is considered to be an essential nutrient for animals (ATSDR 1990a), and it is important for growth and reproduction. The toxicity of manganese can be affected by pH and water hardness.

3.3.11.1 Plants. In a four-day study conducted using duckweed (*Lemna minor*), an EC₅₀ (reduction in growth) was reported of 31,000 µg/L (Wang 1986 in AQUIRE 1990). Lewis et al. (1979) studied the species composition of freshwater phytoplankton populations when exposed to manganese. Population composition was altered at 0.1 mg/L manganese. Soil concentrations of 1,500 to 3,000 mg/kg reported as phytotoxic to all plant species (Kabata-Pendias and Pendias 1984).

3.3.11.2 Aquatic Organisms. In a study conducted by Doudoroff and Katz (1953), brook trout were killed within 24 hours when exposed to concentrations of manganese of 6,250 µg/L. Rainbow trout have a reported LC₅₀ of 2,910 µg/L (Pickering et al. 1983) (dose used to calculate TRV). *Daphnia* spp. have a reported 16 percent reproductive impairment in water with concentrations of 4,100 µg/L (Biesinger and Christensen 1972 in Lewis et al. 1979) (dose used to calculate TRV).

3.3.11.3 Birds. Vohra and Kratzer (1968 in NAS 1980) exposed young turkeys to dietary manganese for 21 days. A NOEL of 4,080 ppm was derived. The maximum tolerable levels of manganese that are recommended by the NAS are 2,000 ppm (250 mg/kg body weight) for poultry. This dose is used as a NOAEL for the derivation of the TRV for birds.

3.3.11.4 Mammals. When fed 9,000 ppm manganese, sheep demonstrated reduced feed intake (Puls 1988). NAS recommends maximum tolerable levels of 1,000 ppm for cattle (15 mg/kg body weight) and sheep (40 mg/kg body weight). A NOAEL of 930 mg/kg-bw/day is reported for rats in ATSDR (1990b). The TRV for mammals is based on this dose.

3.3.12 Zinc

Zinc was determined to be a COC in soils/sediments, but not in surface water. Zinc is considered to be an essential nutrient for animals Eisler (1993), and is necessary for plant growth. Deficiencies of zinc in the diet may retard growth in animals (Eisler 1993).

3.3.12.1 Plants. According to information presented in Eisler (1993), plants that are sensitive to zinc concentrations die when soil levels are in excess of 100 mg/kg. Plant tissue content is also toxic in excess of 178 mg/kg (Eisler 1988). The amount of zinc absorbed from soil in plants is dependent upon soil specific characteristics. COE (1991) report that several species of plants find average concentrations of zinc of 270 mg/kg in soil to be phytotoxic.

3.3.12.2 Aquatic Organisms. Zinc was determined not to be a COC in water and as a result, aquatic organisms are not evaluated quantitatively.

3.3.12.3 Birds. When ducks were fed 2,500 to 3,000 mg/kg ration of zinc, or alternately, force-fed zinc at 742 mg zinc/kg body weight, survival was reduced (Eisler 1993). Chickens were more resistant to zinc exposure; 8,000 mg zinc/kg ration was lethal to chicks (Eisler 1993). Elevated levels of zinc (20 g zinc/kg ration) are given to poultry to induce molting and subsequently reduce egg deposition (Eisler 1993). An acute NOAEL for chickens of 125 mg/kg-bw/day was reported by Morrison Knudsen/ICF (1993). The TRV for birds is based on this dosage.

3.3.12.4 Mammals. According to Eisler, zinc is relatively non-toxic to mammals (as would be expected for an essential trace element). There is a large range in concentrations between normal dietary intakes and those concentrations expected to cause harm. Adult male rats, when fed zinc at levels of 500 mg/kg diet, were adversely affected: spermatogenesis was arrested, and testes enlarged (Eisler 1993). Zinc concentrations of 6,820 mg zinc/kg ration suppressed rat growth and produced changes in the pancreas (Eisler 1993). A chronic NOAEL for laboratory rats of 50 mg/kg-bw/day is reported by MK/ICF (1993). The TRV is derived from this toxicity value.

3.3.13 Characterization of Effects

In this section toxicity information is presented for representative ecological receptors that will be evaluated in the Risk Characterization section (3.4) of this report. Potential impacts to aquatic receptors are evaluated by comparing exposure concentrations to TRVs. Potential impacts to terrestrial wildlife are evaluated for selected representative species based on comparisons of estimated exposures to TRVs. TRVs for the representative aquatic species are presented in Tables 3-20 and 3-21. Exposure to the selected representative species will be primarily through diet; which may include plants, fish and aquatic invertebrates, soils and surface water. TRVs are derived for COC that are elevated in surface water and soil/sediments. TRVs for the representative bird species are presented in Tables 3-22 and 3-23, for the threatened and endangered species in Tables 3-24 and 3-25, and for the representative mammal species in Tables 3-26 and 3-27.

3.3.13.1 Toxicity Reference Values. TRVs are derived by selecting toxicity values from the literature and then extrapolating to the species of concern. Uncertainty factors (UF) and body scaling factors are used in the extrapolation process as described below. Tables 3-20 to 3-27 present the information used to derive the TRVs.

(1) The first step is to select an appropriate toxicity value from the scientific literature for each combination of chemical and representative or protected species. Test species most similar to the species of concern are preferred. A secondary emphasis is given to tests conducted over a significant portion of the animal's natural lifespan (e.g., chronic tests), when available. It should be kept in mind that the goal of the risk assessment is to evaluate the ecological significance of any potential contamination and the use of TRVs is in keeping with this goal.

(2) The toxicity value is then modified through application of uncertainty factors associated with the quality of toxicity data to derive a NOAEL. If a chronic NOAEL or No

TABLE 3-20. TOXICITY REFERENCE VALUES FOR REPRESENTATIVE SPECIES OF AQUATIC ORGANISMS AT THE BARTER ISLAND INSTALLATION (METALS)

CHEMICAL OF CONCERN	REPRESENTATIVE SPECIES	STUDY TYPE	DOSE $\mu\text{g/L}$	TEST SPECIES	NOAEL UF	INTERSPECIES UF	PROTECTED SPECIES UF	TRV mg/L	REFERENCE
Aluminum	Arctic char	LC ₅₀	3680	brook trout	20	2	1	92.0	COE 1991
Aluminum	nine-spined stickleback	LD ₅₀	3680	brook trout	20	2	1	92.0	COE 1991
Aluminum	<i>Daphnia</i> spp.	chronic reproductive impairment LOAEL	1020	<i>D. magna</i>	10	1	1	102.0	COE 1991
Manganese	Arctic char	LC ₅₀	2910	rainbow trout	20	2	1	73.0	COE 1991
Manganese	nine-spined stickleback	LC ₅₀	2910	rainbow trout	20	2	1	73.0	COE 1991
Manganese	<i>Daphnia</i> spp.	reproductive impairment LOAEL	4100	<i>Daphnia</i> spp.	20	1	1	205.0	COE 1991
Iron	Arctic char	EPA chronic water quality criteria	1000	all aquatic life	1	1	1	1,000.0	COE 1991
Iron	nine-spined stickleback	EPA chronic water quality criteria	1000	all aquatic life	1	1	1	1,000.0	COE 1991
Iron	<i>Daphnia</i> spp.	EPA chronic water quality criteria	1000	all aquatic life	1	1	1	1,000.0	COE 1991

TABLE 3-21. TOXICITY REFERENCE VALUES FOR REPRESENTATIVE SPECIES OF AQUATIC ORGANISMS AT THE BARTER ISLAND INSTALLATION (ORGANICS)

CHEMICAL OF CONCERN	REPRESENTATIVE SPECIES	STUDY TYPE	DOSE $\mu\text{g/L}$	TEST SPECIES	NOAEL UF	INTERSPECIES UF	PROTECTED SPECIES UF	TRV $\mu\text{g/L}$	REFERENCE
DRPH	Arctic char	LC ₅₀	38,600	fathead minnow	20	2	1	965.0	Hedtke and Puglis 1982
DRPH	nine-spined stickleback	LC ₅₀	38,600	fathead minnow	20	2	1	965.0	Hedtke and Puglis 1982
DRPH	<i>Daphnia</i> spp.	LOAEL	5600	<i>D. pulex</i>	20	1	1	280.0	Hedtke and Puglis 1982

TABLE 3-22. TOXICITY REFERENCE VALUES FOR REPRESENTATIVE SPECIES OF BIRDS AT THE BARTER ISLAND INSTALLATION (METALS)

CHEMICAL OF CONCERN	REPRESENTATIVE SPECIES	STUDY TYPE	DOSE mg/kg-BW/DAY	TEST SPECIES	NOAEL UF	SCALING FACTOR	INTERSPECIES UF	PROTECTED SPECIES UF	TRV mg/kg-BW/DAY	REFERENCE
Aluminum	Lapland longspur	LD ₅₀	60.8	chicken	20	0.26	2	2	2.9	COE 1991
Aluminum	brant	LD ₅₀	60.8	chicken	20	0.93	2	2	0.8	COE 1991
Aluminum	glaucous gull	LD ₅₀	60.8	chicken	20	0.97	2	2	0.8	COE 1991
Aluminum	pectoral sandpiper	LD ₅₀	60.8	chicken	20	0.37	2	2	2.1	COE 1991
Iron	Lapland longspur	maximum tolerable dietary level	125.0	chicken	10	0.26	2	2	12.0	NAS 1980
Iron	brant	maximum tolerable dietary level	125.0	chicken	10	0.93	2	2	3.4	NAS 1980
Iron	glaucous gull	maximum tolerable dietary level	125.0	chicken	10	0.97	2	2	3.2	NAS 1980
Iron	pectoral sandpiper	maximum tolerable dietary level	125.0	chicken	10	0.37	2	2	8.4	NAS 1980
Lead	Lapland longspur	NOAEL	12.0	all birds	1	na	na	2	6.0	COE 1991
Lead	brant	NOAEL	12.0	all birds	1	na	na	2	6.0	COE 1991
Lead	glaucous gull	NOAEL	12.0	all birds	1	na	na	2	6.0	COE 1991
Lead	pectoral sandpiper	NOAEL	12.0	all birds	1	na	na	2	6.0	COE 1991

TABLE 3-22. TOXICITY REFERENCE VALUES FOR REPRESENTATIVE SPECIES OF BIRDS AT THE BARTER ISLAND INSTALLATION (METALS) (CONTINUED)

CHEMICAL OF CONCERN	REPRESENTATIVE SPECIES	STUDY TYPE	DOSE mg/kg-BW/DAY	TEST SPECIES	NOAEL UF	SCALING FACTOR	INTERSPECIES UF	PROTECTED SPECIES UF	TRV mg/kg-BW/DAY	REFERENCE
Manganese	Lapland longspur	LOAEL	250	chicken	10	0.26	2	2	240.30	NAS 1980
Manganese	brant	LOAEL	250	chicken	10	0.93	2	2	67.0	NAS 1980
Manganese	glaucous gull	LOAEL	250	chicken	10	0.97	2	2	64.0	NAS 1980
Manganese	pectoral sandpiper	LOAEL	250	chicken	10	0.37	2	2	169.0	NAS 1980
Zinc	Lapland longspur	NOAEL	125	chicken	10	0.60	2	2	5.0	MK/ICF 1993
Zinc	brant	NOAEL	125	chicken	10	2.16	2	2	1.0	MK/ICF 1993
Zinc	glaucous gull	NOAEL	125	chicken	10	2.23	2	2	1.0	MK/ICF 1993
Zinc	pectoral sandpiper	NOAEL	125	chicken	10	0.85	2	2	4.0	MK/ICF 1993

TABLE 3-23. TOXICITY REFERENCE VALUES FOR REPRESENTATIVE SPECIES OF BIRDS AT THE BARTER ISLAND INSTALLATION (ORGANICS)

CHEMICAL OF CONCERN	REPRESENTATIVE SPECIES	STUDY TYPE	DOSE mg/kg-BW/DAY	TEST SPECIES	NOAEL UF	SCALING FACTOR	INTERSPECIES UF	PROTECTED SPECIES UF	TRV mg/kg-BW/DAY	REFERENCE
DRPH	Lapland longspur	decreased weight gain LOAEL	6000	mallard	10	0.29	2	2	517.0	Hartung 1964
DRPH	brant	decreased weight gain LOAEL	6000	mallard	10	1.07	2	2	140.0	Hartung 1964
DRPH	glaucous gull	decreased weight gain LOAEL	6000	mallard	10	1.10	2	2	136.0	Hartung 1964
DRPH	pectoral sandpiper	decreased weight gain LOAEL	6000	mallard	10	0.42	2	2	357.0	Hartung 1964
Ethylbenzene	Lapland longspur	NA								
Ethylbenzene	brant	NA								
Ethylbenzene	glaucous gull	NA								
Ethylbenzene	pectoral sandpiper	NA								
Naphthalene	Lapland longspur	NA								
Naphthalene	brant	NA								
Naphthalene	glaucous gull	NA								
Naphthalene	pectoral sandpiper	NA								
2-Methylnaphthalene	Lapland longspur	NA								
2-Methylnaphthalene	brant	NA								
2-Methylnaphthalene	glaucous gull	NA								
2-Methylnaphthalene	pectoral sandpiper	NA								
PCBs	Lapland longspur	LD ₅₀	2,699	mallard	20	0.29	2	2	116.0	Heath et al. 1972

TABLE 3-23. TOXICITY REFERENCE VALUES FOR REPRESENTATIVE SPECIES OF BIRDS AT THE BARTER ISLAND INSTALLATION (ORGANICS) (CONTINUED)

CHEMICAL OF CONCERN	REPRESENTATIVE SPECIES	STUDY TYPE	DOSE mg/kg-BW/DAY	TEST SPECIES	NOAEL UF	SCALING FACTOR	INTERSPECIES UF	PROTECTED SPECIES UF	TRV mg/kg-BW/DAY	REFERENCE
PCBS	brant	LD ₅₀	2,699	mallard	20	1.07	2	2	32.0	Heath et al. 1972
PCBs	glaucous gull	LD ₅₀	2,699	mallard	20	1.10	2	2	31.0	Heath et al. 1972
PCBs	pectoral sandpiper	LD ₅₀	2,699	mallard	20	0.42	2	2	80.0	Heath et al. 1972
Trimethylbenzene	Lapland longspur	NA								
Trimethylbenzene	brant	NA								
Trimethylbenzene	glaucous gull	NA								
Trimethylbenzene	pectoral sandpiper	NA								
Xylenes (total)	Lapland longspur	maximum dietary exposure level	625	Japanese quail	10	0.60	2	2	26.0	Hill and Camardese 1986
Xylenes (total)	brant	maximum dietary exposure level	625	Japanese quail	10	2.16	2	2	7.2	Hill and Camardese 1986
Xylenes (total)	glaucous gull	maximum dietary exposure level	625	Japanese quail	10	2.23	2	2	7.0	Hill and Camardese 1986
Xylenes (total)	pectoral sandpiper	maximum dietary exposure level	625	Japanese quail	10	0.85	2	2	18.4	Hill and Camardese 1986

TABLE 3-24. TOXICITY REFERENCE VALUES FOR THREATENED AND ENDANGERED SPECIES AT THE BARTER ISLAND INSTALLATION (METALS)

CHEMICAL OF CONCERN	PROTECTED SPECIES	STUDY TYPE	DOSE mg/kg-BW/DAY	TEST SPECIES	NOAEL UF	SCALING FACTOR	INTERSPECIES UF	T & E UF	TRV mg/kg-BW/DAY	REFERENCE
Aluminum	peregrine falcon	LD ₅₀	60.8	chicken	20	0.80	2	2	1.0	COE 1991
Aluminum	spectacled eider	LD ₅₀	60.8	chicken	20	0.95	2	2	0.8	COE 1991
Aluminum	Steller's eider	LD ₅₀	60.8	chicken	20	0.80	2	2	1.0	COE 1991
Iron	peregrine falcon	maximum tolerable dietary level	125	chicken	10	0.80	2	2	3.9	NAS 1980
Iron	spectacled eider	maximum tolerable dietary level	125	chicken	10	0.95	2	2	3.3	NAS 1980
Iron	Steller's eider	maximum tolerable dietary level	125	chicken	10	0.80	2	2	3.9	NAS 1980
Lead	peregrine falcon	chronic NOAEL	12.0	all birds	1	NA	NA	2	6.0	COE 1991
Lead	spectacled eider	chronic NOAEL	12.0	all birds	1	NA	NA	2	6.0	COE 1991
Lead	Steller's eider	chronic NOAEL	12.0	all birds	1	NA	NA	2	6.0	COE 1991
Manganese	peregrine falcon	maximum tolerable dietary level	250	chicken	10	0.80	2	2	7.8	NAS 1980
Manganese	spectacled eider	maximum tolerable dietary level	250	chicken	10	0.95	2	2	6.6	NAS 1980
Manganese	Steller's eider	maximum tolerable dietary level	250	chicken	10	0.80	2	2	7.8	NAS 1980
Zinc	peregrine falcon	NOAEL	125	chicken	10	0.80	2	2	4.0	Roberson and Schaible 1960
Zinc	spectacled eider	NOAEL	125	chicken	10	0.95	2	2	3.0	Roberson and Schaible 1960
Zinc	Steller's eider	NOAEL	125	chicken	10	0.80	2	2	4.0	Roberson and Schaible 1960

TABLE 3-25. TOXICITY REFERENCE VALUES FOR THREATENED AND ENDANGERED SPECIES AT THE BARTER ISLAND INSTALLATION (ORGANICS)

CHEMICAL OF CONCERN	PROTECTED SPECIES	STUDY TYPE	DOSE mg/kg-BW/DAY	TEST SPECIES	NOAEL UF	SCALING FACTOR	INTERSPECIES UF	T & E UF	TRV mg/kg-BW/DAY	REFERENCE
DRPH	peregrine falcon	decreased weight gain LOAEL	6000	mallard	10	0.91	2	2	165.0	Hartung 1964
DRPH	spectacled eider	decreased weight gain LOAEL	6000	mallard	10	1.08	2	2	139.0	Hartung 1964
DRPH	Steller's eider	decreased weight gain LOAEL	6000	mallard	10	0.91	2	2	165.0	Hartung 1964
Ethylbenzene	peregrine falcon	NA								
Ethylbenzene	spectacled eider	NA								
Ethylbenzene	Steller's eider	NA								
Naphthalene	peregrine falcon	NA								
Naphthalene	spectacled eider	NA								
Naphthalene	Steller's eider	NA								
2-Methylnaphthalene	peregrine falcon	NA								
2-Methylnaphthalene	spectacled eider	NA								
2-Methylnaphthalene	Steller's eider	NA								
PCBs	peregrine falcon	LD ₅₀	2,699	mallard	20	0.91	2	2	37.0	Heath et al. 1972
PCBs	spectacled eider	LD ₅₀	2,699	mallard	20	1.08	2	2	31.0	Heath et al. 1972
PCBs	Steller's eider	LD ₅₀	2,699	mallard	20	0.91	2	2	37.0	Heath et al. 1972

TABLE 3-25. TOXICITY REFERENCE VALUES FOR THREATENED AND ENDANGERED SPECIES AT THE BARTER ISLAND INSTALLATION (ORGANICS) (CONTINUED)

CHEMICAL OF CONCERN	PROTECTED SPECIES	STUDY TYPE	DOSE mg/kg-BW/DAY	TEST SPECIES	NOAEL UF	SCALING FACTOR	INTERSPECIES UF	T & E UF	TRV mg/kg-BW/DAY	REFERENCE
Trimethylbenzene	peregrine falcon	NA								
Trimethylbenzene	spectacled eider	NA								
Trimethylbenzene	Steller's eider	NA								
Xylenes (total)	peregrine falcon	maximum dietary exposure level	625	Japanese quail	10	1.85	2	2	8.4	Hill and Camardese 1986
Xylenes (total)	spectacled eider	maximum dietary exposure level	625	Japanese quail	10	2.19	2	2	7.1	Hill and Camardese 1986
Xylenes (total)	Steller's eider	maximum dietary exposure level	625	Japanese quail	10	1.84	2	2	8.5	Hill and Camardese 1986

TABLE 3-26. TOXICITY REFERENCE VALUES FOR REPRESENTATIVE SPECIES OF MAMMALS AT THE BARTER ISLAND INSTALLATION (METALS)

CHEMICAL OF CONCERN	REPRESENTATIVE SPECIES	STUDY TYPE	DOSE mg/kg-BW/DAY	TEST SPECIES	NOAEL UF	SCALING FACTOR	INTERSPECIES UF	PROTECTED SPECIES UF	TRV mg/kg-BW/DAY	REFERENCE
Aluminum	brown lemming	acute systemic NOAEL	108	rat	10	0.65	2	1	8.0	ATSDR 1990a
Aluminum	Arctic fox	acute systemic NOAEL	108	rat	10	2.91	2	1	2.0	ATSDR 1990a
Iron	brown lemming	acute hepatic LOAEL	750	rat	20	0.65	2	1	29.0	COE 1991
Iron	Arctic fox	acute hepatic LOAEL	750	rat	20	2.91	2	1	6.0	COE 1991
Lead	brown lemming	LD ₅₀	12	rat	20	0.65	2	1	0.5	Eisler 1993
Lead	Arctic fox	LD ₅₀	12	rat	20	2.91	2	1	0.1	Eisler 1993
Manganese	brown lemming	chronic systemic NOAEL	930	rat	1	0.65	2	1	715.0	ATSDR 1990b
Manganese	Arctic fox	chronic systemic NOAEL	930	rat	1	2.91	2	1	160.0	ATSDR 1990b
Zinc	brown lemming	chronic reproductive NOAEL	50	rat	1	0.65	2	1	38.0	Sutter and Nelson 1937 in ATSDR 1989
Zinc	Arctic fox	chronic reproductive NOAEL	50	rat	1	2.91	2	1	9.0	Sutter and Nelson 1937 in ATSDR 1989

TABLE 3-27. TOXICITY REFERENCE VALUES FOR REPRESENTATIVE SPECIES OF MAMMALS AT THE BARTER ISLAND INSTALLATION (ORGANICS)

CHEMICAL OF CONCERN	REPRESENTATIVE SPECIES	STUDY TYPE	DOSE mg/kg-BW/DAY	TEST SPECIES	NOAEL UF	SCALING FACTOR	INTERSPECIES UF	PROTECTED SPECIES UF	TRV mg/kg-BW/DAY	REFERENCE
DRPH	brown lemming	LD ₅₀	7380	rat	20	0.65	2	1	284.0	Beck et al. 1982
DRPH	Arctic fox	LD ₅₀	7380	rat	20	2.91	2	1	63.0	Beck et al. 1982
Ethylbenzene	brown lemming	LD ₅₀	5460	rat	20	0.65	2	1	210.00	Clayton and Clayton 1981
Ethylbenzene	Arctic fox	LD ₅₀	5460	rat	20	2.91	2	1	47.0	Clayton and Clayton 1981
Naphthalene	brown lemming	NOAEL	50	rat	1	0.65	2	1	38.0	HSDB 1994
Naphthalene	Arctic fox	NOAEL	50	rat	1	2.91	2	1	9.0	HSDB 1994
2-Methyl-naphthalene	brown lemming	NA								
2-Methyl-naphthalene	Arctic fox	NA								
PCBs	brown lemming	LD ₅₀	>75	rat	20	0.65	2	1	3.0	Hudson et al. 1984
PCBs	Arctic fox	LD ₅₀	>75	rat	20	2.91	2	1	0.6	Hudson et al. 1984
Trimethylbenzene	brown lemming	NA								
Trimethylbenzene	Arctic fox	NA								
Xylenes (total)	brown lemming	LD ₅₀	4,300	rat	20	0.65	2	1	165.0	Clayton and Clayton 1981
Xylenes (total)	Arctic fox	LD ₅₀	4,300	rat	20	2.91	2	1	37.0	Clayton and Clayton 1981

Observed Effect Level (NOEL) is available, then the toxicity value is used with an uncertainty factor of one (i.e., no adjustment) because these values have the lowest uncertainty. If chronic data are unavailable, acute or subchronic toxicity data are modified by uncertainty factors to extrapolate to chronic effects. Based on Harding Lawson Associates (1992), the following strategy was derived for uncertainty factors for extrapolating study results to chronic NOELs: 10 for chronic LOEL values, 10 for subchronic NOEL values, and 20 for subchronic LOEL values. LC₅₀ and LD₅₀ values are extrapolated to chronic NOELs by a factor of 20.

(3) A body-scaling factor is used to extrapolate the estimated NOEL for the test species to a NOEL for the species of concern. Klaassen et al. (1986) have indicated that dose expressed on a per unit surface area basis may be more appropriate than dose per unit body weight. The underlying assumption is that a toxicant acts on a physiologic surface and that the toxic effect increases as the ratio of chemical to surface area increases. The scaling factor (SIF) accounts for differences in the mass to surface area ratios between species. In this assessment the scaling factor is calculated using the following equation (Mantel and Schneiderman 1975): $SIF = (\text{weight of representative species} / \text{weight of test species})^{1/3}$.

(4) An uncertainty factor of 2 is used to account for interspecies variation in sensitivity. This value is based on the methodology used in Harding Lawson Associates (1992).

The relationships between the values given in the tables can be summarized as:

DOSE, divided by NOAEL UF, divided by SCALING FACTOR (if applicable), divided by INTERSPECIES UF, divided by PROTECTED SPECIES UF equals TRV.

(5) An uncertainty factor of 2 was used to account for additional sensitivity of state and/or federally protected species. This value is based on Harding Lawson Associates (1992). Migratory birds are Federally protected and include all the representative avian and protected species selected for this assessment.

The TRVs for birds and mammals in this assessment were derived using the five steps described above. The process is similar for the representative aquatic organisms, with the exception of the scaling factor. Scaling factors are not necessary for aquatic species because they are in direct contact with the medium (i.e., water) that is the contaminant pathway.

3.4 RISK CHARACTERIZATION FOR ECOLOGICAL RECEPTORS

In this section, potential risks to ecological receptors are evaluated. Potential risks to plants are evaluated based on the contaminant concentrations in the soil/sediment and information from the literature. Potential risks to aquatic organisms, birds, and mammals are estimated by comparing estimated exposures to TRVs. This method is known as the Quotient Method (EPA 1992a). The Quotient Method divides the estimated exposure concentration by the associated TRV to derive the TRV ratio. If the TRV ratio is less than 1.0, then adverse effects are not

expected. Conversely, if the TRV ratio is equal to or greater than 1.0 a potential for adverse effects exists. The confidence level of the risk estimate is increased as the magnitude of the ratio departs from 1.0. For example, there is greater confidence in a risk estimate where the TRV ratio is 0.1 or 10, than in a TRV ratio such as 0.9 or 1.1.

The characterization of risk focuses on the Assessment and Measurement Endpoints. These endpoints are selected and discussed in keeping with the Framework for Ecological Risk Assessment guidance (EPA 1992a). The Assessment Endpoints for the Barter Island ERA are changes in: the populations of the plant representative species (*Carex* spp., *Salix* spp. and *Eriophorum* spp.); the populations of aquatic representative species (*Daphnia* spp., nine-spined stickleback and arctic char); the populations of avian representative species (Lapland longspur, brant, glaucous gull and pectoral sandpiper); and, the populations of mammalian representative species (brown lemming and arctic fox). In addition, individual (as opposed to populations for other species) peregrine falcons, spectacled eiders and Steller's eiders are Assessment Endpoints because of their federal listing as threatened species (Steller's eider is proposed for listing as threatened). The Measurement Endpoints are the species themselves, used to characterize risk through the comparison of the estimated exposures to COCs (presented in the Ecological Exposure Assessment, Sections 3.2.5 to 3.2.7) with the TRVs.

Potential ecological risks are presented in the following sections: Section 3.4.1 addresses representative species of plants; Section 3.4.2 considers aquatic organisms; Section 3.4.3 addresses representative species of birds (and endangered and threatened species); and Section 3.4.4 discusses representative species of mammals. The potential ecological risks are summarized in Tables 3-28 through 3-30.

3.4.1 Potential Risks to Representative Species of Plants

In determining the risks to plants at the Barter Island installation, a qualitative comparison was made of soil/sediment and surface water chemical concentrations and plant toxicity information in the literature. Table 3-28 summarizes these comparisons. There is a great deal of uncertainty in this phase of the assessment because of the differences in degree of uptake between plant species (Walker et al. 1978). However, the concentrations of contaminants onsite can be compared on the level of orders of magnitude. This comparison allows broad trends to be observed in order to determine whether a potential risk may exist.

Information is generally lacking concerning the toxicity of the COC at Barter Island and how they relate to the representative species of plants. As a result, when comparisons of TRVs for site-specific species and chemicals are not possible, comparisons of related chemicals with other plant species are made.

As seen in Table 3-28, the concentrations of metals, except zinc, found in the soil/sediment and water at Barter Island are at least one order of magnitude lower than reported toxicity values for various plants. Concentrations of zinc at the site are substantially lower than the reported toxicity values. Zinc uptake by plants was shown to be significantly lower at 16° than at 32° C (Giordano and Mortvedt 1980). These comparisons are not definitive in judging the toxicity of metals to the specific representative plant species, but the risk to *Carex* spp., *Salix* spp., *Eriophorum* spp., and

TABLE 3-28. COMPARISON OF CONCENTRATIONS OF POTENTIAL CONTAMINANTS TO TOXICITY INFORMATION FOR PLANTS AT THE BARTER ISLAND INSTALLATION

CHEMICAL (COC media)	PLANT	EXPOSURE LEVEL	EFFECT ON PLANT	BARTER ISLAND EXPOSURE	REFERENCE
ALUMINUM (COC in water)	Eurasian milfoil	2500 µg/L in water	decreased root weights	318 µg/L	COE 1991
	rice/soybeans	8000 µg/L in water	toxic to shoot	313 µg/L	COE 1991
IRON (COC in water)	rice	100,000- 500,000 µg/L; >500,000 µg/L	toxic; highly toxic	8,580 µg/L	COE 1991
MANGANESE (COC in water)	duckweed	31000 µg/L in water	EC ₅₀	574 µg/L	COE 1991
LEAD (COC in soil)	weed spp.	500 mg/kg in soil	reduced pollen germination	26.5 mg/kg	COE 1991
	weed spp.	46 mg/kg	normal germination	26.5 mg/kg	COE 1991
ZINC (COC in soil)	several spp.	270 mg/kg (avg) in soil	phytotoxic	77.3 mg/kg	COE 1991
VOCs	green algae	4,600 µg/L for ethylbenzene 2,290,000 µg/L for methylene chloride, in water	EC ₅₀	821 µg/L as DRPH	COE 1991

TABLE 3-29. RISK CHARACTERIZATION OF REPRESENTATIVE SPECIES OF AQUATIC ORGANISMS AT THE BARTER ISLAND INSTALLATION

SPECIES	ESTIMATED EXPOSURE CONCENTRATION μg/liter	TRV μg/liter	RATIO OF CONCENTRATION TO TRV
ALUMINUM			
Arctic char	318	92.0	3.1
nine-spined stickleback	318	92.0	3.1
<i>Daphnia</i> spp.	318	102.0	3.4
IRON			
Arctic char	8,580	1000.0	8.6
nine-spined stickleback	8,580	1000.0	8.6
<i>Daphnia</i> spp.	8,580	1000.0	8.6
MANGANESE			
Arctic char	574	73.0	7.9
nine-spined stickleback	574	73.0	7.9
<i>Daphnia</i> spp.	574	205.0	2.8
DRPH			
Arctic char	719	965.0	0.7
nine-spined stickleback	719	965.0	0.7
<i>Daphnia</i> spp.	719	560.0	1

TABLE 3-30. RISK CHARACTERIZATION OF REPRESENTATIVE BIRD, MAMMAL, AND PROTECTED SPECIES AT THE BARTER ISLAND INSTALLATION

SPECIES	ESTIMATED DAILY DOSE mg/kg-BW/DAY	TRV mg/kg-BW/DAY	TRV RATIO
ALUMINUM			
Lapland longspur	3×10^{-2}	2.9	1×10^{-2}
brant	3×10^{-3}	0.8	4×10^{-3}
glaucous gull	5×10^{-3}	0.8	6×10^{-3}
pectoral sandpiper	3×10^{-2}	2.1	1×10^{-4}
peregrine falcon	2×10^{-3}	1.0	2×10^{-3}
spectacled eider	2×10^{-4}	0.8	2×10^{-4}
Steller's eider	2×10^{-4}	1.0	2×10^{-4}
brown lemming	2×10^{-2}	8.0	3×10^{-3}
Arctic fox	3×10^{-4}	2.0	2×10^{-4}
IRON			
Lapland longspur	8×10^{-1}	12.0	7×10^{-2}
brant	8×10^{-2}	3.4	2×10^{-2}
glaucous gull	1×10^{-1}	3.2	3×10^{-2}
pectoral sandpiper	8×10^{-1}	8.4	1×10^{-3}
peregrine falcon	5×10^{-2}	3.9	1×10^{-2}
spectacled eider	4×10^{-3}	3.3	1×10^{-3}
Steller's eider	5×10^{-3}	3.9	1×10^{-3}
brown lemming	5×10^{-1}	29.0	2×10^{-2}
Arctic fox	7×10^{-3}	6.0	1×10^{-3}
LEAD			
Lapland longspur	1×10^{-1}	6.0	2×10^{-2}
brant	3×10^{-2}	6.0	5×10^{-3}
glaucous gull	3×10^{-2}	6.0	5×10^{-3}
pectoral sandpiper	4×10^{-1}	6.0	7×10^{-2}
peregrine falcon	3×10^{-3}	6.0	5×10^{-4}
spectacled eider	1×10^{-3}	6.0	2×10^{-4}
Steller's eider	2×10^{-2}	6.0	3×10^{-4}

NC Not calculated because toxicity values were not available.

TABLE 3-30. RISK CHARACTERIZATION OF REPRESENTATIVE BIRD, MAMMAL, AND PROTECTED SPECIES AT THE BARTER ISLAND INSTALLATION (CONTINUED)

SPECIES	ESTIMATED DAILY DOSE mg/kg-BW/DAY	TRV mg/kg-BW/DAY	TRV RATIO
brown lemming	8×10^{-1}	0.5	1.6
Arctic fox	4×10^{-6}	0.1	4×10^{-3}
MANGANESE			
Lapland longspur	5×10^{-2}	24.0	2×10^{-3}
brant	6×10^{-3}	6.7	1×10^{-3}
glaucous gull	1×10^{-2}	6.4	2×10^{-3}
pectoral sandpiper	6×10^{-2}	16.9	4×10^{-3}
peregrine falcon	4×10^{-3}	7.8	1×10^{-3}
spectacled eider	3×10^{-4}	6.6	5×10^{-5}
Steller's eider	4×10^{-4}	7.8	5×10^{-5}
brown lemming	4×10^{-2}	715.0	1×10^{-4}
Arctic fox	5×10^{-4}	160.0	3×10^{-6}
ZINC			
Lapland longspur	4	5.0	7×10^{-1}
brant	1	1.0	1.1
glaucous gull	3×10^{-1}	1.0	3×10^{-1}
pectoral sandpiper	2	4.0	6×10^{-1}
peregrine falcon	9×10^{-3}	4.0	2×10^{-3}
spectacled eider	2×10^{-2}	3.0	7×10^{-3}
Steller's eider	2×10^{-2}	4.0	5×10^{-3}
brown lemming	5	38.0	2×10^{-1}
Arctic fox	1×10^{-3}	9.0	1×10^{-4}
DRPH			
Lapland longspur	4	517.0	7×10^{-3}
brant	1	140.0	8×10^{-3}
glaucous gull	1	136.0	1×10^{-4}
pectoral sandpiper	20	357.0	4×10^{-2}
peregrine falcon	1×10^{-1}	165.0	6×10^{-4}

NC Not calculated because toxicity values were not available.

TABLE 3-30. RISK CHARACTERIZATION OF REPRESENTATIVE BIRD, MAMMAL, AND PROTECTED SPECIES AT THE BARTER ISLAND INSTALLATION (CONTINUED)

SPECIES	ESTIMATED DAILY DOSE mg/kg-BW/DAY	TRV mg/kg-BW/DAY	TRV RATIO
spectacled eider	5×10^{-2}	139.0	4×10^{-4}
Steller's eider	7×10^{-2}	165.0	4×10^{-4}
brown lemming	10	284.0	4×10^{-2}
Arctic fox	2×10^{-2}	63.0	3×10^{-4}
ETHYLBENZENE			
Lapland longspur	5×10^{-3}	NC	NC
brant	1×10^{-3}	NC	NC
glaucous gull	4×10^{-4}	NC	NC
pectoral sandpiper	3×10^{-3}	NC	NC
peregrine falcon	2×10^{-5}	NC	NC
spectacled eider	3×10^{-5}	NC	NC
Steller's eider	3×10^{-5}	NC	NC
brown lemming	6×10^{-2}	210.0	3×10^{-4}
Arctic fox	2×10^{-6}	47.0	4×10^{-8}
NAPHTHALENE			
Lapland longspur	4×10^{-2}	NC	NC
brant	1×10^{-2}	NC	NC
glaucous gull	4×10^{-3}	NC	NC
pectoral sandpiper	4×10^{-2}	NC	NC
peregrine falcon	3×10^{-4}	NC	NC
spectacled eider	2×10^{-4}	NC	NC
Steller's eider	3×10^{-4}	NC	NC
brown lemming	9×10^{-1}	38.0	2×10^{-2}
Arctic fox	3×10^{-5}	9.0	3×10^{-6}
2-METHYLNAPHTHALENE			
Lapland longspur	2×10^{-2}	NC	NC
brant	5×10^{-3}	NC	NC
glaucous gull	3×10^{-3}	NC	NC

NC Not calculated because toxicity values were not available.

TABLE 3-30. RISK CHARACTERIZATION OF REPRESENTATIVE BIRD, MAMMAL, AND PROTECTED SPECIES AT THE BARTER ISLAND INSTALLATION (CONTINUED)

SPECIES	ESTIMATED DAILY DOSE mg/kg-BW/DAY	TRV mg/kg-BW/DAY	TRV RATIO
pectoral sandpiper	3×10^{-2}	NC	NC
peregrine falcon	2×10^{-4}	NC	NC
spectacled eider	1×10^{-4}	NC	NC
Steller's eider	1×10^{-4}	NC	NC
brown lemming	2×10^{-1}	NC	NC
Arctic fox	2×10^{-5}	NC	NC
TRIMETHYLBENZENE			
Lapland longspur	1×10^{-2}	NC	NC
brant	3×10^{-3}	NC	NC
glaucous gull	2×10^{-3}	NC	NC
pectoral sandpiper	2×10^{-2}	NC	NC
peregrine falcon	1×10^{-4}	NC	NC
spectacled eider	7×10^{-5}	NC	NC
Steller's eider	9×10^{-5}	NC	NC
brown lemming	4×10^{-1}	NC	NC
Arctic fox	1×10^{-5}	NC	NC
PCBs			
Lapland longspur	1×10^{-3}	116.0	9×10^{-6}
brant	3×10^{-4}	32.0	9×10^{-6}
glaucous gull	5×10^{-4}	31.0	2×10^{-5}
pectoral sandpiper	6×10^{-3}	80.0	8×10^{-5}
peregrine falcon	5×10^{-5}	37.0	1×10^{-6}
spectacled eider	2×10^{-5}	31.0	7×10^{-7}
Steller's eider	2×10^{-5}	37.0	5×10^{-7}
brown lemming	5×10^{-3}	3.0	2×10^{-3}
Arctic fox	6×10^{-6}	0.6	1×10^{-5}

NC Not calculated because toxicity values were not available.

TABLE 3-30. RISK CHARACTERIZATION OF REPRESENTATIVE BIRD, MAMMAL, AND PROTECTED SPECIES AT THE BARTER ISLAND INSTALLATION (CONTINUED)

SPECIES	ESTIMATED DAILY DOSE mg/kg-BW/DAY	TRV mg/kg-BW/DAY	TRV RATIO
XYLENES			
Lapland longspur	2×10^{-2}	26.0	8×10^{-4}
brant	6×10^{-3}	7.2	8×10^{-4}
glaucous gull	2×10^{-3}	7.0	3×10^{-4}
pectoral sandpiper	2×10^{-2}	18.4	1×10^{-3}
peregrine falcon	1×10^{-4}	8.4	1×10^{-5}
spectacled eider	1×10^{-4}	7.1	1×10^{-5}
Steller's eider	2×10^{-4}	8.5	2×10^{-5}
brown lemming	2×10^{-1}	165.0	1×10^{-3}
Arctic fox	2×10^{-5}	37.0	5×10^{-7}

NC Not calculated because toxicity values were not available.

Vaccinium spp. is likely to be low. The concentrations of VOCs at the site are substantially lower than toxicity values reported by Galassi et al. (1988 in COE 1991) and Hutchinson et al. (1980 in COE 1991) and listed in Table 3-28. These VOCs are not expected to be present at significant levels in most plants because of their volatility, absorption to soil particles, metabolism, or degradation rates in soil (Kostecki and Calabrese 1989). Overall, the potential risk to representative species of plants can be characterized as low.

3.4.2 Potential Risks to Representative Species of Aquatic Organisms

Estimates of exposure for aquatic organisms are based on the average concentrations of COC in surface water samples (Section 3.1). The TRVs for aquatic species are presented in Table 3-20. The TRV ratios are calculated by dividing the estimated exposure concentration by the TRV. Table 3-29 presents the results of the risk characterization for aquatic organisms. The following paragraphs summarize the potential risks to aquatic organisms from each COC in surface water (DRPH, aluminum, iron and manganese).

The TRV ratios for aluminum in surface water ranged from 3.1 to 3.4. These ratios indicate that a potential risk to aquatic organisms may exist from aluminum concentrations in surface water. Although these TRV ratios indicate a potential risk to aquatic species, some qualifying factors should be considered. Aluminum was selected as a COC because it exceeded the action level concentration of 87 $\mu\text{g/L}$ (Federal Ambient Water Quality Criteria). The concentration of aluminum used as the exposure estimate (318 $\mu\text{g/L}$) is within the range of the background concentrations (100 to 350 $\mu\text{g/L}$). In addition, aluminum was detected in only 5 of 10 samples. These facts, combined with the uncertainty factors used in establishing the TRV, suggest minimal risk to aquatic species from aluminum in surface waters.

Iron TRV ratios are 8.6 for the three aquatic organisms. Although these elevated TRV ratios may indicate risk, it is not likely that the arctic char or nine-spined stickleback are found in the locations where iron concentrations are elevated (i.e., intermittent drainages). Also, the TRV ratios are based on total iron concentrations. The risk to *Daphnia* spp. may be better represented by dissolved iron concentrations. The background concentrations for dissolved iron are from several DEW Line installations range from 100 to 1,600 $\mu\text{g/L}$. The average concentration of dissolved iron in the Barter Island detected samples is 1,421 $\mu\text{g/L}$. Not only is the average dissolved iron concentration below the maximum background concentration at the DEW Line installations, iron is an essential nutrient and the uncertainties used in establishing the TRV value may overestimate the toxicity of iron. With these facts in mind, iron should be viewed as a low risk to the aquatic organisms.

Manganese TRV ratios range from 2.8 for *Daphnia* spp. to 7.9 for the Arctic char and the nine-spined stickleback. This indicates a potential risk to these species, however some mitigating factors should be considered. The average manganese concentration at the Barter Island installation of 574 $\mu\text{g/L}$ is only slightly greater than the maximum background concentration of 510 $\mu\text{g/L}$. These background concentrations approximate normal manganese concentrations in surface waters (USGS 1985). Therefore, it is likely the potential risk to aquatic species from manganese at the Barter Island installation is minimal.

The TRV ratios for DRPH in surface water range from 0.7 to 1.0 (Table 3-29). The TRV ratio for the fish species exposed to DRPH is 0.7. This indicates that DRPH do not pose potential risks to the Arctic char and nine-spined stickleback. The TRV ratio of DRPH for *Daphnia* spp. is 1.0. This ratio indicates that *Daphnia* spp. may be at risk in surface water. However, given the equal ratio, the uncertainties of the risk assessment and the TRV, and the presence of elevated levels of DRPH in only 8 of 28 samples, the likelihood of negative effects is relatively low.

3.4.3 Potential Risks to Representative Species of Birds

The avian TRV ratios for aluminum, iron, lead, manganese, DRPH, PCBs, and xylenes are below 1.0 for all species of birds evaluated, including the protected species (peregrine falcon, spectacled eider, and Steller's eider) (see Table 3-30). The TRV ratio for zinc (1.1) indicates that the brant may be at risk, however the average concentration of 76 mg/kg is below the maximum background concentration. Overall, the TRV ratios indicate that there is little risk to bird species from these chemicals at the Barter Island installation. No toxicity values were available for ethylbenzene, naphthalene, 2-methylnaphthalene, and trimethylbenzene. The sampling detected ethylbenzene in 19 of 86, naphthalene in 5 of 30, 2-methylnaphthalene in 3 of 18, and trimethylbenzene in 7 of 30 soil/sediment samples. The frequency of detections, combined with the relatively low concentrations and resulting minimal exposures (see Table 3-30, estimated daily dose column) indicate that the risk from these chemicals is low.

3.4.4 Potential Risks to Representative Species of Mammals

TRV ratios for the brown lemming and the arctic fox were less than one, except for the lead TRV ratio for the brown lemming which was 1.6 (Table 3-30). A TRV ratio greater than one indicates that the estimated chemical intake rate exceeds the intake rate expected to be without adverse effect in the subject species. Thus, the lemming intake rate of lead may be of concern, however there are mitigating circumstances. The maximum background concentration for lead is 22 mg/kg and the average lead concentration in the samples collected is 26.6 mg/kg. Lead was detected in 8 of 16 samples. There is a significant amount of uncertainty in the B_v coefficient used to estimate lead uptake by the plants in the lemming diet. Furthermore, some of the samples collected were sediments that may be submerged, and therefore not available to the lemming. These circumstances and the closeness of the lead TRV ratio to one suggest that the potential for adverse effects from the exposure of lemmings to lead may be negligible.

Because the necessary toxicity values were not available for mammals in the published literature, TRV ratios were not calculated for 1,2,4- and 1,3,5-trimethylbenzene, or for 2-methylnaphthalene. These chemicals are not expected to contribute significantly to the potential for adverse effects in mammals. They were observed at relatively low concentrations and detection frequencies; trimethylbenzenes were observed in 7 of 30 samples, and 2-methylnaphthalene was observed in 3 of 10 samples. Furthermore, the toxicity of these compounds for ecological receptors is expected to be low (Verschuere 1983).

3.4.5 Potential Future Risks

Future ecological risks at the Barter Island installation were based on the assumptions that all buildings may be removed from the 14 sites and the gravel pads will remain in place. Future risks at the ten sites (with the exception of those from PCB contamination at the Heated Storage site discussed in the risk estimate paragraph below) that have potentially suitable habitat for the representative species are expected to be as low as or lower than current risks because the exposure pathways are not likely to change, and the concentrations of COCs are likely to diminish over time. These sites include the Old Landfill (LF01), POL Catchment (LF03), Current Landfill (LF04), Contaminated Ditch (SD08), Old Runway Dump (LF12), Heated Storage (SS13), Garage (SS14), Fuel Tanks (ST18), Old Dump Site (LF19), and Bladder Diesel Spill (SS20).

The remaining four sites, the Weather Station Building (SS15), White Alice Facility (SS16), POL Tanks (ST17), and JP-4 Spill (SS21) were not included in the ecological risk assessment because suitable habitats are not present at these sites under current conditions. Each of these sites is characterized by gravel pads and is generally devoid of vegetation and surface water. The lack of suitable habitat precludes complete exposure pathways. It is possible that pathways could exist in the future. Two factors play key roles in determining the potential future risks at these sites: the revegetation of the gravel pads and the fate of the COCs over time.

Revegetation of gravel pads in arctic Alaska has been studied by Bishop and Chapin (1989a, 1989b) and the information that follows is based on their research. Revegetation is dependent on several factors, including the availability of water and mineral nutrients and the dispersal ability of the recolonizing species. The large particle size found in gravel pad soils contributes to reduced moisture retention (in an environment with only 10-20 cm annual precipitation) and low organic matter content that restricts germination of prospective colonizing seeds. If revegetation occurs as a result of natural processes, riparian species are the most likely to successfully colonize gravel pads. Typically, none of the predominant species in the wet or moist tundra habitats are responsible for natural revegetation on gravel pads. Riparian gravel bars are similar in physical composition to gravel pads and riparian areas are subject to frequent natural disturbance, resulting in a characteristic flora that has a higher proportion of pioneer species. Bishop and Chapin (1989b) found that at 16 gravel pads abandoned for 10 years, the average revegetative cover was 2.7%. Half these gravel pads were within 0.2 km of the Sagavanirktok River and its riparian habitat provided the seed sources for revegetation.

Based on the relatively low availability of riparian recolonizing seed sources near the Barter Island sites and the inherent low water and mineral nutrient availability at the gravel pads, the time required for revegetation to provide suitable habitat for the representative species at the sites is expected to be at least 20 years, with revegetation times of 50 to 100 years possible. During this time, any COCs present may undergo changes that affect the potential exposure and risk to representative species. These changes will vary with the nature of specific COC. The COCs at the four sites are DRPH, GRPH, VOCs and SVOCs characteristic of DRPH (e.g., toluene, isopropyltoluene, ethylbenzene, xylenes, naphthalene, 2-methylnaphthalene, and substituted benzenes), and PCBs.

DRPH and GRPH are complex mixtures of aliphatic and aromatic hydrocarbons with exposure potentials based on the environmental fate of individual components of the mixture. DRPH are hydrocarbon molecules with more carbon atoms than GRPH, therefore DRPH is relatively less volatile than GRPH. Most GRPH will evaporate from soil, whereas DRPH will biodegrade in soil. For example, a single application of about 21, 14, or 13 g/kg soil of home heating oil no. 2 to outdoor plots consisting of silt loam, sandy loam and clay loam was degraded by 86%, 90%, and 86%, respectively, after one year (Raymond et al. 1975, 1976 in ATSDR 1993a).

Jet fuel (JP-4) is highly volatile, however in soil volatilization accounted for 7% of the hydrocarbon loss compared with 93% for biodegradation (Coho 1990 in ATSDR 1993b). Soil type, temperature and jet fuel concentration affected biodegradation, but the half-life of JP-4 in clay at 27°C was shown to be 3.5 weeks (ATSDR 1993b).

PCBs, on the other hand, have very high chemical, thermal, and biological stability in addition to low vapor pressure (Manahan 1994). Conversion of highly substituted PCBs to molecules with one or two chlorines is done relatively slowly by anaerobic bacteria (Manahan 1994). Therefore, biodegradation may not be an effective process for reducing the PCB concentrations at White Alice Facility to be protective of the environment. In addition, PCBs have a high potential for bioaccumulation (see Section 3.2.7.1 and Table 3-6 for information on bioaccumulation), which may result in exposure to PCBs of ecological receptors in the trophic web.

Risk estimates for potential future exposures at these sites are presented below. The estimates are based on the revegetation information and changes that can be expected to occur to the COCs. The hydrocarbons and related volatile and semivolatile compounds (DRPH, GRPH, VOCs, and SVOCs) detected at the Weather Station Building, White Alice Facility, POL Tanks, and JP-4 Spill may undergo volatilization and biodegradation such that their concentrations over 20 to 50 years will diminish to levels that are not expected to present significant risk to ecological receptors. The 20 to 50-year time span is a conservative estimate of the time required for these sites to provide suitable habitat for representative species and for potential exposure pathways to develop. The potential future risk attributable to PCBs found at the White Alice Facility is estimated to be high at the S01, S06 and S07 sample locations (see Figure 2-12 for the sample locations). The risk estimate is based on elevated concentrations of PCBs (exceeding action levels), which are unlikely to degrade and PCBs may bioaccumulate. Although the PCB concentration at sample location 2S09 is below the action level (8.7 mg/kg compared with the action level of 10 mg/kg), the potential future risk is estimated to be moderate because of the potential for bioaccumulation. The four sample locations are grouped together, adjacent to the concrete pad on which the transformer sits. It is recommended that a limited excavation of PCB contaminated soils be conducted, in conjunction with the building removal, to mitigate the potential future risk associated with the PCBs at the White Alice Facility. PCBs are also found at the S01, S02, S03, S04 and SD01 sample locations at the Heated Storage site (SS13). Future risk that may be associated with the potential bioaccumulation of PCBs at these locations is estimated to be moderate, and limited excavation of PCB contaminated soils is also recommended for these locations.

3.5 ECOLOGICAL RISK ASSESSMENT UNCERTAINTY ANALYSIS

As with any risk assessment, there is great uncertainty associated with the estimates of ecological risk for the sites at the Barter Island installation. The risk estimates are based on a number of assumptions regarding exposure and toxicity. In general, the primary sources of uncertainty are the following:

- Environmental Sampling and Analysis;
- Selection of COC;
- Exposure Parameter Estimation; and
- Toxicological Data.

A complete understanding of the uncertainties associated with risk estimates is critical to placing the predicted risks in proper perspective. The most significant sources of uncertainty associated with the estimates of risk for the Barter Island installation sites are summarized in the following sections.

3.5.1 Environmental Sampling and Analysis

The principal source of uncertainty in the analytical data (for the ERA) stems from the sampling approach and the subsequent calculation of exposure concentrations. Sampling at the Barter Island installation was conducted in a systematic manner, designed to characterize localized contaminated areas or "hot spots". The site's potential source areas are therefore well characterized, however there are limited data regarding the peripheral areas (areas to which ecological receptors are most likely to be exposed). In order to compensate for this non-random sampling methodology in the calculation of exposure concentrations, the exposure assessment used the average concentration of COCs.

The methods of calculating the average concentrations were the same for organic and inorganic data. In calculating the average concentration of chemicals at the site, non-detected chemicals were entered at one-half of the quantitation limit, as per EPA guidance (EPA 1989a). Sampling was designed to characterize "hot spots" at each of the sites. Therefore, the average concentrations of COCs tend to be biased high because sampling was generally concentrated in areas of the site where significant contamination exists or was suspected. The use of total metal concentrations in surface water to estimate risk is a conservative approach because dissolved metal concentrations are generally significantly less than total metal concentrations. Therefore, the average concentrations used to estimate exposure to metals in surface water may overestimate potential risk.

In addition, there is uncertainty inherent in using measurements of TPH (or DRPH, GRPH, and RRPH) for risk assessments. The analytical techniques are not specific to petroleum (i.e., they detect other organics, including naturally-occurring ones) (Von Burg 1993). Moreover, the toxicity of these groups of petroleum hydrocarbons is determined by the toxicity of their individual constituents. When petroleum compounds are released to the environment, they tend to weather or transform readily. For example, the lighter fractions (such as BTEX) will volatilize to the atmosphere more readily than the heavier fractions (such as decane, pyrene or benzo(a)pyrene).

The lighter fractions are thought to be the more toxic (Wong et al. 1981; O'Brien 1978; Kauss and Hutchinson 1975; Soto et al. 1975). Therefore the toxicity of DRPH, GRPH, and RRPH is expected to change over time, depending upon the attenuation mechanisms occurring in the environment. As a result, the toxicity of the petroleum hydrocarbons measured at the Barter Island installation is unknown. Use of toxicity values reported in the literature probably contributes to an overestimation of the risk, because it is likely that the most toxic components of the mixtures detected have volatilized to the atmosphere over time.

3.5.2 Selection of Chemicals for Evaluation

The selection of COCs in the ERA was based upon a comparison to background concentrations and action levels, and an evaluation of the frequency of detection. This provided a conservative screen of COCs, therefore it is unlikely that any chemicals presenting an ecological risk were omitted.

3.5.3 Exposure Assessment

Exposures were estimated based on literature-based life history information for the selected representative species. There is moderate uncertainty associated with the exposure information. Food and water ingestion rates were not available for some animals and had to be estimated from regression equations. Incidental ingestion of soils and sediments may occur while animals are foraging in these media, and it is uncertain how much soil and/or sediment is actually ingested. There are significant uncertainties associated with the estimates of how extensively a receptor will use the site, which were based on home range information and site-specific considerations of habitat quality. As noted in the discussion of Estimation of Percent Ingested Onsite, Section 3.2.7.2, the conversion of population density values as substitutes for home ranges adds uncertainty to the risk assessment. The conversion was necessary because home range data is lacking for some of the representative species. In addition, there is uncertainty associated with the habitat suitability determinations. These determinations were made using installation and site maps and photographs taken on the ground by the sampling team. The final determination of habitat suitability was based on the professional judgement of a wildlife biologist.

There is some uncertainty associated with the diet compositions estimated from information obtained from the literature. A good example of this type of uncertainty is the unpredictable fluctuation in the populations of the brown lemmings and their predators (i.e., Arctic fox, glaucous gull). As the numbers of prey increase, predator populations may experience numerical and density increases well beyond the values reported in the literature. When prey populations decrease, predation pressure can shift to diet items that are not considered "normal", and do not represent dietary intakes reported in the literature. Wildlife, and their interactions with the environment around them, are dynamic. Stochastic events, natural or anthropogenic, may cause behavior and/or habits to differ markedly from the "expected or norm". Deviations from typical behavior cause uncertainty when evaluating wildlife and ecosystems.

There is some uncertainty associated with exposure estimates for plants. Plant uptake of COC was derived from a regression equation using the K_{ow} of the COC (Table 3-6). This calculation estimates the concentration of chemicals in the vegetative portion of plants. Actual

concentrations of the COC in plant tissue will vary depending upon actual chemical uptake, species of plant, and other site-specific factors (such as soil organic carbon). It is important to note that maximum acceptable tissue concentrations in plants were not available for comparison with these estimated concentrations. As a result, it is uncertain whether the estimates are phytotoxic. However, the overall effect of this source of uncertainty in the risk assessment is low as is the ecological risk to plants.

3.5.4 Toxicological Data

One of the largest sources of uncertainty in risk assessment is from the toxicological data. Often there are not relevant studies for the specific representative species or endpoints. As a result, extrapolations are made, which introduce uncertainty into the risk estimate. These extrapolations incorporate uncertainty factors into the calculation of TRVs. The purpose of the uncertainty factors is to incorporate some margin of error into the risk estimate, in order to arrive at a "safe" level of exposure to which onsite exposure concentrations may be compared. These techniques introduce into the risk assessment a tendency to overestimate rather than underestimate the risk, as conservative estimates were made in estimating toxicity values.

Toxicity values for plants, water, soils and sediments are based on literature values. Toxicity in soils and sediments is affected by the bioavailability of a given chemical. Toxicity of metals in water is based, in part, upon the speciation of the element. As a result, site-specific bioavailability or toxicity may differ from that in the studies used to estimate potential toxic effects. Therefore actual toxicity of chemicals at the Barter Island installation may be different from the values reported in the literature. In addition, the sensitivity of receptors on site may be different from the sensitivity of the species reported in the literature. This contributes to the overall uncertainty of the risk assessment.

There is also a great deal of uncertainty in assessing the toxicity of a mixture of chemicals. In this ERA, the effects of exposure from each contaminant have been considered separately. However, these substances occur together at the site and organisms may be exposed to mixtures of the chemicals. Prediction of how these mixtures of toxicants will interact must be based on an understanding of the mechanisms of such interactions. The interactions of the individual components of chemical mixtures may occur during absorption, distribution, metabolism, excretion, or activity at the receptor site. Individual compounds may interact chemically, yielding a new toxic component or causing a change in the biological availability of an existing component, or may interact by causing different effects at different receptor sites. Suitable data are not currently available to characterize rigorously the effects of chemical mixtures, so chemicals present at the site were evaluated independently. This approach of assessing risk associated with mixtures of chemicals does not account for any additive, synergistic or antagonistic interactions among the chemicals considered. However, as discussed in Section 3.6, the risk assessment yielded a low potential for ecological risks and it is unlikely that additive effects of chemicals are a concern.

3.6 SUMMARY OF ECOLOGICAL RISK

The potential risks to ecological receptors are summarized in this section based on the information presented in Sections 3.1 through 3.4. The reader is referred to these sections for more details on the assessment. Conclusions regarding potential risks must be viewed in the context of the uncertainties associated with the assessment and the available risk information. The available risk information includes chemical data, exposure estimates, and literature-based toxicity information.

Table 3-31 presents a summary of the ecological risks at the Barter Island installation. The table includes the potential risk to each ecological group evaluated, the COC that contributed to the risk, and the site(s) where the COCs were detected at relatively high concentrations.

3.6.1 Potential Risks to Representative Plants

A qualitative comparison was made of site soil/sediment chemical concentrations and plant toxicity information. The risk to plants is characterized by using comparative information from the literature and BCF (B_v).

The spacial distribution of the contaminants at Barter Island is very uneven. The locations that have elevated concentrations of potential contaminants are few and spatially distinct. In addition, during the RI field activities no stressed vegetation was observed and Alaska cotton grass (*Eriophorum* species) was flourishing in heavily contaminated areas of the POL Catchment, site LF03. Based on visual observations and the ecological assessment, risks to plants are low.

3.6.2 Potential Risks to Representative Aquatic Species

Potential risks to aquatic species were evaluated by comparing toxicity information from the literature with the average exposure concentrations of potential contaminants in surface water. TRV ratios for aquatic organisms indicate that risks may exist from aluminum, iron, manganese, and DRPH. The TRV ratios for fish and aquatic invertebrates for aluminum are 3.1 and 3.4, and for manganese are 7.9 and 2.8, respectively. The TRV ratio for iron for aquatic organisms is 8.6. In the case of the metals, the average concentrations that were used as estimated exposure point concentrations were near, or in some cases within, the ranges of the background concentrations compiled from studies done at seven DEW Line installations. It is possible that the metals detected are representative of elevated natural background concentrations rather than a result of any activities at the Barter Island installation.

In addition, the potential risks to aquatic species was based on total metal concentrations, if dissolved metal concentrations were used in the risk evaluation the risks would be significantly less than noted above.

The TRV ratio of DRPH for *Daphnia* spp. is 1. This indicates that risk may exist, but the small number of locations that are contaminated indicates that the risk to the aquatic ecosystem is low.

TABLE 3-31. SUMMARY OF POTENTIAL ECOLOGICAL RISKS

ECOLOGICAL GROUP	POTENTIAL RISK	COC	SITES
Plants	Low	Zinc in soil/sediment	Heated Storage (SS13) Garage (SS14)
Aquatic Organisms	Low	Aluminum, iron, manganese, and DRPH in surface water	Old Landfill (LF01) Current Landfill (LF04) Contaminated Ditch (SD08) Old Dump Site (LF19)
Birds	Low	Zinc in soil/sediment	Heated Storage (SS13) Garage (SS14)
Mammals	Low	Lead in soil/sediment	Garage (SS14)
Birds/Mammals	Bioaccumulation	PCBs in soil/sediment	Heated Storage (SS13) White Alice Facility (SS16)

3.6.3 Potential Risks to Representative Species of Birds and Mammals

The risks to representative species of birds and mammals were evaluated using the Quotient Method. This method compares the estimated dose with TRVs. The resulting ratios indicate, with two exceptions, that the risks to birds and mammals are low. The two instances of TRV ratios greater than 1.0 involve lead for the brown lemming, and zinc for the brant. These TRV ratios (1.6 for the lemming and 1.1 for the brant) are close to 1.0 and, considering the uncertainties involved and the conservative nature of the assessment, the ultimate risks to these two species is minimal. Future risk that may be associated with the potential bioaccumulation of PCBs is estimated to be moderate. Based on the results of the assessment, the bird and mammal populations at the Barter Island installation are not expected to be at risk from the COCs.

3.6.4 Potential Ecological Risks

The objective of this ERA is to evaluate the potential risk to the representative species at the site at the Barter Island DEW Line installation. This assessment indicates that, although there are a few instances of minimal potential risk to individual species, overall the potential risks presented by the COCs are very low.

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APPENDIX A
RISK CHARACTERIZATION SPREADSHEETS

TABLE A-1. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Water Ingestions
 Endpoint: Noncancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: Old Landfill
 File: LF01WANG.WK1

Exposure Assumptions		DEW Line Worker		Native Northern Adult	Native Northern Child
Water Ingestion	(L/day)	2		2	NA
Exposure Frequency	(days/year)	180		180	NA
Exposure Duration	(years)	10		55	NA
Conversion Factor	(kg/kg)	1		1	NA
Body Weight	(kg)	70		70	NA
Averaging Time	(days)	3,650		20,075	NA

Chemical	Oral RfD	Concentration Water (mg/L)	ADD by Receptor Group (mg/kg-day)			Hazard Quotient		
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult	Native Northern Child
Manganese	0.005	1.5	2.11e-02	2.11e-02	NA	4.23e+00	4.23e+00	NA
					HAZARD INDEX	4.227	4.227	NA

TABLE A-2. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Soil Ingestions
 Endpoint: Noncancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: POL Catchment (LF03)
 File: LF03SONC.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Soil Ingestion Rate	(mg/day)	50	100	200
Exposure Frequency	(days/year)	30	30	30
Exposure Duration	(years)	10	49	6
Conversion Factor	(kg/mg)	0.000001	0.000001	0.000001
Body Weight	(kg)	70	70	15
Averaging Time	(days)	3,650	17,885	2,190

Chemical	Oral RfD	Concentration Soil (mg/kg)	ADD by Receptor Group (mg/kg-day)			Hazard Quotients	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult/Child
DRPH	0.08	28,600	1.68e-03	3.36e-03	3.13e-02	2.10e-02	4.34e-01
Tetrachloro- ethene	0.01	5.42	3.18e-07	6.36e-07	5.94e-06	3.18e-05	6.58e-04
			HAZARD INDEX			0.021	0.434

TABLE A-3. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Soil Ingestions
 Endpoint: Cancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: POL Catchment (LF03)
 File: LF03SOCA.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Soil Ingestion Rate	(mg/day)	50	100	200
Exposure Frequency	(days/year)	30	30	30
Exposure Duration	(years)	10	49	6
Conversion Factor	(kg/mg)	0.000001	0.000001	0.000001
Body Weight	(kg)	70	70	15
Averaging Time	(days)	25,550	25,550	25,550

Chemical	Carcinogen Oral Slope Factor	Concentration Soil (mg/kg)	LADD by Receptor Group (mg/kg-day)			Cancer Risk	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult/Child
Tetrachloro-ethene	0.052	5.42	4.55e-08	4.45e-07	5.09e-07	2.36e-09	4.96e-08
			CANCER RISK			2e-09	5e-08

TABLE A-4. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Water Ingestions
 Endpoint: Noncancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: POL Catchment (LF03)
 File: LF03WANC.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Water Ingestion	(L/day)	2	2	NA
Exposure Frequency	(days/year)	180	180	NA
Exposure Duration	(years)	10	55	NA
Conversion Factor	(kg/mg)	1	1	NA
Body Weight	(kg)	70	70	NA
Averaging Time	(days)	3,650	20,075	NA

Chemical	Oral RfD	Concentration Water (mg/L)	ADD by Receptor Group (mg/k-day)				Hazard Quotient		
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult	Native Northern Child	
DRPH	0.08	1.77	2.49e-02	2.49e-02	NA	3.12e-01	3.12e-01	NA	
GRPH	0.2	0.367	5.17e-03	5.17e-03	NA	2.59e-02	2.59e-02	NA	
			HAZARD INDEX			0.338	0.338	NA	

TABLE A-5. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Water Ingestions
 Endpoint: Cancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: POL Catchment (LF03)
 File: LF03WACA.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Water Ingestion	(L/day)	2	2	NA
Exposure Frequency	(days/year)	180	180	NA
Exposure Duration	(years)	10	55	NA
Conversion Factor	(kg/mg)	1	1	NA
Body Weight	(kg)	70	70	NA
Averaging Time	(days)	25,550	20,075	NA

Chemical	Carcinogen Oral Slope Factor	Concentration Water (mg/L)	LADD by Receptor Group (mg/k-day)				Cancer Risk	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult	Native Northern Child
GRPH	0.0017	0.367	7.39e-04	5.17e-03	NA	1.26e-06	8.79e-06	NA
Benzene	0.029	0.0027	5.43e-06	3.80e-05	NA	1.58e-07	1.10e-06	NA
			CANCER RISK			1e-06	1e-05	NA

TABLE A-6. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Water Ingestions
 Endpoint: Noncancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: Current Landfill (LF04)
 File: LF04WANC.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Water Ingestion	(L/day)	2	2	NA
Exposure Frequency	(days/year)	180	180	NA
Exposure Duration	(years)	10	55	NA
Conversion Factor	(kg/mg)	1	1	NA
Body Weight	(kg)	70	70	NA
Averaging Time	(days)	3,650	20,075	NA

Chemical	Oral RfD	Concentration Water (mg/L)	ADD by Receptor Group (mg/k-day)				Hazard Quotient	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult	Native Northern Child
Manganese	0.005	1.8	2.54e-02	2.54e-02	NA	5.07e+00	5.072	NA
			HAZARD INDEX			5.072	5.072	NA

TABLE A-7. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Water Ingestions
 Endpoint: Cancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: Current Landfill (LF04)
 File: LF04WACA.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Water Ingestion	(L/day)	2	2	NA
Exposure Frequency	(days/year)	180	180	NA
Exposure Duration	(years)	10	55	NA
Conversion Factor	(kg/mg)	1	1	NA
Body Weight	(kg)	70	70	NA
Averaging Time	(days)	25,550	20,075	NA

Chemical	Carcinogen Oral Slope Factor	Concentration Water (mg/L)	LADD by Receptor Group (mg/k-day)				Cancer Risk	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult	Native Northern Child
Trichloroethene	0.011	0.036	7.25e-05	5.07e-04	NA	7.97e-07	5.58e-06	NA
			CANCER RISK			8e-07	6e-06	NA

TABLE A-8. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Soil Ingestions
 Endpoint: Noncancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: Contaminated Ditch (SD08)
 File: SD08SONC.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Soil Ingestion Rate	(mg/day)	50	100	200
Exposure Frequency	(days/year)	30	30	30
Exposure Duration	(years)	10	49	6
Conversion Factor	(kg/mg)	0.000001	0.000001	0.000001
Body Weight	(kg)	70	70	15
Averaging Time	(days)	3,650	17,885	2,190

Chemical	Oral RfD	Concentration Soil (mg/kg)	ADD by Receptor Group (mg/kg-day)			Hazard Quotients	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult/Child
DRPH	0.08	2,260	1.33e-04	2.65e-04	2.48e-03	1.66e-03	3.43e-02
GRPH	0.2	171	1.00e-05	2.01e-05	1.87e-04	5.02e-05	1.04e-03
Beryllium	0.005	3.2	1.88e-07	3.76e-07	3.51e-06	3.76e-05	7.77e-04
			HAZARD INDEX			0.002	0.036

TABLE A-9. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Soil Ingestions
 Endpoint: Cancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: Contaminated Ditch (SD08)
 File: SD08SOCA.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Soil Ingestion Rate	(mg/day)	50	100	200
Exposure Frequency	(days/year)	30	30	30
Exposure Duration	(years)	10	49	6
Conversion Factor	(kg/mg)	0.000001	0.000001	0.000001
Body Weight	(kg)	70	70	15
Averaging Time	(days)	25,550	25,550	25,550

Chemical	Carcinogen Oral Slope Factor	Concentration Soil (mg/kg)	LADD by Receptor Group (mg/kg-day)			Cancer Risk	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult/Child
GRPH	0.0017	171	1.43e-06	1.41e-05	1.61e-05	2.44e-09	5.12e-08
Beryllium	4.3	3.2	2.68e-08	2.63e-07	3.01e-07	1.15e-07	2.42e-06
			CANCER RISK			1e-07	2e-06

TABLE A-10. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Soil Ingestions
 Endpoint: Noncancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: Heated Storage (SS13)
 File: SS13SONC.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Soil Ingestion Rate	(mg/day)	50	100	200
Exposure Frequency	(days/year)	30	30	30
Exposure Duration	(years)	10	49	6
Conversion Factor	(kg/mg)	0.000001	0.000001	0.000001
Body Weight	(kg)	70	70	15
Averaging Time	(days)	3,650	17,885	2,190

Chemical	Oral RfD	Concentration Soil (mg/kg)	ADD by Receptor Group (mg/kg-day)			Hazard Quotients	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult/Child
DRPH	0.08	3,580	2.10e-04	4.20e-04	3.92e-03	2.63e-03	5.43e-02
GRPH	0.2	423	2.48e-05	4.97e-05	4.64e-04	1.24e-04	2.57e-03
RRPH	0.08	2400	1.41e-04	2.82e-04	2.63e-03	1.76e-03	3.64e-02
Aroclor 1254	0.00002	2.72	1.60e-07	3.19e-07	2.98e-06	7.98e-03	1.65e-01
HAZARD INDEX						0.012	0.258

TABLE A-11. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Soil Ingestions
 Endpoint: Cancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: Heated Storage (SS13)
 File: SS13SOCA.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Soil Ingestion Rate	(mg/day)	50	100	200
Exposure Frequency	(days/year)	30	30	30
Exposure Duration	(years)	10	49	6
Conversion Factor	(kg/mg)	0.000001	0.000001	0.000001
Body Weight	(kg)	70	70	15
Averaging Time	(days)	25,550	25,550	25,550

Chemical	Carcinogen Oral Slope Factor	Concentration Soil (mg/kg)	LADD by Receptor Group (mg/kg-day)			Cancer Risk	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult/Child
GRPH	0.0017	423	3.55e-06	3.48e-05	3.97e-05	6.03e-09	1.27e-07
PCBs	7.7	2.72	2.28e-08	2.24e-07	2.55e-07	1.76e-07	3.69e-06
			CANCER RISK			2e-07	4e-06

TABLE A-12. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Water Ingestions
 Endpoint: Noncancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: Heated Storage Building (SS13)
 File: SS13WANC.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Water Ingestion	(L/day)	2	2	NA
Exposure Frequency	(days/year)	180	180	NA
Exposure Duration	(years)	10	55	NA
Conversion Factor	(kg/mg)	1	1	NA
Body Weight	(kg)	70	70	NA
Averaging Time	(days)	3,650	20,075	NA

Chemical	Oral RfD	Concentration Water (mg/L)	ADD by Receptor Group (mg/k-day)			Hazard Quotient		
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult	Native Northern Child
DRPH	0.08	5.76	8.12e-02	8.12e-02	NA	1.01e+00	1.01e+00	NA
Tetrachloro-ethene	0.01	0.012	1.69e-04	1.69e-04	NA	1.69e-02	1.69e-02	NA
Manganese	0.005	0.54	7.61e-03	7.61e-03	NA	1.52e+00	1.52e+00	NA
						2.55	2.55	NA

TABLE A-13. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Water Ingestions
 Endpoint: Cancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: Heated Storage Building (SS13)
 File: SS13WACA.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Water Ingestion	(L/day)	2	2	NA
Exposure Frequency	(days/year)	180	180	NA
Exposure Duration	(years)	10	55	NA
Conversion Factor	(kg/mg)	1	1	NA
Body Weight	(kg)	70	70	NA
Averaging Time	(days)	25,550	20,075	NA

Chemical	Carcinogen Oral Slope Factor	Concentration Water (mg/L)	LADD by Receptor Group (mg/k-day)				Cancer Risk		
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult	Native Northern Child	
Benzene	0.029	0.0069	1.39e-05	9.72e-05	NA	4.03e-07	2.82e-06	NA	
Tetrachloroethene	0.052	0.012	2.42e-05	1.69e-04	NA	1.26e-06	8.79e-06	NA	
			CANCER RISK			2e-06	1e-05	NA	

TABLE A-14. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Soil Ingestions
 Endpoint: Noncancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: Garage (SS14)
 File: SS14SONC.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Soil Ingestion Rate	(mg/day)	50	100	200
Exposure Frequency	(days/year)	30	30	30
Exposure Duration	(years)	10	49	6
Conversion Factor	(kg/mg)	0.000001	0.000001	0.000001
Body Weight	(kg)	70	70	15
Averaging Time	(days)	3,650	17,885	2,190

Chemical	Oral RfD	Concentration Soil (mg/kg)	ADD by Receptor Group (mg/kg-day)			Hazard Quotients	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult/Child
DRPH	0.08	12,400	7.28e-04	1.46e-03	1.36e-03	9.10e-03	1.88e-01
RRPH	0.08	27,000	1.59e-03	3.17e-03	2.96e-02	1.98e-02	4.09e-01
GRPH	0.2	700	4.11e-05	8.22e-05	7.67e-04	2.05e-04	4.25e-03
bis-(2ethylhexyl) phthalate	0.02	4.6	2.70e-07	5.40e-07	5.04e-06	1.35e-05	2.79e-04
HAZARD INDEX						0.029	0.602

TABLE A-15. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Soil Ingestions
 Endpoint: Cancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: Garage (SS14)
 File: SS14SOCA.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Soil Ingestion Rate	(mg/day)	50	100	200
Exposure Frequency	(days/year)	30	30	30
Exposure Duration	(years)	10	49	6
Conversion Factor	(kg/mg)	0.000001	0.000001	0.000001
Body Weight	(kg)	70	70	15
Averaging Time	(days)	25,550	25,550	25,550

Chemical	Carcinogen Oral Slope Factor	Concentration Soil (mg/kg)	LADD by Receptor Group (mg/kg-day)			Cancer Risk	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult/Child
GRPH	0.0017	700	5.87e-06	5.75e-05	6.58e-05	9.98e-09	2.10e-07
Benzene	0.029	1.4	1.17e-08	1.15e-07	1.32e-07	3.41e-10	7.15e-09
bis(2-ethylhexyl) phthalate	0.014	4.6	3.86e-08	3.78e-07	4.32e-07	5.40e-10	1.13e-08
			CANCER RISK			1e-08	2e-07

TABLE A-16. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Soil Ingestions
 Endpoint: Noncancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: Weather Station Building (SS15)
 File: SS15SONC.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Soil Ingestion Rate	(mg/day)	50	100	200
Exposure Frequency	(days/year)	30	30	30
Exposure Duration	(years)	10	49	6
Conversion Factor	(kg/mg)	0.000001	0.000001	0.000001
Body Weight	(kg)	70	70	15
Averaging Time	(days)	3,650	17,885	2,190

Chemical	Oral RfD	Concentration Soil (mg/kg)	ADD by Receptor Group (mg/kg-day)			Hazard Quotients	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult/Child
DRPH	0.08	8,420	4.94e-04	9.89e-04	9.23e-03	6.18e-03	1.28e-01
GRPH	0.2	1,020	5.99e-05	1.20e-04	1.12e-03	2.99e-04	6.19e-03
			HAZARD INDEX			0.006	0.134

TABLE A-17. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Soil Ingestions
 Endpoint: Cancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: Weather Station Building (SS15)
 File: SS15SOCA.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Soil Ingestion Rate	(mg/day)	50	100	200
Exposure Frequency	(days/year)	30	30	30
Exposure Duration	(years)	10	49	6
Conversion Factor	(kg/mg)	0.000001	0.000001	0.000001
Body Weight	(kg)	70	70	15
Averaging Time	(days)	25,550	25,550	25,550

Chemical	Carcinogen Oral Slope Factor	Concentration Soil (mg/kg)	LADD by Receptor Group (mg/kg-day)			Cancer Risk	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult/Child
GRPH	0.0017	1,020	8.55e-06	8.38e-05	9.58e-05	1.45e-08	3.05e-07
			CANCER RISK			1e-08	3e-07

TABLE A-18. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Soil Ingestions
 Endpoint: Noncancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: White Alice Facility (SS16)
 File: SS16SONC.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Soil Ingestion Rate	(mg/day)	50	100	200
Exposure Frequency	(days/year)	30	30	30
Exposure Duration	(years)	10	49	6
Conversion Factor	(kg/mg)	0.000001	0.000001	0.000001
Body Weight	(kg)	70	70	15
Averaging Time	(days)	3,650	17,885	2,190

Chemical	Oral RfD	Concentration Soil (mg/kg)	ADD by Receptor Group (mg/kg-day)			Hazard Quotients	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult/Child
Aroclor 1254	0.00002	52	3.05e-06	6.11e-06	5.70e-05	1.53e-01	3.15e+00
			HAZARD INDEX			0.153	3.155

TABLE A-19. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Soil Ingestions
 Endpoint: Cancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: White Alice Facility (SS16)
 File: SS16SOCA.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Soil Ingestion Rate	(mg/day)	50	100	200
Exposure Frequency	(days/year)	30	30	30
Exposure Duration	(years)	10	49	6
Conversion Factor	(kg/mg)	0.000001	0.000001	0.000001
Body Weight	(kg)	70	70	15
Averaging Time	(days)	25,550	25,550	25,550

Chemical	Carcinogen Oral Slope Factor	Concentration Soil (mg/kg)	LADD by Receptor Group (mg/kg-day)			Cancer Risk	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult/Child
PCB	7.7	52	4.36e-07	4.27e-06	4.88e-06	3.36e-06	7.05e-05
			CANCER RISK			3e-06	7e-05

TABLE A-20. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Soil Ingestions
 Endpoint: Noncancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: POL Tanks (ST17)
 File: ST17SONC.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Soil Ingestion Rate	(mg/day)	50	100	200
Exposure Frequency	(days/year)	30	30	30
Exposure Duration	(years)	10	49	6
Conversion Factor	(kg/mg)	0.000001	0.000001	0.000001
Body Weight	(kg)	70	70	15
Averaging Time	(days)	3,650	17,885	2,190

Chemical	Oral RfD	Concentration Soil (mg/kg)	ADD by Receptor Group (mg/kg-day)			Hazard Quotients	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult/Child
DRPH	0.08	1,670	9.80e-05	1.96e-04	1.83e-03	1.23e-03	2.53e-02
GRPH	0.2	295	1.73e-05	3.46e-05	3.23e-04	8.66e-05	1.79e-03
			HAZARD INDEX			0.001	0.027

TABLE A-21. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Soil Ingestions
 Endpoint: Cancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: POL Tanks (ST17)
 File: ST17SOCA.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Soil Ingestion Rate	(mg/day)	50	100	200
Exposure Frequency	(days/year)	30	30	30
Exposure Duration	(years)	10	49	6
Conversion Factor	(kg/mg)	0.000001	0.000001	0.000001
Body Weight	(kg)	70	70	15
Averaging Time	(days)	25,550	25,550	25,550

Chemical	Carcinogen Oral Slope Factor	Concentration Soil (mg/kg)	LADD by Receptor Group (mg/kg-day)			Cancer Risk	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult/Child
GRPH	0.0017	295	2.47e-06	2.42e-05	2.77e-05	4.21e-09	8.83e-08
			CANCER RISK			4e-09	9e-08

TABLE A-22. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Soil Ingestions
 Endpoint: Noncancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: Old Dump Site (LF19)
 File: LF19SONC.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Soil Ingestion Rate	(mg/day)	50	100	200
Exposure Frequency	(days/year)	30	30	30
Exposure Duration	(years)	10	49	6
Conversion Factor	(kg/mg)	0.000001	0.000001	0.000001
Body Weight	(kg)	70	70	15
Averaging Time	(days)	3,650	17,885	2,190

Chemical	Oral RfD	Concentration Soil (mg/kg)	ADD by Receptor Group (mg/kg-day)			Hazard Quotients	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult/Child
RRPH	0.08	5,800	3.41e-04	6.81e-04	3.36e-03	4.26e-03	8.80e-02
DRPH	0.08	580	3.41e-05	6.81e-05	6.36e-04	4.26e-04	8.80e-03
			HAZARD INDEX			0.005	0.097

TABLE A-23. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Soil Ingestions
 Endpoint: Noncancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: JP-4 Spill (SS21)
 File: SS21SONC.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Soil Ingestion Rate	(mg/day)	50	100	200
Exposure Frequency	(days/year)	30	30	30
Exposure Duration	(years)	10	49	6
Conversion Factor	(kg/mg)	0.000001	0.000001	0.000001
Body Weight	(kg)	70	70	15
Averaging Time	(days)	3,650	17,885	2,190

Chemical	Oral RfD	Concentration Soil (mg/kg)	ADD by Receptor Group (mg/kg-day)			Hazard Quotients	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult/Child
DRPH	0.08	1,300	7.63e-05	1.53e-04	1.42e-03	9.54e-04	1.97e-02
GRPH	0.2	300	1.76e-05	3.52e-05	3.29e-04	8.81e-05	1.82e-03
			HAZARD INDEX			0.001	0.022

TABLE A-24. DEW LINE INSTALLATION RISK ASSESSMENT SPREADSHEET

Route: Soil Ingestions
 Endpoint: Cancer
 Assumptions: Site-specific
 Installation: Barter Island
 Site: JP-4 Spill (SS21)
 File: SS21SOCA.WK1

Exposure Assumptions		DEW Line Worker	Native Northern Adult	Native Northern Child
Soil Ingestion Rate	(mg/day)	50	100	200
Exposure Frequency	(days/year)	30	30	30
Exposure Duration	(years)	10	49	6
Conversion Factor	(kg/mg)	0.000001	0.000001	0.000001
Body Weight	(kg)	70	70	15
Averaging Time	(days)	25,550	25,550	25,550

Chemical	Carcinogen Oral Slope Factor	Concentration Soil (mg/kg)	LADD by Receptor Group (mg/kg-day)			Cancer Risk	
			DEW Line Worker	Native Northern Adult	Native Northern Child	DEW Line Worker	Native Northern Adult/Child
GRPH	0.0017	300	2.52e-06	2.47e-05	2.82e-05	4.28e-09	8.98e-08
Benzene	0.029	3.9	3.27e-08	3.21e-07	3.66e-07	9.49e-10	1.99e-08
			CANCER RISK			5e-09	1e-07

APPENDIX B
TOXICITY PROFILES

BENZENE

Benzene is readily absorbed following oral and inhalation exposure (EPA 1985). The toxic effects of benzene in humans and other animals following exposure by inhalation include central nervous system effects, hematological effects, and immune system depression. In humans, acute exposures to high concentrations of benzene vapors have been associated with dizziness, nausea, vomiting, headache, drowsiness, narcosis, coma, and death (NAS 1976). Chronic exposure (at least 20 years of worker exposure) to benzene vapors [1-100 ppm 8-hour time-weighted-average (TWA)] can produce reduced leukocyte, platelet, and red blood cell counts (EPA 1993). Benzene induced tumors of the zymbal gland, oral cavity, leukemia and lymphoma in rodents chronically exposed by gavage to doses in the range of 25-500 mg/kg/day (Huff et al. 1989, NTP 1986, Maltoni et al. 1989). Many studies have also described a causal relationship between exposure to benzene by inhalation (either alone or in combination with other chemicals) and leukemia in humans (IARC 1982, Rinsky et al. 1981, Ott et al. 1978, Wong et al. 1983).

Applying EPA's criteria for evaluating the overall evidence of carcinogenicity to humans, benzene is classified in Group A (Human Carcinogen) based on adequate evidence of carcinogenicity from epidemiological studies. EPA (1993) derived an oral cancer slope factor of 2.9×10^{-2} (mg/kg/day)⁻¹ and an inhalation unit risk of 8.3×10^{-6} (ug/m³)⁻¹ for benzene. These values were based on several studies in which increased incidences of nonlymphocytic leukemia were observed in humans occupationally exposed to benzene principally by inhalation (Rinsky et al. 1981, Ott et al. 1978, Wong et al. 1983). Equal weight was given to cumulative dose and weighted cumulative dose as well as to relative and absolute risk model forms (EPA 1993). EPA (1993) is currently reviewing both oral and inhalation RfDs for benzene, for which the status is pending.

The National Research Council's Committee on Toxicology has set a one-hour Emergency Exposure Guidance Level (EEGL), for benzene at 50 ppm (200 mg/m³) (NRC 1986). Formerly known as just EEL, the EEGL is defined as a ceiling limit for an unpredicted single exposure lasting one to 24 hours whose occurrence is expected to be rare in the lifetime of any person. It is designed to avoid substantial decrements in performance during emergencies and takes into account the statistical likelihood of a non-incapacitative, reversible effect in exposed populations (NRC 1986). A health criterion for acute inhalation exposure to benzene of 20 mg/m³ can be derived from the EEGL by combining it with a safety factor of 10 to account for the healthy worker effect which assumes employed persons are generally healthier than the general population.

Benzene References

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BIS(2-ETHYLHEXYL)PHTHALATE

Bis(2-ethylhexyl)phthalate, also known as di-ethylhexyl phthalate (DEHP), is readily absorbed following oral or inhalation exposure (EPA 1980). Acute oral exposure (600 - 2,000 mg/kg/day DEHP) to rodents, guinea pigs and monkeys has resulted in adverse effects of the livers and kidney (weight changes and enzyme induction) (ATSDR 1991). Chronic exposure to relatively high concentrations (200 mg/kg/day) of DEHP in the diet can cause retardation of growth and increased liver and kidney weights in laboratory animals (NTP 1982, EPA 1980, Carpenter et al. 1953). Effects on the liver at the cellular level were noted in rats at doses as low as 10 to 50 mg/kg/day DEHP (Ganning et al. 1989, Mitchell et al. 1985, Short et al. 1987) while dogs receiving doses of 59 mg/kg/day DEHP for one year had no observed changes in liver weight or structure (Carpenter et al. 1953, ATSDR 1991). Single oral doses of 4,882 and 9,756 mg/kg DEHP administered to pregnant rats on day 12 of gestation caused a dose-related increase in dead and resorbed fetuses and a number of malformations in the survivors (Ritter et al. 1989). DEHP is lipophilic and has the potential to be transported in maternal milk and thus have an impact on postnatal development (ATSDR 1991). Studies in rodents exposed to doses in the range of 200 to 2,800 mg/kg/day DEHP indicate that the testes are a primary target tissue resulting in increased testicular weights, tubular atrophy, decreased male fertility and abnormal sperm (ATSDR 1991). Several chronic feeding studies in rodents indicate that lifetime exposure to 300 to 1,000 mg/kg/day DEHP can cause liver tumors in rats and mice (Kluwe et al. 1982, Rao et al. 1987, 1990).

EPA (1993a) classified DEHP in Group B2--Probable Human Carcinogen. EPA (1993) calculated an oral cancer slope factor for DEHP of $1.4 \times 10^{-2} \text{ (mg/kg/day)}^{-1}$ based on data from the NTP (1982) study in which liver tumors were noted in mice. EPA recommended an oral reference dose (RfD) for DEHP of $2 \times 10^{-2} \text{ mg/kg/day}$ for both chronic (EPA 1993a) and subchronic (EPA 1993b) exposures based on a study by Carpenter et al. (1953) in which increased liver weight was observed in female guinea pigs exposed to 19 mg/kg bw/day in the diet for 1 year; an uncertainty factor of 1,000 was used to develop both RfDs.

Bis(2-ethylhexyl)phthalate References

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BERYLLIUM

Beryllium is not readily absorbed by any route of exposure. Occupational exposure to beryllium results in bone, liver, and kidney depositions (EPA 1986). In humans, acute respiratory effects due to beryllium exposure include rhinitis, pharyngitis, tracheobronchitis, and acute pneumonitis. Dermal exposure to soluble beryllium compounds can cause contact dermatitis, ulceration, and granulomas (Hammond and Beliles 1980). Ocular effects include conjunctivitis and corneal ulceration from splash burns. The most common clinical symptom caused by chronic beryllium exposure is granulomatous lung inflammation (IARC 1980, EPA 1986). Chronic skin lesions sometimes appear after a long latent period in conjunction with the pulmonary effects. Systemic effects from beryllium exposure may include right heart enlargement with accompanying cardiac failure, liver and spleen enlargement, cyanosis, digital clubbing, and kidney stone development (EPA 1986, Schroeder and Mitchner 1975). Beryllium has been shown to be carcinogenic in experimental animals resulting primarily in lung and/or bone tumors when given by injection, intratracheal administration, or inhalation (EPA 1986). Following lifetime exposure in drinking water, slight increases in the incidence of gross tumors of all sites combined were observed in rats (Schroeder and Mitchner 1975). Several epidemiological studies have suggested that occupational exposure to beryllium may result in an increased lung cancer risk although the data are inconclusive (EPA 1986, Wagoner et al. 1980).

EPA (1993a) classified beryllium in Group B2--(Probable Human Carcinogen) based on increased incidences of lung cancer and osteosarcomas in animals. EPA (1993) calculated an inhalation cancer unit risk of $2.4 \times 10^{-3} \text{ (ug/m}^3\text{)}^{-1}$ based on the relative risk for lung cancer, estimated from an epidemiological study by Wagoner et al. (1980). EPA (1993) established an oral cancer slope factor of $4.3 \text{ (mg/kg/day)}^{-1}$ based on the induction of gross tumors (all sites combined) in rats chronically administered beryllium sulfate in their drinking water (Schroeder and Mitchner 1975). EPA (1993) also developed an oral reference dose (RfD) for beryllium of $5 \times 10^{-3} \text{ mg/kg/day}$ based on a study by Schroeder and Mitchner (1975) in which rats exposed to 0.54 mg/kg/day beryllium sulfate (the highest dose tested) in drinking water for a lifetime did not exhibit adverse effects. An uncertainty factor of 100 was used to develop the RfD.

Beryllium References

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TRICHLOROETHENE

Absorption of trichloroethene (TCE) from the gastrointestinal tract is virtually complete. Absorption following inhalation exposure is proportional to concentration and duration of exposure (EPA 1985). TCE is a central nervous system depressant following acute and chronic exposures. In humans, single oral doses of 15 to 25 ml (21 to 35 grams) of TCE have resulted in vomiting and abdominal pain, followed by transient unconsciousness (Stephens 1945). High-level exposure can result in death due to respiratory and cardiac failure (EPA 1985). Hepatotoxicity has been reported in human and animal studies following acute exposure to TCE (EPA 1985). Nephrotoxicity has been observed in animals following acute exposure to TCE vapors (ACGIH 1986, Torkelson and Rowe 1981). Subacute inhalation exposures of mice have resulted in transient increased liver weights (Kjellstrand et al. 1983a,b). Industrial use of TCE is often associated with adverse dermatological effects including reddening and skin burns on contact with the liquid form, and dermatitis resulting from vapors. These effects are usually the result of contact with concentrated solvent, however, and no effects have been reported following exposure to TCE in dilute, aqueous solutions (EPA 1985). TCE has caused significant increases in the incidence of hepatocellular carcinomas in mice (NCI 1976) and renal tubular-cell neoplasms in rats exposed by gavage (NTP 1983), and pulmonary adenocarcinomas in mice following inhalation exposure (Fukuda et al. 1983, Maltoni et al. 1986). TCE was mutagenic in *Salmonella typhimurium* and in *E. coli* (strain K-12), utilizing liver microsomes for activation (Greim et al. 1977).

EPA is currently reviewing the carcinogenicity of TCE. The EPA Environmental Criteria and Assessment Office (ECAO) currently classifies TCE as a Group B2/C--Probable/Possible Human Carcinogen based on inadequate evidence in humans and sufficient evidence of carcinogenicity from animal studies. ECAO (1992) reported an oral cancer potency factor of 1.1×10^{-2} (mg/kg/day)⁻¹ based on two gavage studies conducted in mice in which an increased incidence of liver tumors were observed (Maltoni et al. 1986, Fukuda et al. 1983). An inhalation cancer unit risk of 1.7×10^{-6} (μg/m³)⁻¹ has been derived for TCE based on an increased incidence of lung tumors in mice exposed via inhalation (ECAO 1992, NCI 1976). The cancer estimates are currently under review by EPA. EPA (1987) developed an oral reference dose (RfD) of 7.35×10^{-3} mg/kg/day based on a subchronic inhalation study in rats in which elevated liver weights were observed following exposure to 55 ppm, 5 days/week for 14 weeks (Kimmerle and Eben 1973). A safety factor of 1,000 was used to calculate the RfD. However, this RfD is currently under review by EPA.

The National Research Council's Committee on Toxicology has set a one-hour Emergency Exposure Guidance Level (EEGL), for trichloroethene at 200 ppm (1,000 mg/m³) (NRC 1988). Formerly known as just EEL, the EEGL is defined as a ceiling limit for an unpredicted single exposure lasting one to 24 hours whose occurrence is expected to be rare in the lifetime of any person. It is designed to avoid substantial decrements in performance during emergencies and takes into account the statistical likelihood of a non-incapacitative, reversible effect in exposed populations (NRC 1988). A health criterion for acute inhalation exposure to trichloroethene of 100 mg/m³ can be derived from the EEGL by combining it with a safety factor of 10 to account for the healthy worker effect which assumes employed persons are generally healthier than the general population.

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POLYCHLORINATED BIPHENYLS (PCBs, including Aroclor 1254)

PCBs are complex mixtures of chlorinated biphenyls. The commercial PCB mixtures that were manufactured in the United States were given the trade name of "Aroclor." Aroclors are distinguished by a four-digit number (for example, Aroclor 1260). The last two digits in the Aroclor 1200 series represent the average percentage by weight of chlorine in the product.

PCBs are readily and extensively absorbed through the gastrointestinal tract and somewhat less readily through the skin; PCBs are presumably readily absorbed from the lungs, but few data are available that experimentally define the extent of absorption after inhalation (EPA 1985). Studies have found oral efficiency on the order of 75 to > 90% in rats, monkeys and ferrets (Albro and Fishbein 1972, Allen et al. 1974, Tanabe et al. 1981, Bleavens et al. 1984, Clevenger et al. 1989). PCBs distribute preferentially to adipose tissue and concentrate in human breast milk due to its high fat content (ATSDR 1991). Dermatitis and chloracne (a disfiguring and long-term skin disease) have been the most prominent and consistent findings in studies of occupational exposure to PCBs. Several studies examining liver function in exposed humans have reported disturbances in blood levels of liver enzymes. Reduced birth weights, slow weight gain, reduced gestational ages, and behavioral deficits in infants were reported in a study of women who had consumed PCB-contaminated fish from Lake Michigan (EPA 1985). Reproductive, hepatic, immunotoxic, and immunosuppressive effects appear to be the most sensitive end points of PCB toxicity in nonrodent species, and the liver appears to be the most sensitive target organ for toxicity in rodents (EPA 1985). For example, adult monkeys exposed to dietary concentrations of 0.028 mg/kg-day Aroclor 1016 for approximately 22 months showed no evidence of overt toxicity; however, the offspring of these monkeys exhibited decreased birth weight and possible neurological impairment (Barsotti and Van Miller 1984, Levin et al. 1988, Schantz et al. 1989, 1991). A number of studies have suggested that PCB mixtures are capable of increasing the frequency of tumors including liver tumors in animals exposed to the mixtures for long periods (Kimbrough et al. 1975, NCI 1978, Schaeffer et al. 1984, Norback and Weltman 1985). Studies have suggested that PCB mixtures can act to promote or inhibit the action of other carcinogens in rats and mice (EPA 1985). It is known that PCB congeners vary greatly in their potency in producing biological effects, such as cancer however, EPA (1993) generally considers Aroclor 1260 to be representative of all PCB mixtures for the evaluation of carcinogenic effects. There is some evidence that mixtures containing highly chlorinated biphenyls are more potent inducers of hepatocellular carcinoma in rats than are mixtures containing less chlorine by weight (EPA 1993).

EPA (1993) classified PCBs as a Group B2 agent (Probable Human Carcinogen) based on sufficient evidence in animal bioassays and inadequate evidence from studies in humans. The EPA (1993) calculated an oral cancer potency factor of $7.7 \text{ (mg/kg/day)}^{-1}$ for PCBs based on the incidence of hepatocellular carcinomas (91%) and neoplastic nodules (4%) in female Sprague-Dawley rats exposed to a diet containing Aroclor 1260 as reported in a study by Norback and Weltman (1985). In the same study, males exhibited a much lower incidence of malignant tumors but a higher incidence of benign tumors (neoplastic nodules). EPA (1993) also calculated a slope factor of $5.7 \text{ (mg/kg/day)}^{-1}$ for malignant tumors alone, which is supported by a risk estimate based on the data of Kimbrough et al. (1975). EPA (1993) derived an oral RfD of $7 \times 10^{-5} \text{ mg/kg/day}$ for Aroclor 1016 based on a 21.8 month oral study conducted in monkeys (Barsotti and Van Miller 1984, Levin et al. 1988, Schantz et al. 1989, 1991). A no-observed-adverse-effect level of 0.25 ppm (0.007 mg/kg/day) for decreased birth weight in offspring was identified from these studies. A safety factor of 100 (3 to account for interspecies extrapolation,

3 to account for sensitive individuals, 3 to account for limitations in the database, and 3 to account for extrapolation from a subchronic to a chronic RfD) was used to calculate the RfD.

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TETRACHLOROETHENE

Tetrachloroethene is absorbed following inhalation (IARC 1979) and oral (EPA 1985a,b) exposure. Tetrachloroethene vapors and liquid also can be absorbed through the skin (EPA 1985a,b). The principal toxic effects of tetrachloroethene in humans and animals following acute and longer-term exposures include central nervous system (CNS) depression and fatty infiltration of the liver and kidney with concomitant changes in serum enzyme activity levels indicative of tissue damage (EPA 1985a,b, Buben and O'Flaherty 1985). Humans exposed to doses of between 136 and 1,018 mg/m³ for 5 weeks develop central nervous system effects, such as lassitude and signs of inebriation (Stewart et al. 1974). The offspring of female rats and mice exposed to high concentrations of tetrachloroethene for 7 hours daily on days 6-15 of gestation developed toxic effects, including a decrease in fetal body weight in mice and a small but significant increase in fetal resorption in rats (Schwetz et al. 1975). Mice also exhibited developmental effects, including subcutaneous edema and delayed ossification of skull bones and sternebrae (Schwetz et al. 1975). In a National Cancer Institute bioassay (NCI 1977), increased incidence of hepatocellular carcinoma were observed in both sexes of B6C3F1 mice administered tetrachloroethylene in corn oil by gavage for 78 weeks. Increased incidence of mononuclear cell leukemia and renal adenomas and carcinomas (combined) have also been observed in long term bioassays in which rats were exposed to tetrachloroethene by inhalation (NTP 1986).

Tetrachloroethene is currently under review by the Carcinogen Risk Assessment Verification Endeavor (CRAVE) and estimates of cancer potency were recently withdrawn by EPA (1992b). However, the EPA Environmental Criteria and Assessment Office (ECAO) (1992a) currently classifies tetrachloroethene as a Group B2/C carcinogen (Probable/Possible Human Carcinogen). ECAO (1992a) has reported an oral slope factor of 5.2×10^{-2} (mg/kg/day)⁻¹ based on liver tumors observed in the NCI (1977) gavage bioassay for mice. An inhalation cancer unit risk of 5.8×10^{-7} (μg/m³)⁻¹ is based on an NTP (1986) bioassay in rats and mice in which leukemia and liver tumors were observed (ECAO 1992a). Both the cancer slope factor and unit risk are currently under review by EPA. EPA (1993) also derived an oral reference dose (RfD) of 1×10^{-2} mg/kg/day for tetrachloroethene based on a 6-week gavage study by Buben and O'Flaherty (1985). In this study, liver weight/body weight ratios were significantly increased in mice and rats treated with 71 mg/kg-day tetrachloroethene but not in animals treated with 14 mg/kg-day. Using a NOAEL of 14 mg/kg/day and applying an uncertainty factor of 1,000 the RfD was derived. EPA (1992b) established a subchronic oral RfD of 1×10^{-1} mg/kg/day, using an uncertainty factor of 100 and based on the same study and effect of concern.

The American Conference of Governmental Industrial Hygienists (ACGIH) has set a Short-Term Exposure Level -- Threshold Limit Value of 200 ppm (1,000 mg/m³) for tetrachloroethene (ACGIH 1991). The STEL-TLV is defined as a 15-minute time-weighted average which should not be exceeded at any time during a work day. A health criterion for acute inhalation exposure to tetrachloroethene of 100 mg/m³ can be derived from the STEL-TLV by combining it with a safety factor of 10 to account for the healthy worker effect which assumes that employed persons are generally healthier than the general population.

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MANGANESE

Manganese is considered to be among the least toxic of the trace metals and, in fact, is considered to be an essential element (NRC 1989). The oral absorption of dietary manganese ranges from 3 to 10% (EPA 1993). However, manganese is absorbed to a greater extent following inhalation exposures. The National Research Council has established a provisional recommended dietary allowance for adults of 2 to 5 mg/day (NRC 1989). The effects following acute exposure to manganese are unknown.

Chronic occupational exposure to manganese dust ($0.02 - 2.6 \text{ mg/m}^3$) has been associated with respiratory symptoms and pneumonitis (Chandra et al. 1981, 1990) and higher levels have been associated with a condition known as manganism, a progressive neurological disease characterized by speech disturbances, tremors, and difficulties in walking. For example, male workers exposed to manganese dioxide, tetroxide and various salts [time-weighted-average (TWA) of total airborne manganese dust ranged from $0.07 - 8.61 \text{ mg/m}^3$] experienced an increased incidence of psychomotor disturbances (e.g., reaction time, hand-eye coordination and hand steadiness) (Roels et al. 1987). Other effects observed in humans occupationally exposed to manganese dust include hematological (Chandra et al. 1981, Flinn et al. 1941, Kesic and Hausler 1954), cardiovascular (Saric and Hrustic 1975) and reproductive effects (Cook et al. 1974, Emara et al 1971, Lauwerys et al 1985, Rodier 1955).

In adults, a safe intake of manganese from dietary sources ranges from 2-10 mg/day (10 mg/day = 0.14 mg/kg/day) (WHO 1973, NRC 1989, Schroeder et al. 1966). Individuals who chronically ingested drinking water from natural wells containing manganese concentrations of 1,600 to 2,300 $\mu\text{g/L}$ (0.06 mg/kg/day), showed a statistically significant increase in minor neurologic effects (neurologic exam scores) (Kondakis et al. 1989). Higher concentrations in drinking water (0.8 mg/kg/day) have resulted in symptoms including lethargy, increased muscle tonus, tremor and mental disturbances (Kawamura et al. 1941).

The apparent differences in manganese toxicity following dietary and drinking water exposures can be attributed to the greater bioavailability of manganese from water (EPA 1993). Chronic oral exposure of rats to manganese chloride can also result in central nervous system dysfunction (Leung et al. 1981, Lai et al. 1982). Chronic inhalation exposure of experimental animals (monkeys, rats, mice, hamsters) has resulted in respiratory effects; however, other studies have demonstrated that these effects may be immunological in origin (ATSDR 1992).

Manganese has not been reported to be teratogenic; however, this metal has been observed to cause depressed reproductive performance and reduced fertility in humans and experimental animals (EPA 1984a). Certain manganese compounds have been shown to be mutagenic in a variety of bacterial tests. Manganese chloride and potassium permanganate can cause chromosomal aberrations in mouse mammary carcinomal cells. Manganese was moderately effective in enhancing viral transformation of Syrian hamster embryo cells (EPA 1984a,b).

EPA (1993a) established a weight-of-evidence classification for manganese of D (not classifiable as to human carcinogenicity). EPA (1993a) derived two separate oral reference doses (RfD). The separate RfDs for food and water indicate a potentially higher bioavailability of manganese from drinking water than from the diet. The RfD associated with oral exposure to drinking water is $5 \times 10^{-3} \text{ mg/kg/day}$ based on a no-observed-adverse-effect-level (NOAEL) of $5 \times 10^{-3} \text{ mg/kg/day}$ for humans (Kondakis et al. 1989). EPA (1993a) also derived an RfD of $1.4 \times 10^{-1} \text{ mg/kg/day}$ for

manganese in food based on a NOAEL of 0.14 mg/kg/day (10 mg/day) in humans chronically exposed to dietary levels (WHO 1973, Schroeder et al. 1966, NRC 1989). The effect of concern was the central nervous system, and an uncertainty factor of one was used to derive both RfDs. The chronic RfD in food was adopted as the subchronic RfD (EPA 1993b). EPA (1993a) derived a chronic inhalation reference concentration (RfC) of 4×10^{-4} mg/m³ based upon an occupational study conducted by Roels et al. (1987) in which respiratory symptoms and psychomotor disturbances were observed. EPA (1993b) adopted the chronic RfC as the subchronic RfC. An uncertainty factor of 900 was used to derive both RfCs.

Manganese References

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APPENDIX C

ESTIMATES EXPOSURE CALCULATIONS FOR ECOLOGICAL RECEPTORS

	Food Concentration		Water Concentration		Soil Concentration		% ingested		Body		Dietary			
	Intak Diet (FI) (kg)	(CD) (mg/kg-bw (A)	(FIxC (WI) (CW) (1/d (mg/l)	(WIXCW (SI) (CS) (B) (kg)	(SIxC (A+B+C (IS) (C) (mg/kg)	on-site (D*I (BW) (E) (kg)	Weight (E/BW=DI) (mg/kg-bw							
ESTIMATED EXPOSURE														
DRPH														
L. longspur	0.01	8.745288	0.06	0	0.821	0.004	0	1041	0.15	0.208	0.5	0.1	0.03	3.843776
brant	0.07	31.50745	2.18	0.1	0.821	0.057	0	1041	5.93	8.171	0.18	1.5	1.31	1.127101
glaucous gull	0.07	3.495896	0.26	0.1	0.821	0.066	0	1041	5.85	6.174	0.3	1.9	1.45	1.281892
pect. sandpiper	0.01	3.495896	0.04	0	0.821	0.008	0	1041	1.5	1.547	0.78	1.2	0.08	15.27512
peregrine falcon	0.05	0	0	0.1	0.821	0.041	0	1041	1.06	1.103	0.1	0.1	0.82	0.134496
spectacled eider	0.07	8.745288	0.63	0.1	0.821	0.057	0	1041	6.14	6.826	0.01	0.1	1.38	0.04964
Steller's eider	0.05	8.745288	0.44	0.1	0.821	0.041	0	1041	4.35	4.836	0.01	0	0.81	0.060076
brown lemming	0.05	0.037334	0	0	0.821	0.006	0	1041	1.25	1.257	0.5	0.6	0.06	11.42388
arctic fox	0.26	0	0	0.4	0.821	0.345	0	1041	7.5	7.84	0.01	0.1	4.95	0.015838
ETHYLBENZENE														
L. longspur	0.01	0.0373	0	0	0	0	0	0.255	0	3E-04	0.5	0	0.03	0.00522
brant	0.07	0.1343	0.01	0.1	0	0	0	0.255	0	0.011	0.18	0	1.31	0.001482
glaucous gull	0.07	0.0149	0	0.1	0	0	0	0.255	0	0.003	0.3	0	1.45	0.000526
pect. sandpiper	0.01	0.0149	0	0	0	0	0	0.255	0	5E-04	0.78	0	0.08	0.005261
peregrine falcon	0.05	0	0	0.1	0	0	0	0.255	0	3E-04	0.1	0	0.82	3.17E-05
spectacled eider	0.07	0.0373	0	0.1	0	0	0	0.255	0	0.004	0.01	0	1.38	3.04E-05
Steller's eider	0.05	0.0373	0	0.1	0	0	0	0.255	0	0.003	0.01	0	0.81	3.66E-05
brown lemming	0.05	0.14922	0.01	0	0	0	0	0.255	0	0.007	0.5	0	0.06	0.063826
arctic fox	0.26	0	0	0.4	0	0	0	0.255	0	0.002	0.01	0	4.95	3.71E-06
TRIMETHYLBENZENE														
L. longspur	0.01	0.26565	0	0	0	0	0	4.2	0	0.002	0.5	0	0.03	0.043357
brant	0.07	0.95634	0.07	0.1	0	0	0	4.2	0.02	0.09	0.18	0	1.31	0.01243
glaucous gull	0.07	0.10626	0.01	0.1	0	0	0	4.2	0.02	0.031	0.3	0	1.45	0.006531
pect. sandpiper	0.01	0.10626	0	0	0	0	0	4.2	0.01	0.007	0.78	0	0.08	0.071401
peregrine falcon	0.05	0	0	0.1	0	0	0	4.2	0	0.004	0.1	0	0.82	0.000522

spectacled eider	0.07	0.26565	0.02	0.1	0	0	0	0	4.2	0.02	0.044	0.01	0	1.38	0.000319
Steller's eider	0.05	0.26565	0.01	0.1	0	0	0	0	4.2	0.02	0.031	0.01	0	0.81	0.000385
brown lemming	0.05	1.0627	0.05	0	0	0	0	0	4.2	0.01	0.053	0.5	0	0.06	0.480559
arctic fox	0.26	0	0	0.4	0	0	0	0	4.2	0.03	0.03	0.01	0	4.95	6.11E-05
XYLENES															
L. longspur	0.01	0.15447	0	0	0	0	0	0	1.07	0	0.001	0.5	0	0.03	0.021654
brant	0.07	0.55609	0.04	0.1	0	0	0	0	1.07	0.01	0.045	0.18	0	1.31	0.006149
glaucous gull	0.07	0.06179	0	0.1	0	0	0	0	1.07	0.01	0.011	0.3	0	1.45	0.002196
pect. sandpiper	0.01	0.06179	0	0	0	0	0	0	1.07	0	0.002	0.78	0	0.08	0.021995
peregrine falcon	0.05	0	0	0.1	0	0	0	0	1.07	0	0.001	0.1	0	0.82	0.000133
spectacled eider	0.07	0.15447	0.01	0.1	0	0	0	0	1.07	0.01	0.017	0.01	0	1.38	0.000126
Steller's eider	0.05	0.15447	0.01	0.1	0	0	0	0	1.07	0	0.012	0.01	0	0.81	0.000152
brown lemming	0.05	0.6179	0.03	0	0	0	0	0	1.07	0	0.029	0.5	0	0.06	0.26445
arctic fox	0.26	0	0	0.4	0	0	0	0	1.07	0.01	0.008	0.01	0	4.95	1.56E-05
NAPHTHALENE															
L. longspur	0.01	0.553125	0	0	0	0	0	0	5	0	0.004	0.5	0	0.03	0.080567
brant	0.07	1.99125	0.14	0.1	0	0	0	0	5	0.03	0.166	0.18	0	1.31	0.022937
glaucous gull	0.07	0.22125	0.02	0.1	0	0	0	0	5	0.03	0.044	0.3	0	1.45	0.009228
pect. sandpiper	0.01	0.22125	0	0	0	0	0	0	5	0.01	0.01	0.78	0	0.08	0.095386
peregrine falcon	0.05	0	0	0.1	0	0	0	0	5	0.01	0.005	0.1	0	0.82	0.000622
spectacled eider	0.07	0.553125	0.04	0.1	0	0	0	0	5	0.03	0.069	0.01	0	1.38	0.000503
Steller's eider	0.05	0.553125	0.03	0.1	0	0	0	0	5	0.02	0.049	0.01	0	0.81	0.000607
brown lemming	0.05	2.2125	0.1	0	0	0	0	0	5	0.01	0.106	0.5	0.1	0.06	0.959659
arctic fox	0.26	0	0	0.4	0	0	0	0	5	0.04	0.036	0.01	0	4.95	7.27E-05
2-METHYLNAPHTHALENE															
L. longspur	0.01	0.10774	0	0	0	0	0	0	1.92	0	1E-03	0.5	0	0.03	0.018146
brant	0.07	0.38787	0.03	0.1	0	0	0	0	1.92	0.01	0.038	0.18	0	1.31	0.005212
glaucous gull	0.07	0.0431	0	0.1	0	0	0	0	1.92	0.01	0.014	0.3	0	1.45	0.002901
pect. sandpiper	0.01	0.0431	0	0	0	0	0	0	1.92	0	0.003	0.78	0	0.08	0.032041

peregrine falcon	0.05	0	0	0.1	0	0	0	1.92	0	0.002	0.1	0	0.82	0.000239
spectacled eider	0.07	0.10774	0.01	0.1	0	0	0	1.92	0.01	0.019	0.01	0	1.38	0.000138
Steller's eider	0.05	0.10774	0.01	0.1	0	0	0	1.92	0.01	0.013	0.01	0	0.81	0.000167
brown lemming	0.05	0.431	0.02	0	0	0	0	1.92	0	0.022	0.5	0	0.06	0.197264
arctic fox	0.26	0	0	0.4	0	0	0	1.92	0.01	0.014	0.01	0	4.95	2.79E-05

PCBs

L. longspur	0.01	4.62E-05	0	0	0	0	0	0.49	0	7E-05	0.5	0	0.03	0.001276
brant	0.07	0.001664	0	0.1	0	0	0	0.49	0	0.003	0.18	0	1.31	0.000401
glaucous gull	0.07	0.000185	0	0.1	0	0	0	0.49	0	0.003	0.3	0	1.45	0.000575
pect. sandpiper	0.01	0.000185	0	0	0	0	0	0.49	0	7E-04	0.78	0	0.08	0.006992
peregrine falcon	0.05	0	0	0.1	0	0	0	0.49	0	5E-04	0.1	0	0.82	6.1E-05
spectacled eider	0.07	0.000462	0	0.1	0	0	0	0.49	0	0.003	0.01	0	1.38	2.13E-05
Steller's eider	0.05	0.000462	0	0.1	0	0	0	0.49	0	0.002	0.01	0	0.81	2.57E-05
brown lemming	0.05	0.001849	0	0	0	0	0	0.49	0	7E-04	0.5	0	0.06	0.006102
arctic fox	0.26	0	0	0.4	0	0	0	0.49	0	0.004	0.01	0	4.95	7.13E-06

ALUMINUM

L. longspur	0.01	0	0	0	0.223	0.001	0	0	0	0.001	0.5	0	0.03	0.020648
brant	0.07	0	0	0.1	0.223	0.016	0	0	0	0.016	0.18	0	1.31	0.002153
glaucous gull	0.07	0	0	0.1	0.223	0.018	0	0	0	0.018	0.3	0	1.45	0.003704
pect. sandpiper	0.01	0	0	0	0.223	0.002	0	0	0	0.002	0.78	0	0.08	0.022018
peregrine falcon	0.05	0	0	0.1	0.223	0.011	0	0	0	0.011	0.1	0	0.82	0.00136
spectacled eider	0.07	0	0	0.1	0.223	0.016	0	0	0	0.016	0.01	0	1.38	0.000114
Steller's eider	0.05	0	0	0.1	0.223	0.011	0	0	0	0.011	0.01	0	0.81	0.000139
brown lemming	0.05	0	0	0	0.223	0.002	0	0	0	0.002	0.5	0	0.06	0.014191
arctic fox	0.26	0	0	0.4	0.223	0.094	0	0	0	0.094	0.01	0	4.95	0.000189

IRON

L. longspur	0.01	0	0	0	8.45	0.042	0	0	0	0.042	0.5	0	0.03	0.782407
brant	0.07	0	0	0.1	8.45	0.592	0	0	0	0.592	0.18	0.1	1.31	0.081586
glaucous gull	0.07	0	0	0.1	8.45	0.676	0	0	0	0.676	0.3	0.2	1.45	0.140346
pect. sandpiper	0.01	0	0	0	8.45	0.085	0	0	0	0.085	0.78	0.1	0.08	0.834304

peregrine falcon	0.05	0	0	0.1	8.45	0.423	0	0	0	0.423	0.1	0	0.82	0.051524
spectacled eider	0.07	0	0	0.1	8.45	0.592	0	0	0	0.592	0.01	0	1.38	0.004302
Steller's eider	0.05	0	0	0.1	8.45	0.423	0	0	0	0.423	0.01	0	0.81	0.005248
brown lemming	0.05	0	0	0	8.45	0.059	0	0	0	0.059	0.5	0	0.06	0.537727
arctic fox	0.26	0	0	0.4	8.45	3.549	0	0	0	3.549	0.01	0	4.95	0.00717
LEAD														
L. longspur	0.01	0.29812	0	0	0	0	0	26.5	0	0.006	0.5	0	0.03	0.105141
brant	0.07	1.1925	0.08	0.1	0	0	0	26.5	0.15	0.234	0.18	0	1.31	0.032217
glaucous gull	0.07	0.1192	0.01	0.1	0	0	0	26.5	0.15	0.158	0.3	0	1.45	0.032749
pect. sandpiper	0.01	0.1192	0	0	0	0	0	26.5	0.04	0.04	0.78	0	0.08	0.390095
peregrine falcon	0.05	0	0	0.1	0	0	0	26.5	0.03	0.027	0.1	0	0.82	0.003296
spectacled eider	0.07	0.29812	0.02	0.1	0	0	0	26.5	0.16	0.178	0.01	0	1.38	0.001292
Steller's eider	0.05	0.29812	0.02	0.1	0	0	0	26.5	0.11	0.126	0.01	0	0.81	0.001564
brown lemming	0.05	1.1925	0.05	0	0	0	0	26.5	0.03	0.085	0.5	0	0.06	0.776932
arctic fox	0.26	0	0	0.4	0	0	0	26.5	0.19	0.191	0.01	0	4.95	0.000385
MANGANESE														
L. longspur	0.01	0	0	0	0.574	0.003	0	0	0	0.003	0.5	0	0.03	0.053148
brant	0.07	0	0	0.1	0.574	0.04	0	0	0	0.04	0.18	0	1.31	0.005542
glaucous gull	0.07	0	0	0.1	0.574	0.046	0	0	0	0.046	0.3	0	1.45	0.009534
pect. sandpiper	0.01	0	0	0	0.574	0.006	0	0	0	0.006	0.78	0	0.08	0.056673
peregrine falcon	0.05	0	0	0.1	0.574	0.029	0	0	0	0.029	0.1	0	0.82	0.0035
spectacled eider	0.07	0	0	0.1	0.574	0.04	0	0	0	0.04	0.01	0	1.38	0.000292
Steller's eider	0.05	0	0	0.1	0.574	0.029	0	0	0	0.029	0.01	0	0.81	0.000357
brown lemming	0.05	0	0	0	0.574	0.004	0	0	0	0.004	0.5	0	0.06	0.036527
arctic fox	0.26	0	0	0.4	0.574	0.241	0	0	0	0.241	0.01	0	4.95	0.000487
ZINC														
L. longspur	0.01	28.987	0.19	0	0	0	0	77.3	0.01	0.202	0.5	0.1	0.03	3.743263
brant	0.07	104.36	7.22	0.1	0	0	0	77.3	0.44	7.662	0.18	1.4	1.31	1.056872
glaucous gull	0.07	11.595	0.86	0.1	0	0	0	77.3	0.43	1.291	0.3	0.4	1.45	0.268089

pect. sandpiper	0.01	11.595	0.13	0	0	0	0	0	77.3	0.11	0.24	0.78	0.2	0.08	2.370546
peregrine falcon	0.05	0	0	0.1	0	0	0	0	77.3	0.08	0.079	0.1	0	0.82	0.009615
spectacled eider	0.07	28.987	2.08	0.1	0	0	0	0	77.3	0.46	2.532	0.01	0	1.38	0.018411
Steller's eider	0.05	28.987	1.46	0.1	0	0	0	0	77.3	0.32	1.787	0.01	0	0.81	0.0222
brown lemming	0.05	115.95	5.22	0	0	0	0	0	77.3	0.09	5.311	0.5	2.7	0.06	48.27736
arctic fox	0.26	0	0	0.4	0	0	0	0	77.3	0.56	0.557	0.01	0	4.95	0.001124

APPENDIX D

REMEDIAL INVESTIGATION ANALYTICAL DATA

TABLE D-1. BACKGROUND ANALYTICAL DATA SUMMARY

Installation: Barter Island Site: Background (BKGD)		Matrix: Soil/Sediment Units: mg/kg		Environmental Samples						Field Blanks		Lab Blanks	
Parameters	Detect Limits	Quant. Limits	Action Levels	Bkgd. Range	S01 & S04 (Replicates)	S02	S03	SD01	SD02	AB01	EB01	TB02	
Laboratory Sample ID Numbers					4203-5	4203-4	4175-2 4203-6	4199-4	4199-1 4179-4	4197-6 4173-9	4175-3 4203-8	4179-5 4199-13	4203 4179 4175 4175
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	µg/L	µg/L	µg/L	mg/kg
DRPH	4.00	4.00	500 ^a	9.55-1,150	588 ^{c,d}	466 ^{c,d}	384J ^{c,d}	116 ^c	9.55J ^c	NA	<200	NA	<4.00
GRPH	0.400	0.400-9	100	<0.400-9	<1.40	<1.20	<0.800	<6.0	<0.400	NA	<20	NA	<0.400
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-1,500	<0.325	<0.750	<0.250	<1,500	<0.100	<1,250			
Benzene	0.020	0.020-0.300	0.5	<0.020-0.300	<0.065	<0.150	<0.050	<0.300	<0.020	<0.250	<1	<1	<0.020
Toluene	0.020	0.020-0.300		<0.020-0.300	<0.085	<0.150	<0.050	<0.300	<0.020	<0.250	1.2	<1	<0.020
Ethylbenzene	0.020	0.020-0.300		<0.020-0.300	<0.065	<0.150	<0.050	<0.300	<0.020	<0.250	<1	<1	<0.020
Xylenes (Total)	0.040	0.040-0.600		<0.040-0.600	<0.130	<0.300	<0.100	<0.600	<0.040	<0.500	<2	<2	<0.040
VOC 8010	0.020	0.020-0.300		<0.020-0.300	<0.065	<0.150	<0.050	<0.300	<0.020	<0.300	<1-9.8	<1-2.5	<0.020
VOC 8280	0.020	0.025-0.050		<0.025-0.500	NA	NA	<0.050	NA	<0.025	NA	<1-1.6	<1J-4.4J	<0.020
SVOC 8270	0.200	0.230-3.50		<0.230-3.50	NA	NA	<3.50	NA	<0.230-0.406	NA	<10	NA	<0.200
Pesticides	0.001	0.001-0.100		<0.001-0.100	<0.06	<0.060	<0.005-0.050	<0.100	<0.001	<0.080	<0.1-1	NA	<0.001-0.020
PCBs	0.020	0.020-0.100	10	<0.020-0.100	<0.060	<0.60	<0.050J	<0.100	<0.020	<0.020	<1	NA	<0.020
TOC				32,000-199,000	NA	NA	199,000	NA	32,000	NA	7,800	NA	<5,000

CT&E Data.

Not analyzed.

Result is an estimate.

The action level for DRPH is based on conversations with ADEC; a final action level has not yet been determined.

The laboratory reported that the Extractable Petroleum Hydrocarbon (EPH) pattern in this sample was not consistent with a middle distillate fuel.

Laboratory reported that the sample is moss and the EPH pattern may be due to biogenic hydrocarbons.

☐ NA
☐ J
☐ a
☐ c
☐ d

TABLE D-1. BACKGROUND ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Background (BKGD)									
Parameters	Matrix: Sediment Units: mg/kg		Quant. Limits	Action Levels	Bkgd. Range	Environmental Sample	Field Blanks		Lab Blanks
	Detect. Limits						AB03	EB08	
Laboratory Sample ID Numbers						2SD03 1746 4616-17	1712	1720 4616-13	#5-9693 #1&2-9693 #1&2-9493 4616 #6-9593 4616
ANALYSES									
DRPH	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	μg/L	μg/L	μg/L
	4.00	4.00	500 ^a	500 ^a	9.55-1,150	1,150 ^{cd}	NA	<1,000J ^b	<1,000J
GRPH	0.400	9.0	100	100	<0.400-<9.0	<9.0 ^c	<50J ^b	<50J ^b	<20
RRPH (Approx.)	4.8	480	2,000 ^a	2,000 ^a	<480	<480	NA	<2000	<2,000
BTEX (8020/8020 Mod.)				10 Total BTEX	<0.250-<1,500	<0.5			
Benzene	0.01	0.1	0.5	0.5	<0.020-<0.300	<0.1	<1	<1	NA
Toluene	0.01	0.1			<0.020-<0.300	<0.1	<3J	<1J	NA
Ethylbenzene	0.01	0.1			<0.020-<0.300	<0.1	<2J	<1	NA
Xylenes (Total)	0.02	0.2			<0.040-<0.600	<0.2	<5J	<2	NA
HVOC (8010 Mod.)	0.05	0.5			<0.5J	<0.5J	<9J	<5	NA
VOC 8260	0.020	0.500			<0.025-<0.500	<0.500	NA	<1	<1-6.6
									<0.020

☐ CT&E Data.
☒ F&B Data.
☒ N/A
☐ J
☐ a
☐ b
☐ c
☐ d

Result is an estimate.

The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined.
 DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.
 These samples were analyzed by F&B also; and DRPH and GRPH were detected at <240^b and <5J^b mg/kg, respectively.
 Laboratory reported that the EPH result indicates possible biogenic hydrocarbon contamination; sample was a very wet moss.

TABLE D-1. BACKGROUND ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Background (BKGD)		Matrix: Soil/Sediment Units: mg/kg		METALS ANALYSES						
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Bkgd. Range	Environmental Samples			Field Blank	Lab Blanks
						S02	SD01		EB01	
Laboratory Sample ID Numbers										
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	4175-2	4179-4		4175-3	4179 4175
Aluminum	0.35	2		1,500-25,000	1,700-8,700	8,700	1,700		<100	<100
Antimony	N/A	2		<7.8-<230	<59-<91	<91	<59		<100	<100
Arsenic	0.11	2-9.1		<4.9-8.5	<9.1-7	<9.1	7		<100	<100
Barium	0.024	1		27-390	27-120	120	27		<50	<50
Beryllium	N/A	3.0-4.6		<2.6-64	<3.0-<4.6	<4.6	<3.0		<50	<50
Cadmium	0.33	3.0-4.6		<3.0-<36	<3.0-<4.6	<4.6	<3.0		<50	<50
Calcium	0.69	4		360-59,000	5,100-12,000	5,100	12,000		<200	<200
Chromium	0.066	1		<4.3-47	3-14	14	3.0		<50	<50
Cobalt	N/A	9.1-59		<5.1-12	<9.1-<59	<9.1	<59		<100	<100
Copper	0.045	1-3.0		<2.7-45	<3.0-14	14	<3.0		<50	<50
Iron	0.50	2		5,400-35,000	7,000-11,000	11,000	7,000		<100	<100
Lead	0.13	5.9-9.1		<5.1-22	<5.9-<9.1	<9.1	<5.9		<100	<100
Magnesium	0.96	4		360-7,400	2,500-4,800	2,500	4,800		<200	<200
Manganese	0.025	1		25-290	76	76	76		<50	<50
Molybdenum	N/A	3.0-4.6		<2.5-<11	<3.0-<4.6	<4.6	<3.0		<50	<50
Nickel	0.11	1		4.2-46	4.2-14	14	4.2		<50	<50
Potassium	23	100-300		<300-2,200	<300-970	970	<300		<5,000	<5,000
Selenium	1.2	9.1-59		<7.8-<170	<9.1-<59	<9.1	<59		<100	<100

☐ CT&E Data.
NA Not available.

TABLE D-1. BACKGROUND ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Background (BKGD)		Matrix: Soil/Sediment Units: mg/kg		METALS ANALYSES						Field Blank		Lab Blanks
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Bkgd. Range	Environmental Samples						
Laboratory Sample ID Numbers						S02	SD01			EB01		
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	4175-2	4179-4			4175-3		4179 4175
Silver	0.53	3.0-4.6		<3-<110	<3.0-<4.6	mg/kg	mg/kg			μg/L		μg/L
Sodium	0.55	5		<160-680	71-230	230	71			<50		<50
Thallium	0.011	0.3-0.45		<0.2-<1.2	<0.3-<0.45	<0.45	<0.3J			<5		<5
Vanadium	0.036	1		6.3-59	7.7-22	22	7.7			<50		<50
Zinc	0.16	1		9.2-95	11-24	24	11			<50		<50

CT&E Data.
Result is an estimate.

☐ J

TABLE D-1. BACKGROUND ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Background (BKGD)		Matrix: Surface Water Units: µg/L		Environmental Samples			Field Blanks			Lab Blanks	
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range	SW01	SW02	AB01	EB02	TB02		
Laboratory Sample ID Numbers					4199-6 4179-2	4199-5 4179-3	4197-6 4173-9	4179-1 4206-4	4197-5 4199-13	4173 4199 4179	
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	
DRPH	100	200		<200	<200	<200	NA	<100	NA	<200	
GRPH	20	20		<20	<20	<20	NA	<20	NA	<20	
BTEX (8020/8020 Mod.)											
Benzene	1	1	5	<1	<1	<1	<1	<1	<1	<1	
Toluene	1	1	1,000	<1	<1	<1	1.2	<1	<1	<1	
Ethylbenzene	1	1	700	<1	<1	<1	<1	<1	<1	<1	
Xylenes (Total)	2	2	10,000	<2	<2	<2	<2	<2	<2	<2	
VOC 8010											
Chloroform	1	1		<1-1.1	<1	1.1	<1	<1-9	<1	<1	
1,2-Dichloroethane	1	1	5	1.3B-2.8B	1.3B	2.8B	1.2	2.9	<1	<1	
VOC 8260											
1,2-Dichloroethane	1	1	5	3U-3.2B	3U	3.2B	1.6	3.0	<1	<1	
SVOC 8270	10	10		<10	<10	<10	NA	<11	NA	<10	
Pesticides	0.05	1		<1	<1	<1	NA	<0.1-<1	NA	<0.1-<1	

CT&E Data.

Not analyzed.

The analyte was detected in the associated blank.

Compound is not present above the concentration listed.

☐ NA B U

TABLE D-1. BACKGROUND ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Background (BKGD)		Matrix: Surface Water Units: µg/L		Action Levels	Bkgd. Range	Environmental Samples			Field Blanks			Lab Blanks
Parameters	Detect. Limits	Quant. Limits				SW01	SW02		AB01	EB02	TB02	
Laboratory Sample ID Numbers						4199-6 4179-2	4199-5 4179-3		4197-6 4173-9	4179-1 4206-4	4197-5 4199-13	4199 4179
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L
PCBs	1	1	0.5	<1	<1	<1	<1	<1	NA	<1	NA	<1
TOC	5,000	5,000		<5,000-12,700	<5,000-12,700	<5,000	12,700		NA	<5,000	NA	<5,000
TSS	100	100-30,000		<30,000-8,000	<30,000-8,000	8,000J	<30,000		NA	4,000	NA	<100
TDS	10,000	10,000-352,000		<352,000-328,000	<352,000-328,000	328,000	<352,000		NA	30,000	NA	<10,000

☐ CT&E Data.
☐ NA
☐ J
 Not analyzed.
 Result is an estimate.

TABLE D-1. BACKGROUND ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Background (BKGD)		Matrix: Surface Water Units: µg/L		METALS ANALYSES: TOTAL (DISSOLVED)					Field Blank		Lab Blank
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Bkgd. Levels	Environmental Samples					
Laboratory Sample ID Numbers						SW01	SW02		EB02		
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	4179-2	4179-3		4179-1		4179
Aluminum	17.4	100		<100-350 (<100-340)	<100 (<100)	<100 (<100)	<100 (<100)		<100 (<100)		<100 (<100)
Antimony	N/A	100	6	<100 (<100)	<100 (<100)	<100 (<100)	<100 (<100)		<100 (<100)		<100 (<100)
Arsenic	5.3	100	50	<100 (<100)	<100 (<100)	<100 (<100)	<100 (<100)		<100 (<100)		<100 (<100)
Barium	1.2	50	2,000	<50-93 (<50-91)	<50-69 (<50-68)	69 (68)	<50 (<50)		<50 (<50)		<50 (<50)
Beryllium	N/A	50	4	<50 (<50)	<50 (<50)	<50 (<50)	<50 (<50)		<50 (<50)		<50 (<50)
Cadmium	1.7	50	5	<50 (<50)	<50 (<50)	<50 (<50)	<50 (<50)		<50 (<50)		<50 (<50)
Calcium	34.5	200		4,500-88,000 (4,100-86,000)	21,000-38,000 (21,000-38,000)	38,000 (38,000)	21,000 (21,000)		<200 (<200)		<200 (<200)
Chromium	3.29	50	100	<50 (<50)	<50 (<50)	<50 (<50)	<50 (<50)		<50 (<50)		<50 (<50)
Cobalt	N/A	100		<100 (<100)	<100 (<100)	<100 (<100)	<100 (<100)		<100 (<100)		<100 (<100)
Copper	2.3	50	1,300	<50 (<50)	<50 (<50)	<50 (<50)	<50 (<50)		<50 (<50)		<50 (<50)
Iron	25	100		180-2,800 (<100-1,600)	280-700 (<100-360)	280 (<100)	700 (360)		<100 (<100)		<100 (<100)
Lead	6.6	100	15	<100 (<100)	<100 (<100)	<100 (<100)	<100 (<100)		<100 (<100)		<100 (<100)

☐ CT&E Data.
N/A Not available.

TABLE D-1. BACKGROUND ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Background (BKGD)			Matrix: Surface Water Units: µg/L		METALS ANALYSES: TOTAL (DISSOLVED)					Field Blank		Lab Blank
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Bkgd. Levels	SW01	SW02	Environmental Samples			EB02	
Laboratory Sample ID Numbers						4179-2	4179-3				4179-1	4179
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L				µg/L	µg/L
Magnesium	47.8	200		<5,000-53,000 (2,600-54,000)	12,000-14,000 (13,000-14,000)	14,000 (14,000)	12,000 (13,000)				<200 (<200)	<200 (<200)
Manganese	1.24	50		<50-510 (<50-120)	<50 (<50)	<50 (<50)	<50 (<50)				<50 (<50)	<50 (<50)
Molybdenum	N/A	50		<50 (<50)	<50 (<50)	<50 (<50)	<50 (<50)				<50 (<50)	<50 (<50)
Nickel	5.5	50	100	<50 (<50)	<50 (<50)	<50 (<50)	<50 (<50)				<50 (<50)	<50 (<50)
Potassium	1,154	5,000		<5,000 (<5,000)	<5,000 (<5,000)	<5,000 (<5,000)	<5,000 (<5,000)				<5,000 (<5,000)	<5,000 (<5,000)
Selenium	62.4	100	50	<100 (<100)	<100 (<100)	<100 (<100)	<100 (<100)				<100 (<100)	<100 (<100)
Silver	2.6	50	50	<50 (<50)	<50 (<50)	<50 (<50)	<50 (<50)				<50 (<50)	<50 (<50)
Sodium	27.7	250		8,400-410,000 (8,200-450,000)	47,000-49,000 (50,000-57,000)	49,000 (50,000)	47,000 (57,000)				<250 (<250)	<250 (<250)
Thallium	0.57	5	2	<5 (<5)	<5 (<5)	<5 (<5)	<5 (<5)				<5J (<5)J	<5 (<5)
Vanadium	1.8	50		<50 (<50)	<50 (<50)	<50 (<50)	<50 (<50)				<50 (<50)	<50 (<50)
Zinc	8.2	50		<50-160 (<50)	<50 (<50)	<50 (<50)	<50 (<50)				<50 (<50)	<50 (<50)

☐ CT&E Data.
☐ N/A
☐ J
 Not available.
 Result is an estimate.

TABLE D-2. OLD LANDFILL ANALYTICAL DATA SUMMARY

Installation: Barter Island Site: Old Landfill (LF01)					Matrix: Soil/Sediment Units: mg/kg										Lab Blanks	
Parameters	Detect Limits	Quant Limits	Action Levels	Bkgd. Levels	Environmental Samples				Field Blanks							
					S01	SD01	SD02	SD03	AB02	EB04	EB05	TB04	TB05			
Laboratory Sample ID Numbers					322 4303-2	1340	1342	1344 4286-5	315 4303-1	311 4302-10	332 392 4303-5	1346 4302-9	375	#6-82493 #3&4-83193 #1&2-82493 4302/4303		
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	μg/L	μg/L	μg/L	μg/L	μg/L	mg/kg		
DRPH	5-11	50-110	500 ^a	9.55-1,150	<500 ^b	<50 ^b	<110 ^b	<110 ^b	NA	NA	<1,000 ^b	NA	NA	<70J		
GRPH	0.2-0.4	2-4	100	<0.4-0.9	NA	<2J ^b	<4J ^b	<4J ^b	<100J ^b	<100J ^b	<100J ^b	<100J ^b	<50J	<2J		
RRPH (Approx.)	10-22	100-220	2,000 ^a	<480	<100	<110	<220	530	NA	NA	<1,000	NA	NA	<100		
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-1.500	NA	<0.10	<0.20	<0.20								
Benzene	0.002-0.004	0.02-0.04	0.5	<0.020-0.300	NA	<0.02	<0.04	<0.04	<1	<1	<1	<1	<1	<0.02		
Toluene	0.002-0.004	0.02-0.04		<0.020-0.300	NA	<0.02	<0.04	<0.04	<1	<1	<1	<1	<1	<0.02		
Ethylbenzene	0.002-0.004	0.02-0.04		<0.020-0.300	NA	<0.02	<0.04	<0.04	<1	<1	<1	<1	<1	<0.02		
Xylenes (Total)	0.004-0.008	0.04-0.08		<0.040-0.600	NA	<0.04	<0.08	<0.08	<2	<2	<2	<2	<2	<0.04		
HVOC (8010 Mod.)	0.002-0.004	0.02-0.04		<0.5J	NA	<0.02J	<0.04J	<0.04J	<1	<1	<1	<1J	<1	NA		
VOC 8260																
1,4-Dichloro- benzene	0.020	0.025-0.030		<0.025-0.500	<0.025J	NA	NA	0.044	<1	<1	<1	<1	NA	<0.020		
Naphthalene	0.020	0.025-0.030		<0.025-0.500	<0.025J	NA	NA	0.105	<1	<1	<1	<1	NA	<0.020		
1,2,4-Trichloro- benzene	0.020	0.025-0.030		<0.025-0.500	<0.025J	NA	NA	0.046	<1	<1	<1	<1	NA	<0.020		
1,2,4-Trimethyl- benzene	0.020	0.025-0.030		<0.025-0.500	<0.025J	NA	NA	0.041	<1	<1	<1	<1	NA	<0.020		
Xylenes (Total)	0.040	0.050-0.060		<0.050-1.000	<0.050J	NA	NA	0.033 ^c	<2	<2	<2	<2	NA	<0.020		

CT&E Data.

F&B Data.

Not analyzed.

Result is an estimate.

The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined.

DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

Result is indicative of p & m xylenes only.

☐ CT&E Data.
☒ F&B Data.
☐ Not analyzed.
☐ Result is an estimate.
☐ The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined.
☐ DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.
☐ Result is indicative of p & m xylenes only.

TABLE D-2. OLD LANDFILL ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Old Landfill (LF01)					Matrix: Soil/Sediment Units: mg/kg										
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples				Field Blanks					Lab Blanks	
					S01	SD01	SD02	SD03	AB02	EB04	EB05	TB04	TB05	#5-82493 4302/4303	#6-82393 4303
Laboratory Sample ID Numbers					322 4303-2	1340	1342	1344 4286-5	315 4303-1	311 4302-10	332 392 4303-5	1346 4302-9	375		
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	µg/L	µg/L	µg/L	µg/L	µg/L	mg/kg	
SVOC 8270	0.200	0.230		<0.23-<3.5	<0.230R	NA	NA	NA	NA	<25-75	<11	NA	NA	<10	<0.200
PCBs	0.01	0.1	10	<0.020-<0.100	<0.5R	<0.1	<0.1	<0.1	NA	NA	≤10	NA	NA	NA	<0.1-<0.5
TOC				32,000-199,000	820	NA	NA	NA	NA	<5,000	<5,000J	NA	NA	<5,000	NA

☐ CT&E Data.
☒ F&B Data.
☒ NA
 Not analyzed.
 Result is an estimate.
 Result has been rejected.

TABLE D-2. OLD LANDFILL ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Old Landfill (LF01)			Matrix: Soil/Sediment Units: mg/kg		METALS ANALYSES							
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Environmental Samples				Field Blanks		Lab Blanks	
					S01	SD03			EB04	EB05		
Laboratory Sample ID Numbers					4303-2	4286-5				4302-10	4303-5	4286 4302
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg				µg/L	µg/L	µg/L
Aluminum	0.35	2		1,500-25,000	1,900	3,600				<100	<100	<100
Antimony	N/A	56-64		<7.8-<230	<56	<64				<100	<100	<100
Arsenic	0.11	5.6-6.4		<4.9-8.5	<5.6	<6.4				<100	<100	<100
Barium	0.024	1		27-390	19	76				<50	<50	<50
Beryllium	N/A	2.8-3.2		<2.6-6.4	<2.8	<3.2				<50	<50	<50
Cadmium	0.33	2.8-3.2		<3.0-<36	<2.8	<3.2				<50	<50	<50
Calcium	0.69	4		360-59,000	4,650	10,600				<200	<200	<200
Chromium	0.066	1		<4.3-47	4.1	6.6				<50	<50	<50
Cobalt	N/A	5.6-6.4		<5.1-12	<5.6	<6.4				<100	<100	<100
Copper	0.045	1		<2.7-45	3.3	14				<50	<50	<50
Iron	0.50	2		5,400-35,000	5,000	12,000				<100	200	<100
Lead	0.13	2-5.6		<5.1-22	<5.6	29				<100	<100	<100
Magnesium	0.96	4		360-7,400	990	2,700				<200	<200	<200
Manganese	0.025	1		25-290	52	170				<50	<50	<50
Molybdenum	N/A	2.8-3.2		<2.5-<11	<2.8	<3.2				<50	<50	<50
Nickel	0.11	1		4.2-46	4.3	8.6				<50	<50	<50
Potassium	23	100-280		<300-2,200	<280	610				<5,000	<5,000	<5,000
Selenium	1.2	5.6-64		<7.8-<170	<5.6	<64				<100	<100	<100

☐ CT&E Data.
N/A Not analyzed.

TABLE D-2. OLD LANDFILL ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Old Landfill (LF01)			Matrix: Soil/Sediment Units: mg/kg		METALS ANALYSES						
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Environmental Samples			Field Blanks		Lab Blanks	
					S01	SD03		EB04	EB05		
Laboratory Sample ID Numbers					4303-2	4286-5		4302-10	4303-5	4286 4302	
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg		μg/L	μg/L	μg/L	
Silver	0.53	2.8-3.2		<3-<110	<2.8	<3.2		<50	<50	<50	
Sodium	0.55	5		<160-680	78	160		<250	<250	<250-267	
Thallium	0.011	0.27-0.28		<0.2-<1.2	<0.27	<0.28		<5	<5	<5	
Vanadium	0.036	1		6.3-59	4.9	11		<50	<50	<50	
Zinc	0.16	1		9.2-95	12	55		<50	<50	<50	

CT&E Data.

TABLE D-2. OLD LANDFILL ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Old Landfill (LF01)													
Parameters	Detect Limits	Quant Limits	Action Levels	Bkgd. Levels	Environmental Samples				Field Blanks				Lab Blanks
					SW01	SW02	SW03	SW04	AB02	EB04	EB05	TB04	
Laboratory Sample ID Numbers					1372 1374	1368 1370	1362 1364 4285-2	1348 1354 4302-1 4285-1	315 4303-1	311 4302-10	332 392 4303-5	1346 4302-9	#3&4-83193 #3&4-82493 4285 4302/4303
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L
DRPH	100	1,000		<200	<1,000 ^B	<1,000 ^B	<1,000 ^B	<1,000 ^B	NA	NA	<1,000 ^B	NA	NA
GRPH	10	100		<20	<100J ^B	<100J ^B	<100J ^B	<100J ^B	<100J ^B	<100J ^B	<100J ^B	<100J ^B	<100J
RRPH (Approx.)	200	2,000		NA	<2,000	<2,000	<2,000	<2,000	NA	NA	<1,000	NA	NA
BTEx (8020/8020 Mod.)													
Benzene	0.1	1	5	<1	<1	<1	<1	<1	<1	<1	<1	<1	<1
Toluene	0.1	1	1,000	<1	56	<1	<1	<1	<1	<1	<1	<1	<1
Ethyl-benzene	0.1	1	700	<1	<1	<1	<1	<1	<1	<1	<1	<1	<1
Xylenes (Total)	0.2	2	10,000	<2	<2	<2	<2	<2	<2	<2	<2	<2	<2
HVOC 8010	0.1	1		<0.5J	<1	<1	<1	<1	<1	<1	<1	<1J	<1J
VOC 8260													
1,2-Dichloroethane	1	1	5	3U-3.2B	NA	NA	3.9B	3.9B	6.3	1.5	3.2	<1	<1
p-Isopropyltoluene	1	1		<1	NA	NA	1.7	<1	<1	<1	<1	<1	<1
Toluene	1	1	1,000	<1	NA	NA	<1	2.5B	2.2	2.3	2.3	<1	<1
SVOC 8270	10	20-22		<10	NA	NA	<22	<20	NA	<25-75	<11	NA	<10

CT&E Data.

F&B Data.

Not analyzed.

The analyte was detected in the associated blank.

Result is an estimate.

Result has been rejected.

Compound is not present above the concentration listed.

DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

AK-RISK\BARTER\APP-D\4109661203\TBLD-2.TBL

TABLE D-2. OLD LANDFILL ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Old Landfill (LF01)													
Matrix: Surface Water Units: µg/L													
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples				Field Blanks				Lab Blanks
					SW01	SW02	SW03	SW04	AB02	EB04	EB05	TB04	
Laboratory Sample ID Numbers					1372 1374	1368 1370	1362 1364 4285-2	1348 1354 4285-1	315 4303-1	311 4302-10	332 392 4303-5	1346 4302-9 4303	4285 4302 4303
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L
PCBs	0.2	2	0.5	<1	<2R	<2J	<2	<2J	NA	NA	<10	NA	NA
TOC	5,000	5,000		<5,000-12,000	NA	NA	53,900J	98,000	NA	<5,000	<5,000J	NA	<5,000
TSS	100	100		<30,000-8,000	NA	NA	8,200	200,000	NA	NA	NA	NA	<100
TDS	10,000	10,000		<352,000-328,000	NA	NA	1,226,000J	2,416,000	NA	NA	NA	NA	<10,000

CT&E Data.
F&B Data.
Not analyzed.
Result is an estimate.
Result has been rejected.

□ ■ NA
J R

TABLE D-2. OLD LANDFILL ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Old Landfill (LF01)		Matrix: Surface Water Units: µg/L		METALS ANALYSES: TOTAL (DISSOLVED)		Environmental Samples				Field Blank		Lab Blanks
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	SW03	SW04				EB04		
Laboratory Sample ID Numbers					4285-2	4285-1				4302-10		4285 4302
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L				µg/L		µg/L
Aluminum	17.4	100		<100-350 (<100-340)	180 (<100)	<100 (<100)				<100		<100 (<100)
Antimony	N/A	50	6	<100 (<100)	<100 (<100)	<100 (<100)				<100		<100 (<100)
Arsenic	5.3	50	50	<100 (<100)	<100 (<100)	<100 (<100)				<100		<100 (<100)
Barium	1.2	50	2,000	<50-93 (<50-91)	120 (89)	120 (110)				<50		<50 (<50)
Beryllium	N/A	50	4	<50 (<50)	<50 (<50)	<50 (<50)				<50		<50 (<50)
Cadmium	1.7	50	5	<50 (<50)	<50 (<50)	<50 (<50)				<50		<50 (<50)
Calcium	34.5	200		4,500-88,000 (4,100-86,000)	130,000 (130,000)	190,000 (180,000)				<200		<200-378 (378)
Chromium	3.29	50	100	<50 (<50)	<50 (<50)	<50 (<50)				<50		<50 (<50)
Cobalt	N/A	100		<100 (<100)	<100 (<100)	<100 (<100)				<100		<100 (<100)
Copper	2.3	50	1,300	<50 (<50)	<50 (<50)	<50 (<50)				<50		<50 (<50)
Iron	25	200		180-2,800 (<100-1,600)	15,000 (240)	3,200 (340)				<100		<100 (<100)
Lead	6.6	100	15	<100 (<100)	<100 (<100)	<100 (<100)				<100		<100 (<100)

☐ CT&E Data.
N/A Not analyzed.

TABLE D-2. OLD LANDFILL ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Old Landfill (LF01)		Matrix: Surface Water Units: µg/L		METALS ANALYSES: TOTAL (DISSOLVED)					Field Blank		Lab Blanks
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	SW03	SW04	Environmental Samples			EB04	
Laboratory Sample ID Numbers					4285-2	4285-1				4302-10	4285 4302
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L				µg/L	µg/L
Magnesium	47.8	50		<5,000-53,000 (2,600-54,000)	55,000 (54,000)	78,000 (76,000)				<200	<200 (<200)
Manganese	1.24	50		<50-510 (<50-120)	1,500 (1,400)	560 (480)				<50	<50 (<50)
Molybdenum	N/A	50		<50 (<50)	<50 (<50)	<50 (<50)				<50	<50 (<50)
Nickel	5.5	50	100	<50 (<50)	<50 (<50)	<50 (<50)				<50	<50 (<50)
Potassium	1,154	5,000		<5,000 (<5,000)	9,400 (9,100)	110,000 (101,000)				<5,000	<5,000 (<5,000)
Selenium	62.4	100	50	<100 (<100)	<100 (<100)	<100 (<100)				<100	<100 (<100)
Silver	2.6	50	50	<50 (<50)	<50 (<50)	<50 (<50)				<50	<50 (<50)
Sodium	27.7	250		8,400-410,000 (8,200-450,000)	160,000 (150,000)	440,000 (410,000)				<250	<250-267 (<250)
Thallium	0.57	5	2	<5 (<5)	<5 (<5)	<5 (<5)				<5	<5 (<5)
Vanadium	1.8	50		<50 (<50)	<50 (<50)	<50 (<50)				<50	<50 (<50)
Zinc	8.2	50		<50-160 (<50)	<50 (<50)	<50 (<50)				<50	<50 (<50)

☐ CT&E Data.
N/A Not analyzed.

TABLE D-2. OLD LANDFILL ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Old Landfill (LF01)			Matrix: Surface Water Units: µg/L										
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples					Field Blank		Lab Blank	
					2SW01	2SW02	2SW03	2SW04					EB06
Laboratory Sample ID Numbers					1663	1664	1665	1666				1688 1690	#5-9693
ANALYSES	µg/L	µg/L			µg/L	µg/L	µg/L	µg/L				µg/L	µg/L
DRPH	100	1,000		<200	<1,000 ^b	<1,000 ^b	<1,000 ^b	<1,000 ^b				<1,000J ^b	<1,000J
RRPH (Approx.)	200	2,000		NA	<2,000	<2,000	<2,000	<2,000				<2,000	<2,000

CT&E Data.

F&B Data.

Not analyzed.

Result is an estimate.

DRPH concentrations reported for these samples are equivalent to diesel range organics (DRO) as defined by ADEC.

□ ■ NA J_b

TABLE D-3 POL CATCHMENT ANALYTICAL DATA SUMMARY

Installation: Barter Island Site: POL Catchment (LF03)		Matrix: Soil Units: mg/kg		Field Blanks										Lab Blanks	
Parameters	Detect. Limits	Quant. Limits	Action Limits	Bkgd. Levels	Environmental Samples					Field Blanks					Lab Blanks
					S01	S01-0.75	S02-0.75	S03-0.75		AB01	AB02	EB03	EB04	TB03	TB04
Laboratory Sample ID Numbers					1290 4302-1	4216-7	4216-8	4216-9		4173-9 4197-6	315 4303-1	4213-4 4215-6	311 4302-10	4211-2	1346 4302-9
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg		μg/L	μg/L	μg/L	μg/L	μg/L	μg/L
DRPH	4.00	4.00	500 ^a	9.55-1,150	5,200 ^b	8.84 ^d	6.95 ^d	4.76 ^d		NA	NA	<100	NA	NA	<100
GRPH	0.400	0.400	100	<0.4-<9.0	19 ^b	<0.400	<0.400	<0.400		NA	<100 ^b	<20	<100 ^b	NA	<20-<100J
RRPH (Approx.)	10-30	100-300	2,000 ^a	<480	180	NA	NA	NA		NA	NA	NA	NA	NA	NA
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-<1.500	1.23J	<0.100	<0.100	<0.100							
Benzene	0.020	0.020	0.5	<0.020-<0.300	<0.02	<0.020	<0.020	<0.020		<1	<1	<1	<1	<1 ^c	<1
Toluene	0.020	0.020		<0.020-<0.300	0.08	<0.020	<0.020	<0.020		1.2	<1	3.2	<1	<1 ^c	<1
Ethylbenzene	0.020	0.020		<0.020-<0.300	0.05	<0.020	<0.020	<0.020		<1	<1	<1	<1	<1 ^c	<1
Xylenes (Total)	0.040	0.040		<0.040-<0.600	1.1J	<0.040	<0.040	<0.040		<2	<2	<2	<2	<2 ^c	<2
VOC B010	0.020	0.020		<0.020-<0.300	NA	<0.020	<0.020	<0.020		<1-9 B	NA	NA	NA	NA	NA
VOC B260															
n-Butylbenzene	0.020	0.200		<0.025-<0.500	3.70	NA	NA	NA		NA	<1	NA	<1	NA	<1
sec-Butylbenzene	0.020	0.200		<0.025-<0.500	1.62	NA	NA	NA		NA	<1	NA	<1	NA	<1
Ethylbenzene	0.020	0.200		<0.025-<0.500	0.313	NA	NA	NA		NA	<1	NA	<1	NA	<1
Isopropylbenzene	0.020	0.200		<0.025-<0.500	0.409	NA	NA	NA		NA	<1	NA	<1	NA	<1
p-Isopropyltoluene	0.020	0.200		<0.025-<0.500	1.69	NA	NA	NA		NA	<1	NA	<1	NA	<1
n-Propylbenzene	0.020	0.200		<0.025-<0.500	0.920	NA	NA	NA		NA	<1	NA	<1	NA	<1

CT&E Data.

F&B Data.

Not analyzed.

NA

Result is an estimate.

The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined.

DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

BTEX determined by 8260 method analysis.

The laboratory reported that the EPH pattern in this sample was not consistent with an unweathered middle distillate fuel.

☐ a
☒ b
☐ c
☐ d

05 MAY 1995

TABLE D-3. POL CATCHMENT ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: POL Catchment (LF03)				Matrix: Sediment Units: mg/kg										Field Blanks				Lab Blanks	
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples														
					SD01	SD02 & SD08 (Replicates)		SD03-1	SD04	SD05	SD06-1	SD07	AB01	EB03	TB03				
Laboratory Sample ID Numbers					4216-11 4212-3	4219-13	4219-12	4215-5	4215-1	4215-4	4219-11	4213-3 4219-10	4173-9 4197-6	4213-4 4215-6	4211-2	4215 4216 4213 4215 4213 4212			
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	μg/L	μg/L	μg/L	mg/kg			
DRPH	4.00	4.00	500 ^a	9.55-1,150	2,250	1,940	2,460	2,610	28,600	10,700	268 ^e	15.2J ^d	NA	<100	NA	<4.00			
GRPH	0.400	0.400-0.600	100	<0.4-9.0	12.7	<0.500	0.429	6.30	78.5	67.9	<0.600	<0.400	NA	<20	NA	<0.400			
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-1.500	1.15	<0.100	<0.100	0.161	4.309	6.282	<0.150	<0.100							
Benzene	0.020	0.020-0.070	0.5	<0.020-0.300	<0.070	<0.020	<0.020	<0.02	<0.035	<0.04	<0.030	<0.020	<2	<1	<1	<0.020			
Toluene	0.020	0.020-0.070		<0.020-0.300	<0.070	<0.020	<0.020	<0.02	0.116B	0.108B	<0.030	<0.020	1.2	3.2	<1 ^c	<0.020			
Ethylbenzene	0.020	0.020-0.070		<0.020-0.300	0.323	<0.020	<0.020	0.034	0.739	0.982	<0.030	<0.020	<1	<1	<1 ^c	<0.020			
Xylenes (Total)	0.040	0.040-0.140		<0.040-0.600	0.827	<0.040	<0.040	0.127	3.57	5.30	<0.060	<0.040	<2	<2	<2 ^c	<0.040			
VOC 8010	0.020	0.020-0.070		<0.020-0.300	<0.070	<0.020	<0.020	NA	NA	NA	<0.030	<0.020	<1.9.8	NA	NA	<0.020			
VOC 8260																			
n-Butylbenzene	0.020	0.120-0.400		<0.025-0.500	0.635	NA	NA	NA	NA	NA	NA	<0.120J	<1	<1	<1	<0.020			
Methylene Chloride	0.020	0.120-0.400	90	<0.025-0.500	<0.400	NA	NA	NA	NA	NA	NA	0.255JB	1.1	2.4	<1	<0.020			
Tetrachloroethene	0.020	0.120-0.400		<0.025-0.500	<0.400	NA	NA	NA	NA	NA	NA	5.42J	<1	<1	<1	<0.020			
1,2,4-Trimethylbenzene	0.020	0.120-0.400		<0.025-0.500	0.638	NA	NA	NA	NA	NA	NA	<0.120J	<1	<1	<1	<0.020			
1,3,5-Trimethylbenzene	0.020	0.120-0.400		<0.025-0.500	0.536	NA	NA	NA	NA	NA	NA	0.379J	<1	<1	<1	<0.020			
SVOC 8270																			
2-Methylnaphthalene	0.200	0.460-1.00		<0.23-3.5	3.44 J	NA	NA	NA	NA	NA	NA	<0.460	NA	<10	NA	<0.200			

☐ CT&E Data.

☐ NA Not analyzed.

☐ B The analyte was detected in the associated blank.

☐ J Result is an estimate.

☐ a The action level for DRPH is based on conversations with ADEC; a final action level has not yet been determined.

☐ c BTEX determined by 8260 method analysis.

☐ d The laboratory reported that the EPH pattern in this sample was not consistent with a middle distillate fuel.

☐ e The laboratory reported that 183 mg/kg of the EPH pattern in this sample was not consistent with a middle distillate fuel.

TABLE D-3. POL CATCHMENT ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: POL Catchment (LF03)				Matrix: Sediment Units: mg/kg											
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples								Field Blanks		Lab Blanks
					SD01	SD02 & SD08 (Replicates)	SD03-1	SD04	SD05	SD06-1	SD07	AB01	EB03	TB03	
Laboratory Sample ID Numbers					4216-11 4212-3	4219-13 4219-12	4215-5	4215-1	4215-4	4219-11	4213-3 4219-10	4173-9 4197-6	4213-4 4215-6	4211-2	4215 4213 4211 4219 4216 4215 4213 4212
					mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	µg/L	µg/L		µg/L
					NA	NA	NA	NA	NA	NA	19,500	NA	<5,000	NA	<5,000
					mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	µg/L	µg/L	µg/L	mg/kg
ANALYSES															
TOC															

☐ CT&E Data.
Not analyzed.

TABLE D-3. POL CATCHMENT ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: POL Catchment (LF03)		Matrix: Sediment Units: mg/kg		Environmental Samples			Field Blanks		Lab Blanks	
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	2SD09	2SD10	AB02	EB06		
Laboratory Sample ID Numbers					1678	1680	315 4303-1	1688/1690	#5-9693 #3&4-82493 #1&2-9493	#3&4-9693
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	μg/L	μg/L	μg/L	mg/kg
DRPH	5-20	50-200	500 ^a	9.55-1,150	<200J ^b	<50J ^b	NA	<1,000J ^b	<1,000J	NA
GRPH	0.2-0.8	2-8	100	<0.400-<9.0	<8J ^b	<2J ^b	<100J ^b	<50J ^b	<50J	<2J
RRPH (Approx.)	10-40	100-400	2,000 ^a	<480	<400	<100	NA	<1,000	<2,000	NA
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-<1.500	<0.44	<0.10				
Benzene	0.002-0.008	0.02-0.08	0.5	<0.020-<0.300	<0.08	<0.02	<1	<1	<1	<0.02
Toluene	0.002-0.008	0.02-0.08		<0.020-<0.300	<0.08	<0.02	<1	<4J	<1	<0.02
Ethylbenzene	0.002-0.008	0.02-0.08		<0.020-<0.300	<0.08	<0.02	<1	<2J	<1	<0.02
Xylenes (Total)	0.004-0.02	0.04-0.2		<0.040-<0.600	<0.2	<0.04	<2	<5J	<2	<0.04

CT&E Data.

F&B Data.

Not analyzed.

Result is an estimate.

The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined.

DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

☐ CT&E Data.
☒ F&B Data.
☐ Not analyzed.
☐ Result is an estimate.

TABLE D-3. POL CATCHMENT ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: POL Catchment (LF03)				Matrix: Surface Water Units: µg/L							
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples			Field Blanks			Lab Blanks
					SW01	SW02	SW03	AB01	EB03	TB03	
Laboratory Sample ID Numbers					4216-6 4212-4	4219-8	4213-2 4219-9	4173-9 4197-6	4213-4 4215-6	4211-2	4219/4216 4215/4213 4212/4211
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L
DRPH	100	100		<200	612 ^{ad}	1,770 ^{ad}	1,200 ^{ad}	NA	<100	NA	<100
GRPH	20	20		<20	367 ^a	<20	<20	NA	<20	NA	<20
BTEX (8020/8020 Mod.)											
Benzene	1	1	5	<1	2.7	<1	<1	<1	<1	<1 ^c	<1
Toluene	1	1	1,000	<1	<1	<1	<1	1.2	3.2	<1 ^c	<1
Ethylbenzene	1	1	700	<1	19	<1	<1	<1	<1	<1 ^c	<1
Xylenes (Total)	1	1	10,000	<2	38.4	<2	<2	<2	<2	<2 ^c	<2
VOC 8010											
1,2-Dichloroethane	1	1	5	1.3B-2.8B	4.6B	NA	NA	1.2	NA	NA	<1
VOC 8260											
Benzene	1	1	5	<1	2.5	NA	<1	<1	<1	<1	<1
n-Butylbenzene	1	1		<1	2.4	NA	<1	<1	<1	<1	<1
sec-Butylbenzene	1	1		<1	1.0	NA	<1	<1	<1	<1	<1
1,2-Dichloroethane	1	1	5	3U-3.2B	3.9B	NA	4.4B	1.6	2.2	<1	<1
Ethylbenzene	1	1	700	<1	18	NA	<1	<1	<1	<1	<1

☐ CT&E Data.
☐ NA Not analyzed.
☐ B The analyte was detected in the associated blank.
☐ U Compound is not present above the concentration listed.
☐ a Total petroleum hydrocarbons in these water samples exceed the 15 µg/L stated for fresh water in ADEC's Water Quality Criteria 18 AAC 70 (ADEC 1989).
☐ c BTEX determined by 8260 method analysis.
☐ d The laboratory reported that the EPH pattern in this sample was not consistent with an unweathered middle distillate fuel.

TABLE D-3. POL CATCHMENT ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: POL Catchment (LF03)		Matrix: Surface Water Units: µg/L		Bkgd. Levels	Environmental Samples			Field Blanks			Lab Blanks
Parameters	Detect. Limits	Quant. Limits	Action Levels		SW01	SW02	SW03	AB01	EB03	TB03	
Laboratory Sample ID Numbers					4216-6 4212-4	4219-8	4213-2 4219-9	4173-9 4197-6	4213-4 4215-6	4211-2	4219/4216 4215/4213 4212/4211
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L
Isopropylbenzene	1	1		<1	2.9	NA	<1	<1	<1	<1	<1
p-Isopropyltoluene	1	1		<1	2.1	NA	<1	<1	<1	<1	<1
Naphthalene	1	1		<1	35	NA	<1	<1	<1	<1	<1
n-Propylbenzene	1	1		<1	3.6	NA	<1	<1	<1	<1	<1
1,2,4-Trimethylbenzene	1	1		<1	19	NA	<1	<1	<1	<1	<1
1,3,5-Trimethylbenzene	1	1		<1	13	NA	<1	<1	<1	<1	<1
Xylenes (Total)	2	2	10,000	<2	33.3	NA	<2	<2	<2	<2	<2
SVOC 8270	10	10-11		<10	<10.6-11	NA	<10	NA	<10	NA	<10
TOC	5,000	5,000		<5,000-12,700	48,100	NA	21,000	NA	<5,000	NA	<5,000
TSS	100	100		<30,000-8,000	71,000	NA	12,000	NA	NA	NA	<100-<200
TDS	10,000	10,000		<352,000-328,000	847,000	NA	637,000	NA	NA	NA	<10,000

☐ CT&E Data.
☐ NA
 Not analyzed.

TABLE D-4. CURRENT LANDFILL ANALYTICAL DATA SUMMARY

Installation: Barter Island Site: Current Landfill (LF04)		Matrix: Soil/Sediment Units: mg/kg		Action Levels	Bkgd. Levels	Environmental Samples				Field Blanks			Lab Blanks
Parameters	Detect Limits	Quant. Limits				S01	S02	SD01	SD02	AB02	EB04	TB04	
Laboratory Sample ID Numbers						326	1334	1336 4286-3	1338 4286-4	315 4303-1	311 4302-10	1346 4302-9	#3&4-82493 #5-82493 #3&4-82493 #1&2-83193 428
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/L	mg/L	μg/L	μg/L
DRPH	6-55	60-550	500 ^a	9.55-1,150	<70 ^b	<70 ^b	<340 ^a	<550 ^b	80 ^b	NA	NA	NA	<50-70J
GRPH	0.2-1	2-10	100	<0.4-8	<2J ^b	<2J ^b	<10J ^a	<6J ^b	<2J ^b	<100J ^b	<100J ^b	<100J ^b	<2J
RRPH (Approx.)	12-56	120-560	2,000 ^a	<480	<140	<140	<560	<340	<120	NA	NA	NA	<100
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-1 500	<1.5	<1.5	<0.5	<0.28	<0.15				
Benzene	0.003-0.01	0.03-0.1	0.5	<0.020-0.300	<0.03	<0.03	<0.1	<0.06	<0.03	<1	<1	<1	<0.02
Toluene	0.003-0.01	0.03-0.1		<0.020-0.300	<0.03	<0.03	<0.1	<0.06	<0.03	<1	<1	<1	<0.02
Ethylbenzene	0.003-0.01	0.03-0.1		<0.020-0.300	<0.03	<0.03	<0.1	<0.06	<0.03	<1	<1	<1	<0.02
Xylenes (Total)	0.006-0.02	0.06-0.2		<0.040-0.600	<0.06	<0.06	<0.2	<0.1	<0.06	<2	<2	<2	<0.04
HVOC 8010	0.003-0.01	0.03-0.1		<0.5J	<0.03J	<0.03J	<0.1J	<0.06J	<0.03J	<1	<1	<1J	NA
VOC 8260													
Toluene	0.020	0.200		<0.025-0.500	NA	NA	NA	<0.200	0.026	<1	<1	<1	<0.020
SVOC 8270	0.200	0.200		<0.230-3.50	NA	NA	NA	<0.200	<0.200	NA	<25-75	NA	<0.200
PCBs	0.01-0.07	0.1-0.7	10	<0.020-0.100	<0.5-0.7	<0.1	<0.1	<0.1	<0.1	NA	NA	NA	<0.1-0.5
TOC				32,000-199,000	NA	NA	NA	196,000	2,860	NA	<5,000	NA	NA

CT&E Data.

F&B Data.

Not analyzed.

Result is an estimate.

The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined.

DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

☐ CT&E Data.
☒ F&B Data.
☐ Not analyzed.
☐ Result is an estimate.

TABLE D-4. CURRENT LANDFILL ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Current Landfill (LF04)		Matrix: Soil/Sediment Units: mg/kg		Environmental Samples		Field Blanks		Lab Blanks	
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	2SD03	2SD04	AB03	EB08	
Laboratory Sample ID Numbers					1742 4616-15	1744 4616-16	1712	1719/1720 4616-13	#5-9693 #1&2-9693 #1&2-9793 4616
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	µg/L	µg/L	mg/kg
DRPH	4.00-6	18.3-60	500	9.55-1,150	<18.3 ^{cd}	<60 ^b	NA	<1,000 ^{jb}	<50
GRPH	0.400-0.2	0.400-2	100	<0.400-9.0	<0.400 ^c	8.1 ^b	<50 ^{jb}	<50 ^{jb}	<0.400-1J
RRPH (Approx.)	12-64	120-640	2,000 ^a	<480	<640	<120	NA	<2,000	<100
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-1.500	0.2J	<0.10			
Benzene	0.002-0.01	0.02-0.1	0.5	<0.020-0.300	<0.1	<0.02	<1	<1	<0.02
Toluene	0.002-0.01	0.02-0.1		<0.020-0.300	<0.1	<0.02	<3J	<1J	<0.02
Ethylbenzene	0.002-0.01	0.02-0.1		<0.020-0.300	<0.1	<0.02	<2J	<1	<0.02
Xylenes (Total)	0.004-0.02	0.04-0.2		<0.040-0.600	0.2J	<0.04	<5J	<2	<0.04
HVOC 8010	0.01-0.05	0.1-0.5		<0.5J	<0.5J	<0.1J	<9J	<5J	<0.1J-10J
VOC 8260	0.020	0.020		<0.025-0.500	<0.020	<0.020	NA	<1-6.6	<0.020

☐ CT&E Data.
☒ F&B Data.
☒ Not analyzed.

Result is an estimate.

The action levels for DRPH and RRPB are based on conversations with ADEC; final action levels have not yet been determined.
 DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.
 Sample was analyzed by F&B also; DRPH and GRPH were detected at <310^b and <5.5^b mg/kg, respectively.
 The laboratory reported that the EPH pattern in this sample was not consistent with a middle distillate fuel.

TABLE D-4. CURRENT LANDFILL ANALYTICAL DATA SUMMARY (CONTINUED)

METALS ANALYSES											
Installation: Barter Island Site: Current Landfill (LF04)		Matrix: Sediment Units: mg/kg									
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Environmental Samples				Field Blank		Lab Blanks
					SD01	SD02				EB04	
Laboratory Sample ID Numbers											
ANALYSES					4286-3	4286-4				4302-10	4286 4302
Aluminum	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg				μg/L	μg/L
	0.35	2		1,500-25,000	7,500	1,900				<100	<100
Antimony	N/A	55-260		<7.8-<230	<260	<55				<100	<100
Arsenic	0.11	5.5-26		<4.9-8.5	<26	<5.5				<100	<100
Barium	0.024	1-30		27-390	120	<30				<50	<50
Beryllium	N/A	2.8-13		<2.6-6.4	<13	<2.8				<50	<50
Cadmium	0.33	2.8-13		<3.0-<36	<13	<2.8				<50	<50
Calcium	0.69	4		360-59,000	10,000	7,200				<200	<200
Chromium	0.066	1-13		<4.3-47	<13	3.4				<50	<50
Cobalt	N/A	5.5-26		<5.1-12	<26	<5.5				<100	<100
Copper	0.045	1		<2.7-45	18	9.9				<50	<50
Iron	0.50	2		5,400-35,000	17,000	4,700				<100	<100
Lead	0.13	2-26		<5.1-22	<26	6.1				<100	<100
Magnesium	0.96	4		360-7,400	2,800	3,800				<200	<200
Manganese	0.025	1		25-290	380	40				<50	<50
Molybdenum	N/A	2.8-13		<2.5-<11	<13	<2.8				<50	<50
Nickel	0.11	1		4.2-46	16	4.6				<50	<50
Potassium	23	280-1,300		<300-2,200	<1,300	<280				<5,000	<5,000
Selenium	1.2	5.6-26		<7.8-<170	<26	<5.6				<100	<100

☐ CT&E Data.
N/A Not available.

TABLE D-4. CURRENT LANDFILL ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Current Landfill (LF04)		Matrix: Sediment Units: mg/kg		METALS ANALYSES						
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Environmental Samples			Field Blank		Lab Blanks
					SD01	SD02		EB04		
Laboratory Sample ID Numbers					4286-3	4286-4		4302-10		4286 4302
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg		µg/L		µg/L
Silver	0.53	2.8-13		<3-<100	<13	<2.8		<50		<50
Sodium	0.55	5		<160-680	870	89		<250		<250-267
Thallium	0.11	0.26-1.3		<0.2-<1.2	<1.3	<0.26		<5		<5
Vanadium	0.036	1		6.3-59	21	6.3		<50		<50
Zinc	0.16	1		9.2-95	65	31		<50		<50

TABLE D-4. CURRENT LANDFILL ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Current Landfill (LF04)		Matrix: Surface Water Units: µg/L								
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples		Field Blanks			Lab Blanks
					SW01	SW02	AB02	EB04	TB04	
Laboratory Sample ID Numbers					1324/1326 4286-1	1328/1330 4286-2	315 4303-1	332 4302-10	1346 4302-9	#5-83193 4303/4302 4286
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L
p-Isopropyltoluene	1	1		<1	<1	1.1	<1	<1	<1	<1
Toluene	1	1	1,000	<1	<1	1.3B	2.2	2.3	<1	<1
Trichloroethene	1	1	5	<1	<1	6	<1	<1	<1	<1
SVOC 8270	10	22		<10	<22	<22	NA	<25-75	NA	<10
PCBs	0.2	2	0.5	<1	<2	<2	NA	NA	NA	NA
TOC	5,000	5,000		<5,000-12,700	23,400	26,600	NA	<5,000	NA	<5,000
TSS	100	200		<30,000-8,000	156,000	22,000	NA	NA	NA	<200
TDS	10,000	10,000		<352,000-328,000	905,000	614,000	NA	NA	NA	<10,000

□ CT&E Data.

■ F&B Data.

NA Not analyzed.

B The analyte was detected in the associated blank.

TABLE D-4. CURRENT LANDFILL ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Current Landfill (LF04)		Matrix: Surface Water Units: µg/L		Action Levels	Bkgd. Levels	Environmental Samples			Field Blanks			Lab Blanks
Parameters	Detect. Limits	Quant. Limits				2SW01	2SW02	2SW03	AB03	EB06	EB08	
Laboratory Sample ID Numbers						1669	1670	1747/1748 4616-14	1712	1688/1690	1719/1720 4616-13	#5-9693 #1&2-9793 #1&2-9493 4616
ANALYSES	µg/L	µg/L		µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L
DRPH	100	1,000			<200	<1,000 ^b	<1,000 ^b	<1,000 ^b	NA	<1,000 ^b	<1,000 ^b	<1,000J
GRPH	5	50			<20	NA	NA	<50J ^b	<50J ^b	<100J ^b	<50J ^b	<2J-<50
RRPH (Approx.)	200	2,000			NA	<2,000	<2,000	<2,000	NA	<2,000	<2,000	<2,000
BTEx (8020/8020 Mod.)												
Benzene	0.1	1	5		<1	NA	NA	<1	<1	<1	<1	<1-<5
Toluene	0.1	1	1,000		<1	NA	NA	<1	<3J	<4J	<1J	<1
Ethylbenzene	0.1	1	700		<1	NA	NA	<1	<2J	<2J	<1	<1
Xylenes (Total)	0.2	2	10,000		<1	NA	NA	<2	<5J	<5J	<2	<2
HVOC 8010												
Trichloroethene	0.1	1	5		NA	NA	NA	8J	<1	<1	<1J	<1-<10J
VOC 8260	1	1			<1-3.2B	NA	NA	<1	NA	NA	<1-6.6	<1

CT&E Data.

F&B Data.

NA

Not analyzed.

The analyte was detected in the associated blank.

Result is an estimate.

DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

TABLE D-4. CURRENT LANDFILL ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Current Landfill (LF04)				Matrix: Surface Water Units: µg/L		METAL ANALYSES				
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Environmental Samples			Field Blank		Lab Blanks
					SW01	SW02			EB04	
Laboratory Sample ID Numbers					4286-1	4286-2				4286 4302
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L			µg/L	µg/L
Aluminum	17.4	100		<100-350 (<100-340)	<740J (<100)J	<100 (<100)			<100	<100 (<100)
Antimony	N/A	100	6	<100 (<100)	<100J (<100)J	<100 (<100)			<100	<100 (<100)
Arsenic	5.3	100	50	<100 (<100)	<100J (<100)J	<100 (<100)			<100	<100 (<100)
Barium	1.2	50	2,000	<50-93 (<50-91)	150J (61)J	<50 (<50)			<50	<50 (<50)
Beryllium	N/A	50	4	<50 (<50)	<50J (<50)J	<50 (<50)			<50	<50 (<50)
Cadmium	1.7	50	5	<50 (<50)	<50J (<50)J	<50 (<50)			<50	<50 (<50)
Calcium	34.5	200		4,500-88,000 (4,100-86,000)	120,000J (120,000)J	58,000 (59,000)			<200	<200 (<200)
Chromium	3.29	50	100	<50 (<50)	<50J (<50)J	<50 (<50)			<50	<50 (<50)
Cobalt	N/A	100		<100 (<100)	<100J (<100)J	<100 (<100)			<100	<100 (<100)
Copper	2.3	50	1,300	<50 (<50)	<50J (<50)J	<50 (<50)			<50	<50 (<50)
Iron	25	100		180-2,800 (<100-1,600)	21,000J (890)J	2,400 (2,200)			<100	<100 (<100)

☐ CT&E Data.
☐ N/A
☐ J
 Not available.
 Result is an estimate.

TABLE D-4. CURRENT LANDFILL ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Current Landfill (LF04)			Matrix: Surface Water Units: µg/L		METAL ANALYSES					
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Environmental Samples			Field Blank		Lab Blanks
					SW01	SW02			EB04	
Laboratory Sample ID Numbers					4286-1	4286-2			4302-10	4286 4302
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L			µg/L	µg/L
Lead	6.6	10-100	15	<100 (<100)	<10J (<100)J	<100 (<100)			<100	<100 (<100)
Magnesium	47.8	200		<5,000-53,000 (2,600-54,000)	41,000J (41,000)J	29,000 (28,000)			<200	<200 (<200)
Manganese	1.24	50		<50-510 (<50-120)	1,800J (260)J	150 (150)			<50	<50 (<50)
Molybdenum	N/A	50		<50 (<50)	<50J (<50)J	<50 (<50)			<50	<50 (<50)
Nickel	5.5	50	100	<50 (<50)	<50J (<50)J	<50 (<50)			<50	<50 (<50)
Potassium	1,154	500		<5,000 (<5,000)	5,200J (<5,000)J	12,000 (10,000)			<5,000	<5,000 (<5,000)
Selenium	62.4	100	50	<100 (<100)	<100J (<100)J	<100 (<100)			<100	<100 (<100)
Silver	2.6	50	50	<50 (<50)	<50J (<50)J	<500 (<50)			<50	<50 (<50)
Sodium	27.7	250		8,400-410,000 (8,200-450,000)	100,000J (100,000)J	100,000 (87,000)			<250	267
Thallium	0.57	5	2	<5 (<5)	<5J (<5)J	<5 (<5)			<5	<5 (<5)
Vanadium	1.8	50		<50 (<50)	<50J (<50)J	<50 (<50)			<50	<50 (<50)
Zinc	8.2	50		<50-160 (<50)	<50J (<50)J	<50 (<50)			<50	<50 (<50)

☐ CT&E Data.
☐ N/A
☐ J Not analyzed.
 Result is an estimate.

TABLE D-5. CONTAMINATED DITCH ANALYTICAL DATA SUMMARY

Installation: Barter Island Site: Contaminated Ditch (SD08)				Matrix: Soil Units: mg/kg		Environmental Samples										Field Blanks			Lab Blanks	
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	S01-3 & S10-3 (Replicates)	S02-3	S03-3.5	S04-2.5	S05-3	S06-2	S07-3	AB01	EB01	TB02						
Laboratory Sample ID Numbers					4199-9	4199-8	4199-10	4199-12	4199-11	4216-2	4216-1	4173-9	4175-3	4179-5	4203 4199 4197	4199 4203 4216				
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	µg/L	µg/L	µg/L	µg/L	mg/kg				
DRPH	4.00	4.00	500 ^a	9.55-1,150	691	535	7.03	234	742 ^c	743 ^e	22.4 ^d	NA	<200	NA	<200	<4.00				
GRPH	0.400	0.400	100	<0.4-<9	134	171	<0.400	163	9.84	23.8	<0.400	NA	<20	NA	<20	<0.400				
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-<1.500	7.14	9.00	0.099	15.54	1.629	<0.852	<0.100									
Benzene	0.020	0.020-0.200	0.5	<0.020-<0.300	<0.100	<0.200	<0.020	<0.100	<0.100	<0.080	<0.020	<1	<1	<1	<1	<0.020				
Toluene	0.020	0.020-0.200		<0.020-<0.300	0.885B	0.182JB	0.024B	0.874B	<0.100	0.532B	<0.020	1.2	<1	<1	<1	<0.020				
Ethylbenzene	0.020	0.020-0.200		<0.020-<0.300	1.59	2.14	0.027	3.01	0.279	<0.080	<0.020	<1	<1	<1	<1	<0.020				
Xylenes (Total)	0.040	0.040-0.400		<0.040-<0.600	5.55	6.86	0.072	12.53	1.35	<0.160	<0.040	<2	<2	<2	<2	<0.040				
VOC 8010	0.020	0.020-0.100		<0.020-<0.300	<0.075	<0.100	<0.020	<0.020	<0.020	NA	NA	<1-9.8	<1-2.5	<1	<1	<0.020				

☐ NA
☐ J
☐ B
☐ a
☐ c
☐ d
☐ e

CT&E Data.

Not analyzed.

Result is an estimate.

The analyte was detected in the associated blank.

The action level for DRPH is based on conversations with ADEC; a final action level has not yet been determined.

The laboratory reported that 213 mg/kg of the EPH pattern in this sample was not consistent with a middle distillate fuel.

The laboratory reported that 19.9 mg/kg of the EPH pattern in this sample was not consistent with a middle distillate fuel.

The laboratory reported that the EPH pattern in this sample was not consistent with an unweathered middle distillate fuel.

TABLE D-5. CONTAMINATED DITCH ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Contaminated Ditch (SD08)				Matrix: Soil Units: mg/kg									
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples				Field Blanks		Lab Blanks		
					2S11-3.5	2S12-4.5	2S13-3.5 & 2S14-3.5 (Replicates)	2S15-4	2S16-1	A803		EB07	
Laboratory Sample ID Numbers					1730	1732	1734	1738	1740	1712	1715 1718 4818-9	#5-9893 4818 #1&2-9893	
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	µg/L	µg/L	µg/L	
DRPH	6	60	500 ^a	9.55-1,150	<60 ^b	<60 ^b	<60 ^b	<60 ^b	<60 ^b	NA	<200 ^c	<200	
GRPH	0.1	1	100	<0.400-<9.0	<1 ^b	<1 ^b	<1 ^b	<1 ^b	<1 ^b	<50 ^b	<20 ^c	<20	
RRPH (Approx.)	12	120	2,000 ^a	<480	<120	<120	<120	<120	<120	NA	<1,000	<2,000	
BTEX (8020/8020 Mod.)			10 Total BTEx	<0.250-<1,500	<0.10	<0.10	<0.10	<0.10	<0.10				
Benzene	0.002	0.02	0.5	<0.020-<0.300	<0.02	<0.02	<0.02	<0.02	<0.02	<1	<1	NA	
Toluene	0.002	0.02		<0.020-<0.300	<0.02	<0.02	<0.02	<0.02	<0.02	<3J	<2J	NA	
Ethylbenzene	0.002	0.02		<0.020-<0.300	<0.02	<0.02	<0.02	<0.02	<0.02	<2J	<1	NA	
Xylenes (Total)	0.004	0.04		<0.040-<0.600	<0.04	<0.04	<0.04	<0.04	<0.04	<5J	<2	NA	

CT&E Data.

F&B Data.

Not analyzed.

Result is an estimate.

The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined. DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC. This sample was analyzed by F&B also; DRPH and GRPH were detected at <1,030^b and <50^b µg/L, respectively.

□ NA
■ J
a
b
c

TABLE D-5. CONTAMINATED DITCH ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Contaminated Ditch (SD08)		Matrix: Sediment Units: mg/kg											
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples					Field Blanks			Lab Blanks
					SD01	SD02	SD03	SD04	SD05 & SD09 (Replicates)	AB01	EB01	TB01	
Laboratory Sample ID Numbers					4198-8 4178-3	4198-9	4198-10	4198-11	4198-12 4173-5	4173-9 4197-6	4203-8 4175-3	4197-7 4173-10	4173 4178 4198
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	µg/L	µg/L	µg/L	mg/kg
DRPH	4.00	4.00	500 ^a	9.55-1,150	<4.00	<4.00	<4.00	7.14 ^c	7.46	NA	<200	NA	<4.00
GRPH	0.400	0.400-2.00	100	<0.4-<8	2.97	<0.400	<0.400	<0.400	<2.00	NA	<20	NA	<0.400
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-<1.500	<0.100	<0.100	<0.100	<0.100	<0.100	<1	<1	<1	<0.020
Benzene	0.020	0.020	0.5	<0.020-<0.300	<0.020	<0.020	<0.020	<0.020	<0.020	<1	<1	<1	<0.020
Toluene	0.020	0.020		<0.020-<0.300	<0.020	<0.020	<0.020	<0.020	<0.020	1.2	<1	<1	<0.020
Ethylbenzene	0.020	0.020		<0.020-<0.300	<0.020	<0.020	<0.020	<0.020	<0.020	<1	<1	<1	<0.020
Xylenes (Total)	0.040	0.040		<0.040-<0.600	<0.040	<0.040	<0.040	<0.040	<0.040	<2	<2	<2	<0.040
VOC 8260	0.020	0.020-0.025		<0.025-<0.500	<0.020	NA	NA	NA	<0.025	<1-1.9	<1J-4.4J	<1	<0.025
SVOC 8270	0.200	0.200-1.00		<0.23-<3.5	NA	NA	NA	NA	<0.200-<1.00	NA	<10	NA	<0.200
TOC				32,000-199,000	2,780	NA	NA	NA	1,670	NA	7,800	NA	NA

☐ CT&E Data.
☐ NA
☐ J
☐ a
☐ c

Not analyzed.
 Result is an estimate.
 The action level for DRPH is based on conversations with ADEC; a final action level has not yet been determined.
 The laboratory reported that the EPH pattern in this sample was not consistent with a middle distillate fuel.

TABLE D-5. CONTAMINATED DITCH ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Contaminated Ditch (SD08)			Matrix: Sediment Units: mg/kg		METALS ANALYSES						
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Environmental Samples				Field Blank		Lab Blanks
					SD01	SD05 & SD09 (Replicates)				EB01	
Laboratory Sample ID Numbers					4178-3	4173-6	4178-4			4175-3	4178 4175 4173
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg			µg/L	µg/L
Aluminum	0.35	2-2,800		1,500-25,000	<2,800	2,400	2,500			<100	<100
Antimony	N/A	54-57		<7.8-<230	<54	<57	<57			<100	<100
Arsenic	0.11	5.4-5.7		<4.9-8.5	<5.4	<5.7	<5.7			<100	<100
Barium	0.024	1		27-390	27	32	44			<50	<50
Beryllium	N/A	1-2.7		<2.6-6.4	<2.7	3.2	2.8			<50	<50
Cadmium	0.33	1-2.8		<3.0-<36	<2.7	<2.8	2.8			<50	<50
Calcium	0.69	4		360-59,000	12,300	5,000	6,900			<200	<200
Chromium	0.066	1		<4.3-<47	4.4	5.3	4.0			<50	<50
Cobalt	N/A	1-57		<5.1-12	<5.4	<57	5.7			<100	<100
Copper	0.045	1		<2.7-45	4.7	5.8	4.1			<50	<50
Iron	0.50	2		5,400-35,000	9,800	12,000	8,700			<100	<100
Lead	0.13	5.4-5.7		<5.1-22	<5.4	<5.7	<5.7			<100	<100
Magnesium	0.96	4		360-7,400	2,600J	1,400	2,300			<200	<200
Manganese	0.025	1		25-290	98J	250	160			<50	<50
Molybdenum	N/A	2.9-5.7		<2.5-<11	<5.4	<2.9	<5.7			<50	<50
Nickel	0.11	1		4.2-46	7.6	6.6	6.1			<50	<50
Potassium	23	100		<300-2,200	440	336	370			<5,000	<5,000

☐ CT&E Data.
☐ N/A Not available.
☐ J Result is an estimate.

TABLE D-5. CONTAMINATED DITCH ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Contaminated Ditch (SD08)		Matrix: Sediment Units: mg/kg		METALS ANALYSES						
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Environmental Samples			Field Blank		Lab Blanks
					SD01	SD05 & SD09 (Replicates)			EB01	
Laboratory Sample ID Numbers					4178-3	4173-6	4178-4		4175-3	4178 4175 4173
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg		µg/L	µg/L
Selenium	1.2	2-5.7		<7.8-<170	<5.4	57	<5.7		<100	<100
Silver	0.53	2.7-2.9		<3-<110	<2.7	<2.9	<2.8		<50	<50
Sodium	0.55	5		<160-680	80	72	65		<250	<250
Thallium	0.011	0.28		<0.2-<1.2	<0.28J	<0.28	<0.28		<5	<5
Vanadium	0.036	1		6.3-59	7.8	8.4	7.7		<50	<50
Zinc	0.16	1		9.2-95	20J	27	27		<50	<50

CT&E Data.
Result is an estimate.

☐ J

TABLE D-5. CONTAMINATED DITCH ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Contaminated Ditch (SD08)				Matrix: Surface Water Units: µg/L								
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples				Field Blanks			Lab Blanks
					SW01	SW02	SW03	SW04 & SW08 (Duplicates)	AB01	EB02	TB01	
Laboratory Sample ID Numbers					4178-2 4197-1	4197-2	4197-3	4175-1 4197-4	4173-1 4197-5	4179-1 4208-4	4197-7 4173-10	4206 4197 4178 4175 4173
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L
DRPH	100	200		<200	<200	<200	<200	<200	<200	NA	NA	<100
GRPH	20	20		<20	<20	<20	<20	<20	<20	NA	NA	<20
BTX (8020/8020 Mod.)												
Benzene	1	1-20	5	<1	<1	<1	<20	<1	<1	<1	<1	<1
Toluene	1	1-20	1,000	<1	<1	<1	<20	<1	<1	1.2	<1	<1
Ethylbenzene	1	1-20	700	<1	<1	<1	<20	<1	<1	<1	<1	<1
Xylenes (Total)	2	2-40	10,000	<2	<2	<2	<40	<2	<2	<2	<2	<2
VOC 8260												
1,2-Dichloroethane	1	1	5	3U-3.2B	5.4B	NA	NA	5.3B	4.4B	1.6	3.0	<1
cis-1,2-Dichloroethene	1	1	70	<1	<1	NA	NA	1.4	1.5	<1	<1	<1
SVOC 8270	10	10		<10	<10	NA	NA	<10	<10	NA	<11	<10
TOC	5,000	5,000		<5,000-12,700	15,000	NA	NA	13,000	11,300	NA	NA	<5,000
TSS	100	200		<30,000-8,000	57,000	NA	NA	6,000	8,000	NA	NA	<200
TDS	10,000	10,000		<352,000-328,000	662,000	NA	NA	610,000	590,000	NA	NA	12,000

☐ CT&E Data.
☐ NA Not analyzed.
☐ B The analyte was detected in the associated blank.
☐ U Compound is not present above the concentration listed.

TABLE D-5. CONTAMINATED DITCH ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Contaminated Ditch (SD08)			Matrix: Surface Water Units: µg/L		METALS ANALYSES: TOTAL (DISSOLVED)							Field Blank		Lab Blanks
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	SW01	SW04 & SW08 (Replicates)							EB01	
Laboratory Sample ID Numbers					4178-2	4175-1	4173-1						4179-1	4179 4178 4175 4173
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L						µg/L	µg/L
Aluminum	17.4	100		<100-350 (<100-340)	1,900 (<100)	<100 (<100)	<100 (<100)						<100	<100 (<100)
Antimony	N/A	100	6	<100 (<100)	<100 (<100)	<100 (<100)	<100 (<100)						<100	<100 (<100)
Arsenic	5.3	100	50	<100 (<100)	<100 (<100)	<100 (<100)	<100 (<100)						<100	<100 (<100)
Barium	1.2	50	2,000	<50-93 (<50-91)	110 (82)	83 (79)	86 (78)						<50	<50 (<50)
Beryllium	N/A	50	4	<50 (<50)	<50 (<50)	<50 (<50)	<50 (<50)						<50	<50 (<50)
Cadmium	1.7	50	5	<50 (<50)	<50 (<50)	<50 (<50)	<50 (<50)						<50	<50 (<50)
Calcium	34.5	100		4,500-88,000 (4,100-86,000)	7,200 (71,000)	63,000 (65,000)	63,000 (61,000)						<200	<200 (<200)
Chromium	3.29	50	100	<50 (<50)	<50 (<50)	<50 (<50)	<50 (<50)						<50	<50 (<50)
Cobalt	N/A	100		<100 (<100)	<100 (<100)	<100 (<100)	<100 (<100)						<100	<100 (<100)

☐ CT&E Data.
N/A Not available.

TABLE D-5. CONTAMINATED DITCH ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Contaminated Ditch (SD08)			Matrix: Surface Water Units: µg/L		METALS ANALYSES: TOTAL (DISSOLVED)							Field Blank		Lab Blanks
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	SW01	SW04 & SW08 (Replicates)							EB01	
Laboratory Sample ID Numbers					4178-2	4175-1	4173-1						4179-1	4179 4178 4175 4173
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L						µg/L	µg/L
Copper	2.3	50	1,300	<50 (<50)	<50 (<50)	<50 (<50)	<50 (<50)						<50 (<50)	<50 (<50)
Iron	25	100		180-2,800 (<100-1,600)	4,200 (<100)	2,200 (180)	2,400 (<100)						<100 (<100)	<100 (<100)
Lead	6.6	100	15	<100 (<100)	<100 (<100)	<100 (<100)	<100 (<100)						<100 (<100)	<100 (<100)
Magnesium	47.8	100		<5,000-53,000 (2,600-54,000)	31,000 (30,000)	27,000 (27,000)	27,000 (27,000)						<200 (<200)	<200 (<200)
Manganese	1.24	100		<50-510 (50-120)	170 (100)	150 (130)	150 (120)						<50 (<50)	<50 (<50)
Molybdenum	N/A	50		<50 (<50)	<50 (<50)	<50 (<50)	<50 (<50)						<50 (<50)	<50 (<50)
Nickel	5.5	50	100	<50 (<50)	<50 (<50)	<50 (<50)	<50 (<50)						<50 (<50)	<50 (<50)
Potassium	1,154	5,000		<5,000 (<5,000)	<5,000 (<5,000)	<5,000 (<5,000)	<5,000 (<5,000)						<5,000 (<5,000)	<5,000 (<5,000)
Selenium	62.4	100	50	<100 (<100)	<100 (<100)	<100 (<100)	<100 (<100)						<100 (<100)	<100 (<100)
Silver	2.6	50	50	<50 (<50)	<50 (<50)	<50 (<50)	<50 (<50)						<50 (<50)	<50 (<50)

☐ CT&E Data.
N/A Not available.

TABLE D-5. CONTAMINATED DITCH ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Contaminated Ditch (SD08)		Matrix: Surface Water Units: µg/L		METALS ANALYSES: TOTAL (DISSOLVED)					Field Blank		Lab Blanks
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	SW01	SW04 & SW08 (Replicates)	Environmental Samples			EB01	
Laboratory Sample ID Numbers					4178-2	4175-1	4173-1			4179-1	4179 4178 4175 4173
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L			µg/L	µg/L
Sodium	27.7	100		8,400-410,000 (8,200-450,000)	120,000 (96,000)	98,000 (91,000)	90,000 (85,000)			<250	<250 (<250)
Thallium	0.57	5	2	<5 (<5)	<5 (<5)	<5 (<5)	<5 (<5)			<5	<5 (<5)
Vanadium	1.8	50		<50 (<50)	<50 (<50)	<50 (<50)	<50 (<50)			<50	<50 (<50)
Zinc	8.2	50		<50-160 (<50)	<50 (<50)	<50 (<50)	<50 (<50)			<50	<50 (<50)

CT&E Data.

TABLE D-6. OLD RUNWAY DUMP ANALYTICAL DATA SUMMARY

Installation: Barter Island Site: Old Runway Dump (LF12)		Matrix: Soil Units: mg/kg		Bkgd. Levels	Environmental Samples			Field Blanks			Lab Blanks	
Parameters	Detect. Limits	Quant. Limits	Action Levels		S01	S02	S03	AB02	EB05	TB05		
Laboratory Sample ID Numbers					416 4305-10	412	414	315 4303-1	332 392 4303-5	375	#3&4-82493 #1&2-82493 4303	#6-82393 #1&2-82493 4305
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	μg/L	μg/L	μg/L	μg/L	mg/kg
DRPH	5	50	500 ^a	9.55-1,150	<50J ^b	<50J ^b	<50J ^b	NA	<100J ^b	NA	NA	<50J
GRPH	0.2	2	100	<0.4-<9	<2J ^b	<2J ^b	<2J ^b	<100J ^b	<100J ^b	<50J ^b	<100J	<2J
RRPH (Approx.)	10	100	2,000 ^a	<480	<100	<100	<100	NA	<1000	NA	NA	<100
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-<1.500	<0.10J	<0.10J	<0.10J					
Benzene	0.002	0.02	0.5	<0.020-<0.300	<0.02J	<0.02J	<0.02J	<1	<1	<1	<1	<0.02J
Toluene	0.002	0.02		<0.020-<0.300	<0.02	<0.02	<0.02	<1	<1	<1	<1	<0.02
Ethylbenzene	0.002	0.02		<0.020-<0.300	<0.02	<0.02	<0.02	<1	<1	<1	<1	<0.02
Xylenes (Total)	0.004	0.04		<0.040-<0.600	<0.04	<0.04	<0.04	<2	<2	<2	<2	<0.04
HVOC (8010 Mod.)	0.002	0.02		<0.50J	<0.02J	<0.02J	<0.02J	<1	<1	<1	<1	<0.02J
VOC 8260	0.020	0.020		<0.025-<0.500	<0.02	NA	NA	<1-6.3	<1-3.2	NA	<1	<0.020
SVOC 8270	0.200	0.200-1.00		<0.230-<3.50	<0.200-<1.00	NA	NA	NA	<1	NA	<10	<0.200
PCBs	0.05	0.5	10	<0.020-<0.100	<0.5	<0.5	<0.5	NA	<10	NA	NA	<0.1-<0.5
Pesticides	0.001-0.05	0.01-0.5		<0.001-<0.100	<0.01J-<0.5J	NA	NA	NA	<0.2-<10	NA	NA	<0.01-<0.5
TOC				32,000-199,000	7,680	NA	NA	NA	<5,000J	NA	<5,000	NA

□

CT&E Data.

■

F&B Data.

NA

Not analyzed.

J

Result is an estimate.

a

The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined.

b

DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

TABLE D-6. OLD RUNWAY DUMP ANALYTICAL DATA SUMMARY (CONTINUED)

METALS ANALYSES									
Installation: Barter Island Site: Old Runway Dump (LF12)			Matrix: Soil Units: mg/kg		Environmental Samples				
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installation	S01			Field Blank	Lab Blanks
Laboratory Sample ID Numbers					4305-10			EB05	
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg			µg/L	4305 4303 µg/L
Aluminum	0.35	2		1,500-25,000	1,700			<100	<100
Antimony	N/A	52		<7.8-<230	<52			<100	<100
Arsenic	0.11	5.2		<4.9-8.5	<5.2			<100	<100
Barium	0.024	1		27-390	16			<50	<50
Beryllium	N/A	2.6		<2.6-6.4	<2.6			<50	<50
Cadmium	0.33	2.6		<3.0-<36	<2.6			<50	<50
Calcium	0.69	4		360-59,000	2,200			<200	<200
Chromium	0.066	1		<4.3-4.7	4.5			<50	<50
Cobalt	N/A	5.2		<5.1-12	<5.2			<100	<100
Copper	0.045	1		<2.7-45	7.3			<50	<50
Iron	0.50	2		5,400-35,000	7,500			200	<100
Lead	0.13	5.2		<5.1-22	<5.2			<100	<100
Magnesium	0.96	4		360-7,400	1,100			<200	<200
Manganese	0.025	1		25-290	67			<50	<50
Molybdenum	N/A	2.6		<2.5-<11	<2.6			<50	<50
Nickel	0.11	1		4.2-46	5.3			<50	<50
Potassium	23	260		<300-2,200	<260			<5,000	<5,000

☐ CT&E Data.
N/A Not available.

TABLE D-6. OLD RUNWAY DUMP ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Old Runway Dump (LF12)		Matrix: Soil Units: mg/kg		METALS ANALYSES									
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installation	Environmental Samples						Field Blank		Lab Blanks
					S01						EB05		
Laboratory Sample ID Numbers					4305-10						4303-5		4305 4303
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg						µg/L		µg/L
Selenium	1.2	52		<7.8-<170	<52						<100		<100
Silver	0.53	2.6		<3-<110	<2.6						<50		<50
Sodium	0.55	5		<160-680	410						<250		<250-267
Thallium	0.011	0.25		<0.2-<1.2	<0.25						<5		<5
Vanadium	0.036	1		6.3-59	5.1						<50		<50
Zinc	0.16	1		9.2-95	16						<50		<50

☐ CT&E Data.

TABLE D-7. HEATED STORAGE ANALYTICAL DATA SUMMARY

Installation: Barter Island Site: Heated Storage (Building 87) (S13)		Matrix: Soil Units: mg/kg											
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples				Field Blanks			Lab Blank	
Laboratory Sample ID Numbers					S01	S02	S03	S04	AB01	EB03A	TB03	4187/4173 4215/4211	4219/4216 4212
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	µg/L	µg/L	µg/L	µg/L	mg/kg
DRPH	4.00	4.00	500 ^a	9.55-1,150	116 ^f	3,580 ^g	1,290 ^d	797 ^e	NA	<100	NA	<100	<4.00
GRPH	0.400	0.400	100	<0.400-9.0	<0.400	423	<0.400	3.65	NA	<20	NA	<20	<0.400
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-1,500	0.021	3.445N	<0.100	0.22					
Benzene	0.020	0.020	0.5	<0.020-0.300	<0.020	<0.020	<0.020	<0.020	<1	<1	<1 ^c	<1	<0.020
Toluene	0.020	0.020		<0.020-0.300	<0.020	0.632N	<0.020	0.03	1.2	<1	<1 ^c	<1	<0.020
Ethylbenzene	0.020	0.020		<0.020-0.300	<0.020	0.283N	<0.020	0.03	<1	<1	<1 ^c	<1	<0.020
Xylenes (Total)	0.040	0.040		<0.040-0.600	0.021	2.53N	<0.040	0.16	<2	<2	<2 ^c	<2	<0.040
VOC 8010	0.020	0.020		<0.020-0.300	<0.020	<0.020	<0.020	<0.020	<1-9.8	NA	NA	<1	<0.020
VOC 8260													
1,3,5-Trimethylbenzene	0.020	0.020		<0.025-0.500	NA	7.66	NA	NA	<1	<1	<1	<1	<0.020
Xylenes (Total)	0.040	0.040		<0.050-1.000	NA	0.124	NA	NA	<2	<2	<2	<2	<0.040
SVOC 8270	0.200	2.10		<0.230-3.50	NA	<2.10	NA	NA	NA	<11	<1	<10	<0.200
PCBs													
Aroclor 1254	0.020	0.020	10	<0.020-0.100	0.932	2.72	2.4	0.316	NA	<1	NA	<1	<0.010

☐ CT&E Data.
☐ Not analyzed.
☐ Result is an estimate.
☐ The analysis indicates the presence of an analyte for which there is presumptive evidence to make a "tentative identification".
☐ The action level for DRPH is based on conversations with ADEC; a final action level has not yet been determined.
☐ BTEX determined by 8260 method analysis.
☐ The laboratory reported that the EPH pattern in this sample was not consistent with a middle distillate fuel.
☐ The laboratory reported that 310 mg/kg of the EPH pattern in this sample was not consistent with a middle distillate fuel.
☐ The laboratory reported that the EPH pattern in this sample was not consistent with an unweathered middle distillate fuel.
☐ The laboratory reported that 1090 mg/kg of the EPH pattern in this sample was not consistent with an unweathered middle distillate fuel.

TABLE D-7. HEATED STORAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Heated Storage (Building 87) (SS13)										Matrix: Soil Units: mg/kg		
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd Levels	Environmental Samples				Field Blanks		Lab Blanks	
					2S05-1	2S06-1	2S07-1			AB03	EB07	
Laboratory Sample ID Numbers					1724 4616-10	1726 4616-12	1728 4616-11		1712	1715 1716 4616-9	#3&4-82493 #1&2-9693 #5-9693 4616	#1&2-9693 #6-9593 4616
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg		µg/L	µg/L	µg/L	mg/kg
DRPH	4.00	4.00-60	500 ^a	9.55-1,150	<60 ^b	<4.00 ^c	1100 ^b		NA	<200 ^d	<200	<50
GRPH	0.400	0.400-1	100	<0.4-9	<1 ^b	<0.400 ^c	63 ^b		<500 ^a	<20 ^d	<20-50J	<0.400-1J
RRPH (Approx.)	12	120	2,000 ^a	<480	2,400	<120	<120		NA	<1,000	<2,000	<100
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-1.500	<0.10	<0.10	0.57J					
Benzene	0.002	0.02	0.5	<0.020-0.300	<0.02	<0.02	<0.02		<1	<1	<1	<0.02
Toluene	0.002	0.02		<0.020-0.300	<0.02	<0.02	<0.02		<3J	<2J	<1	<0.02
Ethylbenzene	0.002	0.02		<0.020-0.300	<0.02	<0.02	0.07		<2J	<1	<1	<0.02
Xylenes (Total)	0.004	0.04		<0.040-0.600	<0.04	<0.04	0.5J		<5J	<2	<2	<0.04
VOC 8260	0.020	0.020-0.070		<0.025-0.500	<0.070	<0.020	<0.020		NA	<1-8.3	<1	<0.020

☐ CT&E Data.
☒ F&B Data.
☐ Not analyzed.
☐ Result is an estimate.
☐ The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined.
☐ DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.
☐ This sample was analyzed by F&B also; DRPH and GRPH were detected at <60^a and <1^b mg/kg, respectively.
☐ This sample was analyzed by F&B also; DRPH and GRPH were detected at <1,000^b and <50^b µg/L, respectively.

TABLE D-7. HEATED STORAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Heated Storage (Building 87) (SS13)		Matrix: Sediment Units: mg/kg								
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples			Field Blanks		
					SD01	SD02	SD03	AB01	EB03A	TB03
Laboratory Sample ID Numbers					4212-2 4216-5	4219-1	4219-4	4173-9 4197-6	4211-1 4215-7	4211-2
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	µg/L	µg/L	µg/L
DRPH	4.00	4.00	500 ^a	9.55-1,150	770J ^f	36.5 ^d	51.4 ^e	NA	<100	NA
GRPH	0.400	0.400	100	<0.400-<9.0	0.653	<0.400	1.17	NA	<20	NA
BTEX (8020/8020 Mod.)			15 Total BTEX	<0.250-<1.500	0.321N	<0.100	0.48			
Benzene	0.020	0.020-0.025	0.5	<0.020-<0.300	<0.025	<0.020	<0.020	<1	<1	<1 ^c
Toluene	0.020	0.020-0.025		<0.020-<0.300	<0.025	<0.020	0.23	1.2	<1	<1 ^c
Ethylbenzene	0.020	0.020-0.025		<0.020-<0.300	0.028N	<0.020	<0.020	<1	<1	<1 ^c
Xylenes (Total)	0.040	0.040-0.050		<0.040-<0.600	0.293N	<0.040	0.25	<2	<2	<2 ^c
VOC 8010	0.020	0.020		<0.020-<0.300	NA	<0.020	<0.020	<1-9.8	NA	NA
VOC 8260										
Toluene	0.020	0.020		<0.025-<0.500	0.029	NA	NA	1.9	3.3	<1
Xylenes (Total)	0.040	0.040		<0.050-<1.000	0.052	NA	NA	<2	<2	<2

☐ NA
☐ J
☐ N
☐ a
☐ c
☐ d
☐ e
☐ f

CT&E Data.

Not analyzed.

Result is an estimate.

The analysis indicates the presence of an analyte for which there is presumptive evidence to make a "tentative identification".

The action level for DRPH is based on conversations with ADEC; a final action level has not yet been determined.

BTEX determined by 8260 method analysis.

The laboratory reported that the EPH pattern in this sample was not consistent with a middle distillate fuel.

The laboratory reported that 40.2 mg/kg of the EPH pattern in this sample was not consistent with a middle distillate fuel.

The laboratory reported that the EPH pattern in this sample was not consistent with an unweathered middle distillate fuel.

TABLE D-7. HEATED STORAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Heated Storage (Building 87) (SS13) Matrix: Sediment Units: mg/kg											
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples			Field Blanks			Lab Blanks
					SD01	SD02	SD03	AB01	EB03A	TB03	
Laboratory Sample ID Numbers					4212-2 4216-5	4219-1	4219-4	4173-9 4197-6	4211-1 4215-7	4211-2	4215 4216 4212
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	µg/L	µg/L	µg/L	mg/kg
SVOC 8270	0.200	1.00		<0.23-<3.5	<1.00	NA	NA	NA	<11	NA	<10 <0.200
PCBs											
Aroclor 1254	0.020	0.02-0.1	10	<0.020-<0.100	0.112	<0.1	<0.02	NA	<1	NA	<10 <0.020

☐ NA
☐ CT&E Data.
 Not analyzed.

TABLE D-7. HEATED STORAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Heated Storage (Building 87) (SS13) Matrix: Sediment Units: mg/kg										
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples			Field Blanks		
					SD06	SD07	SD08	AB01	EB01	TB02
Laboratory Sample ID Numbers										
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	4203-2	4203-3	4203-1	4173-9 4197-6	4203-8 4175-3	4179-5 4199-13
DRPH	4.00	4.00	500 ^a	9.55-1,150	37.3	32.5	115	μg/L	μg/L	μg/L
GRPH	0.400	0.400	100	<0.400-<9.0	4.45	2.7	8.94	NA	<200	NA
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-<1.500	0.043	<0.125	0.224			
Benzene	0.020	0.020-0.025	0.5	<0.020-<0.300	<0.020	<0.025	<0.020	<1	<1	<1
Toluene	0.020	0.020-0.025		<0.020-<0.300	<0.020	<0.025	<0.020	1.2	<1	<1
Ethylbenzene	0.020	0.020-0.025		<0.020-<0.300	<0.020	<0.025	0.045	<1	<1	<1
Xylenes (Total)	0.040	0.040-0.050		<0.040-<0.600	0.043	<0.050	0.179	<2	<2	<2
VOC 8010	0.020	0.020-0.025		<0.020-<0.300	<0.020	<0.025	<0.020	<1-9.8	<1-2.5	<1
Pesticides (8080)	0.001	0.020		<0.001-<0.100	NA	NA	<0.02	NA	<1J	NA
										<0.1-<1

CT&E Data.

Not analyzed.

Result is an estimate.

The action level for DRPH is based on conversations with ADEC; a final action level has not yet been determined.

NA
J
a

TABLE D-7. HEATED STORAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Heated Storage (Building 87) (SS13)			Matrix: Soil/Sediment Units: mg/kg		METALS ANALYSES						
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Environmental Samples				Field Blank		Lab Blanks
					S02	SD01			EB03A		
Laboratory Sample ID Numbers					4212-5	4212-2			4211-1		4212 4211
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg			µg/L		µg/L
Aluminum	0.35	2		1,500-25,000	3,000	2,900			<100		<100
Antimony	N/A	53-54		<7.8-<230	<54	<53			<100		<100
Arsenic	0.11	5.3-5.4		<4.9-8.5	<5.4	<5.3			<100		<100
Barium	0.024	1		27-390	43	23			<50		<50
Beryllium	N/A	2.7		<2.6-6.4	<2.7	<2.7			<50		<50
Cadmium	0.33	1-2.7		<3.0-<36	<2.7	2.7			<50		<50
Calcium	0.69	4		360-59,000	14,000	5,100			<200		<200
Chromium	0.066	1		<4.3-47	17	11			<50		<50
Cobalt	N/A	5.3-5.4		<5.1-12	<5.4	<5.3			<100		<100
Copper	0.045	1		<2.7-45	12	7.5			<50		<50
Iron	0.50	2		5,400-35,000	9,600	8,000			<100		<100
Lead	0.13	2		<5.1-22	33	33			<100		<100
Magnesium	0.96	4		360-7,400	4,900	2,400			<200		<200
Manganese	0.025	1		25-290	96	56			<50		<50
Molybdenum	N/A	2.7		<2.5-<11	<2.7	<2.7			<50		<50
Nickel	0.11	1		4.2-46	7.9	6.6			<50		<50
Potassium	23	100		<300-2,200	580	360			<5,000		<5,000
Selenium	1.2	53-54		<7.8-<170	<54	<53			<100		<100

☐ CT&E Data.
N/A Not available.

TABLE D-7. HEATED STORAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Heated Storage (Building 87) (SS13)				Matrix: Soil/Sediment Units: mg/kg		METALS ANALYSES					
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Environmental Samples				Field Blank		Lab Blanks
					S02	SD01			EB03A		
Laboratory Sample ID Numbers					4212-5	4212-2			4211-1		4212 4211
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg			µg/L		µg/L
Silver	0.53	2.7		<3-<110	<2.7	<2.7			<50		<50
Sodium	0.55	5		<160-680	60	38			<250		<250
Thallium	0.011	<0.27		<0.2-<1.2	<0.27	<0.27			<5		<5
Vanadium	0.036	1		6.3-59	9.1	14			<50		<50
Zinc	0.16	1		9.2-95	180	500			<50		<50

TABLE D-7. HEATED STORAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Heated Storage (Building 87) (SS13)		Matrix: Surface Water Units: µg/L		Environmental Samples			Field Blanks			Lab Blanks
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	SW01	SW02	AB01	EB03A	TB03	
Laboratory Sample ID Numbers					4212-1 4216-4	4213-1 4219-7	4173-9 4197-6	4211-1 4215-7	4211-2	4219 4216/4213 4212/4211 4197/4173
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L
DRPH	100	100		<200	5,760 ^a	1,300 ^{ad}	NA	<100	NA	<100
GRPH	20	20		<20	<20	6.9 ^a	NA	<20	NA	<20
BTEX (8020/8020 Mod.)										
Benzene	1	1	5	<1	<1	6.9	<1	<1	<1 ^c	<1
Toluene	1	1	1,000	<1	<1	<1	1.2	<1	<1 ^c	<1
Ethylbenzene	1	1	700	<1	<1	3.1	<1	<1	<1 ^c	<1
Xylenes (Total)	2	2	10,000	<2	<2	12.8	<2	<2	<2 ^c	<2
VOC 8260										
Benzene	1	1	5	<1	<1	6.4	<1	<1	<1	<1
Chloromethane	1	1		<1	<1	4.2	<1	<1	<1	<1
1,2-Dichloroethane	1	1	5	3U-3.2B	6.1B	9.1	1.6	<1	<1	<1
Ethylbenzene	1	1	700	<1	<1	2.5	<1	<1	<1	<1
Naphthalene	1	1		<1	<1	1.6	<1	<1	<1	<1

☐ CT&E Data.
☐ Not analyzed.
☐ NA
☐ B U a c d

The analyte was detected in the associated blank.
 Compound is not present above the concentration listed.
 Total petroleum hydrocarbons in these water samples exceed the 15 µg/L stated for fresh water in ADEC's Water Quality Criteria 18 AAC 70 (ADEC 1989).
 BTEX determined by 8260 method analysis.
 The laboratory reported that the EPH pattern in this sample was not consistent with an unweathered middle distillate fuel.

TABLE D-7. HEATED STORAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Heated Storage (Building 87) (SS13)				Matrix: Surface Water Units: µg/L							
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples			Field Blanks			Lab Blanks
					SW01	SW02		AB01	EB03A	TB03	
Laboratory Sample ID Numbers					4212-1 4216-4	4213-1 4219-7		4173-9 4197-6	4211-1 4215-7	4211-2	4215 4212/4211 4197/4173
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L		µg/L	µg/L	µg/L	µg/L
Tetrachloroethene	1	1	5	<1	<1	12		<1	<1	<1	<1
1,2,4-Trimethylbenzene	1	1		<1	<1	1.4		<1	<1	<1	<1
1,3,5-Trimethylbenzene	1	1		<1	<1	1.3		<1	<1	<1	<1
Xylenes (Total)	2	2	10,000	<2	<2	10.4		<2	<2	<2	<2
SVOC	10	10		<10	<10	<10		NA	<11	NA	<10

☐ CT&E Data.
☐ NA
 Not analyzed.

TABLE D-7. HEATED STORAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Heated Storage (Building 87) (SS13)				Matrix: Surface Water Units: µg/L										
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples				Field Blanks			Lab Blanks		
					SW05	SW06	SW07		AB01	EB02	TB02			
Laboratory Sample ID Numbers					4206-2	4206-3	4206-1		4197-6	4206-4	4199-13	4206 4199 4197		
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L		µg/L	µg/L	µg/L	µg/L		
DRPH	100	100		<200	334 ^a	196 ^a	349 ^a		NA	<100	NA	<100		
GRPH	20	20		<20	<20	<20	<20		NA	<20	NA	<20		
BTEX (8020/8020 Mod.)														
Benzene	1	1	5	<1	<1	<1	<1		<1	<1	<1	<1		<1
Toluene	1	1	1,000	<1	<1	<1	<1		1.2	<1	<1	<1		<1
Ethylbenzene	1	1	700	<1	<1	<1	3.2		<1	<1	<1	<1		<1
Xylenes (Total)	2	2	10,000	<2	<2	3.8	12.5		<2	<2	<2	<2		<2
VOC 8010														
1,2-Dichloroethane	1	1	5	3U-3.2B	3.3B	3.8B	<1		1.2	2.9	<1	<1		<1

☐ CT&E Data.

☐ N/A

☐ Not analyzed.

☐ The analyte was detected in the associated blank.

☐ Compound is not present above the concentration listed.

☐ Total hydrocarbons in these water samples exceed the 15 $\mu\text{g/L}$ stated for fresh water in ADEC's Water Quality Criteria 18 AAC 70 (ADEC 1989).

TABLE D-7. HEATED STORAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Heated Storage (Building 87) (SS13)				Matrix: Surface Water Units: µg/L		METALS ANALYSES: TOTAL (DISSOLVED)						Field Blank		Lab Blanks
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations		SW01	SW02						EB03A	
Laboratory Sample ID Numbers						4212-1	4213-1						4211-1	4213 4212 4211
ANALYSES	µg/L	µg/L	µg/L	µg/L		µg/L	µg/L					µg/L		µg/L
Aluminum	17.4	100		<100-350 (<100-340)		160 (<100)	210 (<100)					<100 (<100)		<100 (<100)
Antimony	N/A	100	6	<100 (<100)		<100 (<100)	<100 (<100)					<100 (<100)		<100 (<100)
Arsenic	5.3	100	50	<100 (<100)		<100 (<100)	<100 (<100)					<100 (<100)		<100 (<100)
Barium	1.2	50	2,000	<50-93 (<50-91)		82 (68)	82 (65)					<50 (<50)		<50 (<50)
Beryllium	N/A	50	4	<50 (<50)		<50 (<50)	<50 (<50)					<50 (<50)		<50 (<50)
Cadmium	1.7	50	5	<50 (<50)		<50 (<50)	<50 (<50)					<50 (<50)		<50 (<50)
Calcium	34.5	200		4,500-88,000 (4,100-86,000)		75,000 (71,000)	130,000 (130,000)					<200 (<200)		<200 (<200)
Chromium	3.29	100	100	<50 (<50)		<50 (<50)	<50 (<50)					<50 (<50)		<50 (<50)
Cobalt	N/A	100		<100 (<100)		<100 (<100)	<100 (<100)					<100 (<100)		<100 (<100)
Copper	2.3	50	1,300	<50 (<50)		<50 (<50)	<50 (<50)					<50 (<50)		<50 (<50)

☐ CT&E Data.
☐ N/A Not available.

TABLE D-7. HEATED STORAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Heated Storage (Building 87) (SS13)				Matrix: Surface Water Units: µg/L		METALS ANALYSES: TOTAL (DISSOLVED)						
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Environmental Samples							Lab Blanks
					SW01	SW02					Field Blank	
Laboratory Sample ID Numbers					4212-1	4213-1					EB03A	
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L					µg/L	µg/L
Iron	25	100		180-2,800 (<100-1,600)	1,600 (280)	12,000 (4,000)					<100 (<100)	<100-792 (792)
Lead	6.6	100	15	<100 (<100)	<100 (<100)	<100 (<100)					<100 (<100)	<100 (<100)
Magnesium	47.8	200		<5,000-53,000 (2,600-54,000)	26,000 (25,000)	46,000 (46,000)					<200 (<200)	<200 (<200)
Manganese	1.24	50		<50-510 (<50-120)	310 (240)	540 (500)					<50 (<50)	<50 (<50)
Molybdenum	N/A	50		<50 (<50)	<50 (<50)	<50 (<50)					<50 (<50)	<50 (<50)
Nickel	5.5	50	100	<50 (<50)	<50 (<50)	<50 (<50)					<50 (<50)	<50-82 (82)
Potassium	1,154	5,000		<5,000 (<5,000)	5,100 (<5,000)	17,000 (17,000)					<5,000 (<5,000)	<5,000 (<5,000)
Selenium	62.4	100	50	<100 (<100)	<100 (<100)	<100 (<100)					<100 (<100)	<100 (<100)
Silver	2.6	50	50	<50 (<50)	<50 (<50)	<50 (<50)					<50 (<50)	<50 (<50)
Sodium	27.7	250		8,400-410,000 (8,200-450,000)	100,000 (110,000)	92,000 (98,000)					<250 (<250)	<250 (<250)
Thallium	0.57	5	2	<5 (<5)	<5 (<5)	<5 (<5)					<5 (<5)	<5 (<5)

☐ CT&E Data.
N/A Not available.

TABLE D-7. HEATED STORAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Heated Storage (Building 87) (SS13)				Matrix: Surface Water Units: µg/L		METALS ANALYSES: TOTAL (DISSOLVED)						
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Environmental Samples					Field Blank		Lab Blanks
					SW01	SW02						
Laboratory Sample ID Numbers					4212-1	4213-1					4211-1	4213 4212 4211
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L					µg/L	µg/L
Vanadium	1.8	50		<50 (<50)	<50 (<50)	<50 (<50)					<50 (<50)	<50 (<50)
Zinc	8.2	50		<50-160 (<50)	<50 (<50)	<50 (<50)					<50 (<50)	<50 (<50)

AK-RISK\BARTER\APP-D\4109661203\TB LD-8.TBL

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The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined. DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC. Result is indicative of o-xylenes only.

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TABLE D-8. GARAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Garage (SS14)					Matrix: Soil Units: mg/kg									
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples					Field Blanks			Lab Blanks	
					S01	S02-2.5	S03-3 & S06-3 (Replicates)	S04-2	AB02	EB05	TB05			
Laboratory Sample ID Numbers					339	341	343	347	345 4301-7	315 4303-1	332 392 4303-5	375	#5-8249 #3&4-82493 #1&2-82493 4301	#6-82393 #1&2-82493 4301
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	μg/L	μg/L	μg/L	μg/L	mg/kg
Pesticides	0.002	0.02-0.5		<0.001-<0.100	<0.021-<0.5J	NA	NA	NA	NA	NA	<0.2-<10	NA	NA	<0.01-<0.5
PCBs	0.05-0.06	0.5-0.6	10	<0.020-<0.100	<0.5	<0.5	<0.6	<0.5	<0.5	NA	<10	NA	NA	<0.1-<0.5

☐ CT&E Data.
☒ F&B Data.
☐ Not analyzed.
☐ Result is an estimate.

TABLE D-8. GARAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Garage (SS14)				Matrix: Soil/Sediment Units: mg/kg		METALS ANALYSES										
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Environmental Samples						Field Blank		Lab Blanks			
					S04-2	SD01						EB05				
Laboratory Sample ID Numbers																4301 4303
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/L	µg/L
Aluminum	0.35	2		1,500-25,000	2,100	2,000									<100	<100
Antimony	N/A	52-53		<7.8-<230	<52	<53									<100	<100
Arsenic	0.11	5.2-5.3		<4.9-8.5	<5.2	<53									<100	<100
Barium	0.024	1		27-390	20	29									<50	<50
Beryllium	N/A	2.6-2.7		<2.6-6.4	<2.6	<2.7									<50	<50
Cadmium	0.33	2.6-2.7		<3.0-<36	<2.6	<2.7									<50	<50
Calcium	0.69	4		360-59,000	9,300J	4,400J									<200	<200
Chromium	0.066	1		<4.3-47	5.1	53									<50	<50
Cobalt	N/A	5.2-5.3		<5.1-12	<5.2	<5.3									<100	<100
Copper	0.045	1		<2.7-45	4.9	16									<50	<50
Iron	0.50	2		5,400-35,000	6,500	6,400									200	<100
Lead	0.13	2		<5.1-22	16J	231J									<100	<100
Magnesium	0.96	4		360-7,400	4,500J	1,900J									<200	<200
Manganese	0.025	1		25-290	63	50									<50	<50
Molybdenum	N/A	2.6-2.7		<2.5-<11	<2.6	<2.7									<50	<50
Nickel	0.11	1		4.2-46	5.1	5.6									<50	<50
Potassium	23	100		<300-2,200	300	310									<5,000	<5,000
Selenium	1.2	5.3-52		<7.8-<170	<52	<5.3									<100	<100

☐ CT&E Data.
☐ N/A Not available.
☐ J Result is an estimate.

TABLE D-8. GARAGE ANALYTICAL DATA SUMMARY (CONTINUED)

METALS ANALYSES													
Installation: Barter Island Site: Garage (SS14)			Matrix: Soil/Sediment Units: mg/kg		Bkgd. Range from 7 DEW Line Installations	Environmental Samples					Field Blank		Lab Blanks
Parameters	Detect. Limits	Quant. Limits	Action Levels	S04-2		SD01					EB05		
Laboratory Sample ID Numbers				4301-7	4301-4					4303-5		4301 4303	
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg					µg/L		µg/L	
Silver	0.53	2.6-2.7		<3-<110	<2.6J	<2.7J				<50		<50	
Sodium	0.55	5		<160-680	66	110				<250		<250-267	
Thallium	0.011	0.25-1.4		<0.2-<1.2	<0.25	<1.4				<5		<5	
Vanadium	0.036	1		6.3-59	7.5	6.6				<50		<50	
Zinc	0.16	1		9.2-95	35	200				<50		<50	

CT&E Data.
Result is an estimate.

☐ J

TABLE D-8. GARAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Garage (SS14)				Matrix: Soil Units: mg/kg								
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples			Field Blanks		Lab Blanks		
					2S05-2	2S07-4	2S08-3	AB03	EB07			
Laboratory Sample ID Numbers					1706 4616-1	1710 4616-3	1708 4616-2	1712	1715 1716 4616-9	#5-9593 #1&2-9693 #1&2-9493 4616	#5-9593 #3&4-9693 4616	
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	μg/L	μg/L	μg/L	mg/kg	
DRPH	4.00-20	50-200	500 ^a	9.55-1,150	<200J ^b	4,820 ^c	<50J ^b	NA	<200 ^d	<1,000J	<4.0	
GRPH	0.2-0.400	0.400-4	100	<0.4-<9	<4J ^b	358 ^c	<2J ^b	<50J ^b	<20 ^d	NA	<0.400-<2J	
RRPH (Approx.)	10-40	100-400	2,000 ^a	<480	<400	<120	<100	NA	<2,000	<2,000	NA	
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-<1.500	<0.15	5.8JN	<0.10					
Benzene	0.002-0.003	0.02-0.03	0.5	<0.020-<0.300	<0.03	1.4N	<0.02	<1	<1	NA	<0.02	
Toluene	0.002-0.003	0.02-0.03		<0.020-<0.300	<0.03	2.3J	<0.02	<3J	<2J	NA	<0.02	
Ethylbenzene	0.002-0.003	0.02-0.03		<0.020-<0.300	<0.03	0.4J	<0.02	<2J	<1	NA	<0.02	
Xylenes (Total)	0.004-0.006	0.04-0.06		<0.040-<0.600	<0.06	1.7J	<0.04	<5J	<2	NA	<0.04	
VOC 8260												
Benzene	0.020	0.020-0.400	0.5	<0.025-<0.500	0.072	<0.400	<0.020	NA	<1	<1	<0.020	
n-Butylbenzene	0.020	0.020-0.400		<0.025-<0.500	<0.040	3.10	<0.02	NA	<1	<1	<0.020	
sec-Butylbenzene	0.020	0.020-0.400		<0.025-<0.500	<0.040	1.28	<0.02	NA	<1	<1	<0.020	
cis-1,2-Dichloroethene	0.020	0.020-0.400		<0.025-<0.500	0.069	<0.400	<0.020	NA	<1	<1	<0.020	

□ CT&E Data.

■ F&B Data.

NA Not analyzed.

J Result is an estimate.

N The analysis indicates the presence of an analyte for which there is presumptive evidence to make a "tentative identification."

a The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined.

b DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

c This sample was analyzed by F&B also; DRPH and GRPH were detected at 8.460J^b and 700J^b mg/kg, respectively.

d This sample was analyzed by F&B also; DRPH and GRPH were detected at <1,000J^b and <50J^b μg/L, respectively.

TABLE D-8. GARAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Garage (SS14)		Matrix: Soil Units: mg/kg		Action Levels	Bkgd. Levels	Environmental Samples			Field Blanks		Lab Blanks	
Parameters	Detect. Limits	Quant. Limits				2S05-2	2S07-4	2S08-3	AB03	EB07		
Laboratory Sample ID Numbers						1706 4616-1	1710 4616-3	1708 4616-2	1712	1715 1716 4616-9	4616	4616
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	μg/L	μg/L	μg/L	mg/kg
Ethylbenzene	0.020	0.020-0.0400			<0.025-<0.500	<0.040	0.550	<0.020	NA	<1	<1	<0.020
Isopropylbenzene	0.020	0.020-0.400			<0.025-<0.500	<0.040	0.637	<0.020	NA	<1	<1	<0.020
p-Isopropyltoluene	0.020	0.020-0.400			<0.025-<0.500	<0.040	2.03	<0.02	NA	<1	<1	<0.020
Methylene Chloride	0.020	0.020-0.400		90	<0.025-<0.500	<0.040	<0.400	0.031B	NA	8.3	<1	<0.020
Naphthalene	0.020	0.020-0.400			<0.025-<0.500	<0.040	46.0	<0.02	NA	<1	<1	<0.020
n-Propylbenzene	0.020	0.020-0.400			<0.025-<0.500	<0.040	1.17	<0.02	NA	<1	<1	<0.020
Tetrachloroethene	0.020	0.020-0.400			<0.025-<0.500	<0.040	<0.400	0.023	NA	<1	<1	<0.020
Toluene	0.020	0.020-0.400			<0.025-<0.500	0.134	<0.400	<0.02	NA	1.7	<1	<0.020
1,2,4-Trimethylbenzene	0.020	0.020-0.400			<0.025-<0.500	<0.040	13.3	<0.02	NA	<1	<1	<0.020
1,3,5-Trimethylbenzene	0.020	0.020-0.400			<0.025-<0.500	<0.040	2.93	<0.02	NA	<1	<1	<0.020
Xylenes (Total)	0.040	0.040-0.800			<0.500-<1.000	<0.080	4.98	<0.04	NA	<2	<2	<0.040
Total BTEX			10		<0.125-<2.500	0.206	5.53	<0.10				

☐ CT&E Data.
☐ NA Not analyzed.
☐ B The analyte was found in the associated blank.

TABLE D-8. GARAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Garage (SS14)				Matrix: Sediment Units: mg/kg							
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Sample			Field Blanks		Lab Blanks	
					SD01			AB02	EB05	TB05	
Laboratory Sample ID Numbers					349 4301-4			315 4303-1	332 392 4303-5	375	#3&4-82493 #1&2-82493 4303
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg			μg/L	μg/L	μg/L	mg/kg
DRPH	70	700	500 ^a	9.55-1,150	5,100 ^b			NA	<1,000 ^b	NA	<50J
GRPH	0.2	2	100	<0.400-<9.0	390J ^b			<100J ^b	<100J ^b	<50J ^b	<2J
RRPH (Approx.)	70	700	2,000 ^a	<480	13,000			NA	<1,000	NA	<100
BTEX (8020/8020 Mod.)				10 Total BTEX	<0.250-<1.500						
Benzene	0.03	0.3	0.5	<0.020-<0.300	<0.3			<1	<1	<1	<0.02J
Toluene	0.03	0.3		<0.020-<0.300	1			<1	<1	<1	<0.02
Ethylbenzene	0.03	0.3		<0.020-<0.300	11			<1	<1	<1	<0.02
Xylenes (Total)	0.06	0.6		<0.040-<0.600	47J			<2	<2	<2	<0.04
HVOC (8010 Mod.)	0.03	0.3		<0.5J	<0.3			<1	<1	<1	<0.02J
VOC 8260											
n-Butylbenzene	0.020	0.200		<0.025-<0.500	4.22			<1	<1	NA	<1
sec-Butylbenzene	0.020	0.200		<0.025-<0.500	1.24			<1	<1	NA	<1
tert-Butylbenzene	0.020	0.200		<0.025-<0.500	0.256			<1	<1	NA	<1
Ethylbenzene	0.020	0.200		<0.025-<0.500	0.728			<1	<1	NA	<1
Isopropylbenzene	0.020	0.200		<0.025-<0.500	0.681			<1	<1	NA	<1

□ CT&E Data.

■ F&B Data.

■ Not analyzed.

Result is an estimate.

The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined.

DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

TABLE D-8. GARAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Garage (SS14)		Matrix: Sediment Units: mg/kg									
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Sample			Field Blanks			Lab Blanks
					SD01			AB02	EB05	TB05	
Laboratory Sample ID Numbers					349 4301-4			315 4303-1	332 392 4303-5	375	4301
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg			µg/L	µg/L	µg/L	mg/kg
p-Isopropyltoluene	0.020	0.200		<0.025-<0.500	2.47			<1	<1	NA	<0.020
Naphthalene	0.020	0.200		<0.025-<0.500	14.9			<1	<1	NA	<0.020
n-Propylbenzene	0.020	0.200		<0.025-<0.500	0.918			<1	<1	NA	<0.020
1,2,4-Trimethylbenzene	0.020	0.200		<0.025-<0.500	14.7			<1	<1	NA	<0.020
1,3,5-Trimethylbenzene	0.020	0.200		<0.025-<0.500	9.32			<1	<1	NA	<0.020
Xylenes (Total)	0.040	0.400		<0.050-<1.000	10.3			<2	<2	NA	<0.040
SVOC 8270											
2-Methylnaphthalene	0.200	2.20		<0.230-<3.50	14.5			NA	<11	NA	<0.200
Phenanthrene	0.200	2.20		<0.230-<3.50	4.79			NA	<11	NA	<0.200
bis(2-Ethylhexyl) Phthalate	0.200	2.20	50	<0.230-<3.50	4.6			NA	<11	NA	<0.200
Fluoranthene	0.200	2.20		<0.230-<3.50	2.28			NA	<11	NA	<0.200
Isophorone	0.200	2.20		<0.230-<3.50	1.49J			NA	<11	NA	<0.200
Naphthalene	0.200	2.20		<0.230-<3.50	7.73			NA	<11	NA	<0.200

☐ CT&E Data.
☐ NA Not analyzed.
☐ J Results in an estimate.

TABLE D-8. GARAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Garage (SS14)		Matrix: Surface Water Units: µg/L									
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Sample				Field Blanks		
					SW01				AB02	EB05	TB05
Laboratory Sample ID Numbers					336/333 4301-1				315 4303-1	332 392 4303-5	375
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L				µg/L	µg/L	µg/L
DRPH	100	1,000		<200	<1,000 ^b				NA	<1,000 ^b	NA
GRPH	10	100		<20	<100 ^b				<100 ^b	<100 ^b	<50J ^b
RRPH (Approx.)	200	2,000		NA	<2,000				NA	<1,000	NA
BTEX (8020/8020 Mod.)											
Benzene	0.1	1	5	<1	<1				<1	<1	<1
Toluene	0.1	1	1,000	<1	<1				<1	<1	<1
Ethylbenzene	0.1	1	700	<1	<1				<1	<1	<1
Xylenes (Total)	0.2	2	10,000	<2	BJ				<2	<2	<2
HVOC (8010 Mod.)	0.1	1		NA	<1				<1	<1	<1
SVOC 8270	10	25		<10	<25				NA	<11	NA
Pesticides	0.2	2		<1	<2J-<50J				NA	<0.2-<10	NA

□ CT&E Data.

■ F&B Data.

NA Not analyzed.

J Result is an estimate.

DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

TABLE D-8. GARAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Garage (SS14)		Matrix: Surface Water Units: µg/L		METALS ANALYSES: TOTAL (DISSOLVED)		Environmental Sample						Field Blank		Lab Blanks
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	SW01								EB05	
Laboratory Sample ID Numbers					4301-1								4303-5	4301 4305
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L								µg/L	µg/L
Aluminum	17.4	100		<100-350 (<100-340)	<100 (<100)								<100	<100 (<100)
Antimony	N/A	100	6	<100 (<100)	<100 (<100)								<100	<100 (<100)
Arsenic	5.3	100	50	<100 (<100)	<100 (<100)								<100	<100 (<100)
Barium	1.2	50	2,000	<50-93 (<50-91)	74 (62)								<50	<50 (<50)
Beryllium	N/A	50	4	<50 (<50)	<50 (<50)								<50	<50 (<50)
Cadmium	1.7	50	5	<50 (<50)	<50 (<50)								<50	<50 (<50)
Calcium	34.5	100		4,500-86,000 (4,100-86,000)	105,000 (110,000)								<200	<200-378 (378)
Chromium	3.29	50	100	<50 (<50)	<50 (<50)								<50	<50 (<50)
Cobalt	N/A	100		<100 (<100)	<100 (<100)								<100	<100 (<100)
Copper	2.3	50	1,300	<50 (<50)	<50 (<50)								<50	<50 (<50)
Iron	25	100		180-2,800 (<100-1,600)	6,000 (5,500)								200	<100 (<100)

☐ CT&E Data.
N/A Not available.

TABLE D-8. GARAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Garage (SS14)			Matrix: Surface Water Units: µg/L		METALS ANALYSES: TOTAL (DISSOLVED)							Field Blank		Lab Blanks
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	SW01							EB05		
Laboratory Sample ID Numbers					4301-1							4303-5		4301 4305
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L							µg/L		µg/L
Lead	6.6	100	15	<100 (<100)	<100J (<100)J							<100		<100 (<100)
Magnesium	47.8	100		<5,000-53,000 (2,600-54,000)	34,000 (32,000)							<200		<200 (<200)
Manganese	1.24	50		<50-210 (<50-120)	490 (460)							<50		<50 (<50)
Molybdenum	N/A	50		<50 (<50)	<50 (<50)							<50		<50 (<50)
Nickel	5.5	50	100	<50 (<50)	<50 (<50)							<50		<50 (<50)
Potassium	1,154	100		<5,000 (<5,000)	9,100 (8,200)							<5,000		<5,000 (<5,000)
Selenium	62.4	100	50	<100 (<100)	<100 (<100)							<100		<100 (<100)
Silver	2.6	50	50	<50 (<50)	<50 (<50)							<50		<50 (<50)
Sodium	27.7	100		8,400-410,000 (8,200-450,000)	130,000 (140,000)							<250		<250-267 (<250)
Thallium	0.57	5	2	<5 (<5)	<5 (<5)							<5		<5 (<5)
Vanadium	1.8	50		<50 (<50)	<50 (<50)							<50		<50 (<50)

☐ CT&E Data.
☐ N/A
☐ J Not available.
 Result is an estimate.

TABLE D-8. GARAGE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Garage (SS14)			Matrix: Surface Water Units: µg/L		METALS ANALYSES: TOTAL (DISSOLVED)										
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Environmental Sample							Field Blank			Lab Blanks
					SW01									EB05	
Laboratory Sample ID Numbers					4301-1								4303-5		4301 4305
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L								µg/L		µg/L
Zinc	8.2	50		<50-160 (<50)	<50 (<50)								<50		<50 (<50)

TABLE D-9. WEATHER STATION BUILDING ANALYTICAL DATA SUMMARY

Installation: Barter Island Site: Weather Station Building (SS15)		Matrix: Soil Units: mg/kg											
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples					Field Blanks			Lab Blanks
Laboratory Sample ID Numbers					S01	S02-2	S03	S04	S05	AB01	EB01	TB01	
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	4198-1	4198-4	4178-1 4198-5	4198-6	4198-7	4173-9 4197-6	4175-3 4203-8	4173-10 4179-7	#5-9693 #1&2-9493 4203 4175
DRPH	4.00	4.00	500 ^a	9.55-1,150	6,100	8,420	2,090	214	107 ^c	NA	<200	NA	<200-<1,000J
GRPH	0.400	0.400	100	<0.400-<9.0	216	1,020	11.1	1.46	5.3	NA	<20	NA	<20-<50J
BTEX (8020/ 8020 Mod.)			10 Total BTEX	<0.250-<1.500	0.459	3.567	0.154	0.023	0.547				
Benzene	0.020	0.020	0.5	<0.020-<0.300	<0.020	<0.020	<0.020	<0.020	<0.020	<1	<1	<1	<1
Toluene	0.020	0.020		<0.020-<0.300	<0.020	0.137	<0.020	<0.020	<0.020	1.2	<1	<1	<1
Ethylbenzene	0.020	0.020		<0.020-<0.300	0.196	1.09	0.056	<0.020	0.071	<1	<1	<1	<1
Xylenes (Total)	0.040	0.040		<0.040-<0.600	0.263	2.34	0.098	0.023 ^d	0.476	<2	<2	<2	<2
VOC 8260													
Naphthalene	0.020	0.020		<0.025-<0.500	NA	NA	0.063	NA	NA	<1	<1J	<1	<1
1,3,5-Trimethylbenzene	0.020	0.020		<0.025-<0.500	NA	NA	0.084	NA	NA	<1	<1J	<1	<1
Xylenes (Total)	0.020	0.020		<0.050-<1.000	NA	NA	0.079	NA	NA	<2	<2J	<2	<2

☐ CT&E Data.
☐ Not analyzed.
☐ Result is an estimate.
☐ The action level for DRPH is based on conversations with ADEC; a final action level has not yet been determined.
☐ The laboratory reported that 6.39 mg/kg of the EPH pattern in this sample was not consistent with a middle distillate fuel.
☐ Result is indicative of p & m xylenes only.

TABLE D-9. WEATHER STATION ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barber Island Site: Weather Station Building (SS15)		Matrix: Soil Units: mg/kg									
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	2506	2507-2	2508-2	Field Blanks			Lab Blanks
Laboratory Sample ID Numbers								AB01	EB06	TB01	
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	µg/L	µg/L	µg/L	mg/kg
DRPH	5-6	50-60	500 ^a	9.55-1,150	<50J ^b	<50J ^b	<60J ^b	NA	1688 1690	4173-10 4179-7	#5-9693 #182-9493 4203 4175
GRPH	0.2	2	100	<0.400-<9.0	<2J ^b	4J ^b	<2J ^b	NA	<50J ^b	NA	<4.00
RRPH (Approx.)	10-12	100-120	2,000 ^a	<480	<100	<100	<120	NA	<2,000	NA	<0.400-<2J
BTEX (8020/ 8020 Mod.)			10 Total BTEX	<0.250-<1.500	<0.10	0.3J	<0.10				NA
Benzene	0.002-0.020	0.02	0.5	<0.020-<0.300	<0.02	<0.02	<0.02	<1	<1	<1	<0.020
Toluene	0.002-0.020	0.02		<0.020-<0.300	<0.02	<0.02	<0.02	1.2	<4J	<1	<0.020
Ethylbenzene	0.002-0.020	0.02		<0.020-<0.300	<0.02	0.1J	<0.02	<1	<2J	<1	<0.020
Xylenes (Total)	0.004-0.040	0.04		<0.040-<0.600	<0.04	0.2J	<0.04	<2	<5J	<2	<0.040

☐ CT&E Data.
☒ F&B Data.
☒ Not analyzed.
☒ Result is an estimate.
☒ The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined.
☒ DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

TABLE D-10. WHITE ALICE FACILITY ANALYTICAL DATA SUMMARY

Installation: Barter Island Site: White Alice Facility (SS16)		Matrix: Soil Units: mg/kg											
Parameters	Detect Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples						Field Blanks		
					S01	S02 Composite	S03 Composite	S04 Composite	S05 Composite	S06 & S07 (Replicates)	AB02	EB05	TB05
Laboratory Sample ID Numbers					320	321	370	371	372	373	374	392 332	375
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	μg/L	μg/L
DRPH	5	50	500 ^a	9.55-1,150	NA	NA	<50.0 ^b	<50.0 ^b	<50.0 ^b	<50.0 ^b	NA	<1,000 ^b	NA
RRPH (Approx.)	10-20	100-200	2,000 ^a	<480	NA	NA	<100	<100	<100	<200	<200	<1,000	NA
PCBs													
Aroclor 1254	0.01-0.05	0.1-0.5	10	<0.020-0.100	52J	<0.5	<0.1	<0.1	<0.1	20J	30J	<10	NA
													NA
													<0.1-0.5

CT&E Data.

F&B Data.

Not analyzed.

Result is an estimate.

The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined.

DRPH concentrations reported for these samples are equivalent to diesel range organics (DRO) as defined by ADEC.

☐ CT&E Data.
☒ F&B Data.
☐ Not analyzed.
☐ Result is an estimate.
☐ The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined.
☐ DRPH concentrations reported for these samples are equivalent to diesel range organics (DRO) as defined by ADEC.

TABLE D-10. WHITE ALICE FACILITY ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: White Alice Facility (SS16)		Matrix: Soil Units: mg/kg		Environmental Samples							Field Blank	Lab Blank
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	2S08-1.5	2S09	2S10 & 2S11 (Replicates)	2S12	2S13-1.5		EB06	
Laboratory Sample ID Numbers					1671	1672	1673	1674	1675	1676	1688 1690	#5-9493
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	µg/L	mg/kg
PCBs												
Aroclor 1254	0.05	0.5	10	<0.020-<0.100	<0.5	8.7J	<0.5	<0.5	<0.5	<0.5	<2	<0.1-<2

☐ CT&E Data.
☒ F&B Data.
 Result is an estimate.

TABLE D-11. POL TANKS ANALYTICAL DATA SUMMARY

Installation: Barter Island Site: POL Tanks (S117)		Matrix: Soil Units: mg/kg											
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples				Field Blanks			Lab Blanks	
					S02	S03	S04 & S06 (Replicates)	S05-1.5	AB02	EB04	TB04		
Laboratory Sample ID Numbers					1292	1294	1296 4302-4	1300 4302-5	315 4303-1	311 4302-10	1346 4302-9	#384-82493 #384-83193 4303 4302	#6-83193 #1&2-83193 4302
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/L	mg/L	μg/L	μg/L	mg/kg
DRPH	8	80	500*	9.55-1,150	<80 ^b	540 ^b	730 ^b	1,670 ^a	NA	NA	NA	NA	<70J
GRPH	0.2-0.4	2.4	100	<0.400-<9.0	<2J ^b	255J ^b	52J ^b	85J ^b	<100J ^b	<100J ^b	<100J ^b	<100J	<2J
RRPH (Approx.)	10-30	100-300	2,000*	<480	<100	<300	<100	<140	NA	NA	NA	NA	<100
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-<1.500	<0.10	1.0	0.4J	1.0J	<0.35				
Benzene	0.002-0.006	0.02-0.06	0.5	<0.020-<0.300	<0.02	<0.08	<0.03	<0.02	<1	<1	<1	<1	<0.02
Toluene	0.002-0.006	0.02-0.06		<0.020-<0.300	<0.02	1.0	<0.03	<0.02	<1	<1	<1	<1	<0.02
Ethylbenzene	0.002-0.006	0.02-0.06		<0.020-<0.300	<0.02	<0.02	0.1	0.4	<1	<1	<1	<1	<0.02
Xylenes (Total)	0.004-0.020	0.04-0.2		<0.040-<0.600	<0.04	<0.04	0.3J	0.6J	<2	<2	<2	<2	<0.04
VOC 8260													
1,2,4-Trimethylbenzene	0.020	0.240-0.250		<0.025-<0.500	NA	NA	0.511	0.344	NA	<1	<1	<1	<0.020
SVOC 8270	0.200	2.30-2.50		<0.23-<3.5	NA	NA	<2.50	<2.30	NA	<25-75	NA	<10	<0.200

CT&E Data.

F&B Data.

Not analyzed.

Result is an estimate.

The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined.
 DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

TABLE D-12. FUEL TANKS ANALYTICAL DATA SUMMARY

Installation: Barter Island Site: Fuel Tanks (ST18)					Matrix: Soil Units: mg/kg											
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples							Field Blanks			Lab Blanks	
					S01	S02	S03 & S10 (Replicates)	S04	S05	S06	S07	AB02	EB04	TB04		
Laboratory Sample ID Numbers					1302	1304 4302-7	1306 1308	1310	1312	1314	1322	315 4303-1	311 4302-10	1346 4301-9	#3&4-82493 #3&4-83192 4303 4302 4301	#1&2-83193 #6-83193 4302
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	μg/L	μg/L	μg/L	μg/L	mg/kg
DRPH	5-10	50-100	500 ^a	9.55-1,150	<80 ^b	<70 ^b	<80 ^b	<100 ^b	<80 ^b	<50 ^b	<70 ^b	NA	NA	NA	NA	<70J
GRPH	0.2	2	100	<0.400-<9.0	<2J ^b	<2J ^b	<2J ^b	<2J ^b	<2J ^b	<2J ^b	<2J ^b	<100J ^b	<100J ^b	<100J ^b	<50J-<100J	<2J
RRPH (Approx.)	10-11	100-110	2,000 ^a	<480	<100	<100	<100	<100	<100	190	<110	NA	NA	NA	NA	<100
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-<1.500	<0.10	<0.10	<0.10	<0.10	<0.10	<0.10	<0.10					
Benzene	0.002	0.02	0.5	<0.020-<0.300	<0.02	<0.02	<0.02	<0.02	<0.02	<0.02	<0.02	<1	<1	<1	<1	<0.02
Toluene	0.002	0.02		<0.020-<0.300	<0.02	<0.02	<0.02	<0.02	<0.02	<0.02	<0.02	<1	<1	<1	<1	<0.02
Ethylbenzene	0.002	0.02		<0.020-<0.300	<0.02	<0.02	<0.02	<0.02	<0.02	<0.02	<0.02	<1	<1	<1	<1	<0.02
Xylenes (Total)	0.004	0.04		<0.040-<0.600	<0.04	<0.04	<0.04	<0.04	<0.04	<0.04	<0.04	<2	<2	<2	<2	<0.04
VOC 8260																
Ethylbenzene	0.020	0.020		<0.025-<0.500	NA	0.020	NA	NA	NA	NA	NA	<1	<1	<1	<1	<0.020
1,2,4-Trimethylbenzene	0.020	0.020		<0.025-<0.500	NA	0.128	NA	NA	NA	NA	NA	<1	<1	<1	<1	<0.020
1,3,5-Trimethylbenzene	0.020	0.020		<0.025-<0.500	NA	0.048	NA	NA	NA	NA	NA	<1	<1	<1	<1	<0.020
Xylenes (Total)	0.040	0.040	10	<0.500-<1.00	NA	0.191	NA	NA	NA	NA	NA	<2	<2	<2	<2	<0.040
SVOC	0.200	0.220		<0.230-<3.50	NA	<0.220	NA	NA	NA	NA	NA	NA	<25-75	NA	<10	<0.200
TOC				32,000-199,000	NA	1,330	NA	NA	NA	NA	NA	NA	<5,000	NA	<5,000	NA

CT&E Data.

F&B Data.

Not analyzed.

Result is an estimate.

The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined. DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

☐ CT&E Data.
☒ F&B Data.
☐ Not analyzed.
☐ Result is an estimate.

TABLE D-12. FUEL TANKS ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Fuel Tanks (ST18)														
Matrix: Soil/Sediment Units: mg/kg														
Parameters	Detect Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples					Field Blanks				Lab Blanks
					S08-0.75	S09-1.5	2S11-2	2S12-2	SD01	AB02	EB04	EB08	TB04	
Laboratory Sample ID Numbers					1316	1318	1754 4616-6	1752	1320 4302-6	315 4303-1	311 4302-10	1719 1720 4616-13	1346 4302-9	#1&2-83193 #3&4-83193 #3&4-82493 #6-83193 #1&2-9883 #5-9693 4616 4303 4302
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	μg/L	μg/L	μg/L	mg/kg
DRPH	5-10	50-100	500*	9.55-1,150	<100 ^b	<60 ^b	490 ^b	<50 ^b	<70 ^b	NA	NA	<1,000 ^b	NA	<70J
GRPH	0.1-0.2	1-2	100	<0.400-≤9.0	NA	<2 ^b	6 ^b	<1 ^b	<2 ^b	<100 ^b	<100 ^b	<50 ^b	<100 ^b	<2J
RRPH (Approx.)	10-15	100-150	2,000*	<480	<150	<100	<100	<100	<110	NA	NA	<2,000	NA	<100
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-≤1,500	NA	<0.10	0.8J	<0.10	<0.10					
Benzene	0.002	0.02	0.5	<0.020-≤0.300	NA	<0.02	0.1J	<0.02	<0.02	<1	<1	<1	<1	<0.02
Toluene	0.002	0.02		<0.020-≤0.300	NA	<0.02	0.1	<0.02	<0.02	<1	<1	<1J	<1	<0.02
Ethylbenzene	0.002	0.02		<0.020-≤0.300	NA	<0.02	0.2	<0.02	<0.02	<1	<1	<1	<1	<0.02
Xylenes (Total)	0.004	0.04		<0.040-≤0.600	NA	<0.04	0.4J	<0.04	<0.04	<2	<2	<2	<2	<0.04
VOC 8260														
Benzene	0.020	0.020	0.5	<0.025-≤0.500	NA	NA	0.091	NA	<0.020	<1	<1	<1	<1	<0.020
Xylenes (Total)	0.020	0.040		<0.050-≤1,000	NA	NA	0.063	NA	<0.040	<2	<2	<2	<2	<0.040
SVOC 8270	0.20	<0.210-0.214		<0.230-≤3.50	NA	NA	<0.214-≤1.67	NA	<0.210	NA	<25-75	NA	NA	<0.200
TOC				32,000-199,000	NA	NA	NA	NA	3,400	NA	<5,000	NA	NA	NA

CT&E Data.

F&B Data.

Not analyzed.

Result is an estimate.

The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined. DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

TABLE D-12. FUEL TANKS ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Fuel Tanks (ST18)		Matrix: Surface Water Units: µg/L		Environmental Sample				Field Blanks			Lab Blanks	
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	SW01			AB02	EB04	TB04		
Laboratory Sample ID Numbers					317 4302-8			315 4303-1	311 4302-10	1346 4302-9	#3&4-83193 #3&4-82493 4303 4302	
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L			µg/L	µg/L	µg/L	µg/L	
GRPH	10	100		<20	<100J ^b			<100J ^b	<100J ^b	<100J ^b	<100J	
BTEX (8020/8020 Mod.)												
Benzene	0.1	1	5	<1	<1			<1	<1	<1	<1	
Toluene	0.1	1	1,000	<1	<1			<1	<1	<1	<1	
Ethylbenzene	0.1	1	700	<1	<1			<1	<1	<1	<1	
Xylenes (Total)	0.2	2	10,000	<2	<2			<2	<2	<2	<2	
VOC 8260	1	1		<1-3.2B	<1-3U			<1-6.3	<1-6.0	<1	<1	
SVOC 8270	10	10		<10	<10			NA	<25-75	NA	<10	
TOC	5,000	5,000		<5,000-12,700	36,300			NA	<5,000	NA	<5,000	
TSS	100	200		<30,000-8,000	8,000J			NA	NA	NA	<200	
TDS	10,000	10,000		<352,000-328,000	711,000J			NA	NA	NA	12,000	

□ CT&E Data.

■ F&B Data.

■ NA

■ Not analyzed.

■ The analyte was detected in the associated blank.

■ Result is an estimate.

■ Compound is not present above the concentration listed.

■ GRPH concentrations reported for these samples are equivalent to gasoline range organics (GRO) as defined by ADEC.

TABLE D-13. OLD DUMP SITE ANALYTICAL DATA SUMMARY

Installation: Barter Island Site: Old Dump Site (LF19)				Matrix: Soil Units: mg/kg								
Parameters	Detect Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples			Field Blanks			Lab Blanks	
					S01-1.5	S02-2	S03	AB02	EB05	TB05		
Laboratory Sample ID Numbers					353 4301-9	355	357	315 4303-1	392/332 4303-5	375	#384-82493 #182-82493 4303	#6-82393 #182-82493 4301
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	μg/L	μg/L	μg/L	μg/L	mg/kg
DRPH	5.0-12	50-120	500 ^a	9.55-1,150	<60J ^a	<50J ^a	<50J ^a	NA	<1,000 ^a	NA	NA	<50J
GRPH	0.2-4.6	2-4.6	100	<0.400-<9.0	<46J ^a	<8J ^a	<5J ^a	<100J ^a	<100J ^a	<50J ^a	<100J	<2J
RRPH (Approx.)	10	100	2,000 ^a	<480	<100	<100	<100	NA	<1,000	NA	NA	<100
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-<1,500	0.08R	<2.14	<1.54					
Benzene	0.002-0.005	0.02-0.05	0.5	<0.020-<0.300	<0.02R	<0.02	<0.02	<1	<1	<1	<1	<0.02J
Toluene	0.002-0.005	0.02-0.05		<0.020-<0.300	<0.02R	<0.02	<0.02	<1	<1	<1	<1	<0.02
Ethylbenzene	0.002-0.005	0.02-0.05		<0.020-<0.300	0.06R	<0.6	<0.2	<1	<1	<1	<1	<0.02
Xylenes (Total)	0.004-0.010	0.04-0.1		<0.040-<0.600	<0.04R	<1.3	<1.3	<2	<2	<2	<1	<0.04
HVOC (8010 Mod.)	0.002-0.005	0.02-0.05		<0.3J	<0.02R	<0.02	<0.02	<1	<1	<1	<2	<0.02J
VOC 8280	0.020	0.020		<0.025-<0.050	<0.020	NA	NA	<1-6.3	<1-3.2	NA	<1	<0.020
SVOC 8270	0.200	0.200-1.00		<0.230-<3.50	<0.200-<1.00	NA	NA	NA	<11	NA	<10	<0.200
Pesticides	0.002-0.05	0.02-0.5		<0.001-<0.100	<0.02J-<0.5J	NA	NA	NA	<0.2-<10	NA	NA	<0.01-<0.5
PCBs	0.05-0.12	0.5-1.2	10	<0.020-<0.100	<0.5	<0.5	<0.5	NA	<10	NA	NA	<0.1-<0.5

CT&E Data.

F&B Data.

Not analyzed.

Result is an estimate.

Result has been rejected.

The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined.

DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

☐ CT&E Data.
☒ F&B Data.
☐ Not analyzed.
☐ Result is an estimate.
☐ Result has been rejected.

TABLE D-13. OLD DUMP SITE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Old Dump Site (LF19)		Matrix: Soil Units: mg/kg									
Parameters	Detect Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples		Field Blanks			Lab Blanks	
Laboratory Sample ID Numbers					S04-2.5 & S06-2.5 (Replicates)	S05-2	AB02	EB05	TB05	#3&4-82493 #1&2-82493 4303	#6-82393 #1&2-82493 4301
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	359 4301-8	361	315 4303-1	392/332 4303-5	375	μg/L	μg/L
DRPH	5.0-12	50-120	500*	9.55-1,150	<50J ^b	<120J ^b	NA	<1,000 ^b	NA	NA	<50J
GRPH	0.2-0.5	2-5	100	<0.4-49	<3J ^b	<5J ^b	<100J ^b	<100J ^b	<50J ^b	<100J	<2J
RRPH (Approx.)	10	100	2,000*	<480	<100	1,800	NA	<1,000	NA	NA	<100
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-1,500	<1.24	<0.81					
Benzene	0.002-0.005	0.02-0.05	0.5	<0.020-0.300	<0.02	<0.05	<1	<1	<1	<1	<0.02J
Toluene	0.002-0.005	0.02-0.05		<0.020-0.300	<0.02	<0.05	<1	<1	<1	<1	<0.02
Ethylbenzene	0.007-0.05	0.07-0.5		<0.020-0.300	<0.5	<0.07	<1	<1	<1	<1	<0.02
Xylenes (Total)	0.05-0.07	0.5-0.7		<0.040-0.600	<0.7	<0.5	<2	<2	<2	<1	<0.04
HVOC (8010 Mod.)	0.002-0.005	0.02-0.05		<0.5J	<0.02	<0.05	<1	<1	<1	<2	<0.020J
VOC 8260	0.020	0.020		<0.025-0.050	<0.020	NA	<1-6.3	<1-3.2	NA	<1	<0.020
SVOC 8270	0.200	0.20-0.710		<0.23-3.5	<0.210-0.710	NA	NA	<11	NA	<10	<0.200
Pesticides	0.002-0.05	0.02-0.5		<0.001-0.100	<0.02J-0.5J	NA	NA	<0.2-10	NA	NA	<0.01-0.5
PCBs	0.05-0.12	0.5-1.2	10	<0.020-0.100	<0.5	<1.20	NA	<10	NA	NA	<0.1-0.5
TOC				32,900-199,000	4,950	NA	NA	5,000J	NA	<5,000	NA

□ CT&E Data.

■ F&B Data.

■ Not analyzed.

■ Result is an estimate.

The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined.
DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

□ ■ NA J a b

TABLE D-13. OLD DUMP SITE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Old Dump Site (LF19)				Matrix: Soil Units: mg/kg		METALS ANALYSES									
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Environmental Samples				Field Blank		Lab Blanks				
					S01-1.5	S04-2.5 & S06-2.5 (Replicates)				EB05					
Laboratory Sample ID Numbers															
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	4301-9	4301-8	4301-10				4303-5	4303-5			
Aluminum	0.35	2		1,500-25,000	2,500	2,900	3,500				μg/L	μg/L			
Antimony	N/A	49-52		<7.8-<230	<51	<52	<49				<100	<100			
Arsenic	0.11	4.9-5.2		<4.9-8.5	<5.1	<5.2	<4.9				<100	<100			
Barium	0.024	1		27-390	16	23	31				<50	<50			
Beryllium	N/A	2.5-2.6		<2.6-6.4	<2.6	<2.6	<2.5				<50	<50			
Cadmium	0.33	2.5-2.6		<3.0-<36	<2.6	<2.6	<2.5				<50	<50			
Calcium	0.69	4		360-59,000	8,400	9,800J	12,000				<200	<200			
Chromium	0.066	1		<4.3-47	5.0	5.4	7.4				<50	<50			
Cobalt	N/A	4.9-5.2		<5.1-12	<5.1	<5.2	<4.9				<100	<100			
Copper	0.045	1		<2.7-45	7.9	5.4	4.8				<50	<50			
Iron	0.50	2		5,400-35,000	7,500	8,000	8,800				200	<100			
Lead	0.13	4.9-5.2		<5.1-22	<51	<5.2J	<4.9				<100	<100			
Magnesium	0.96	4		360-7,400	1,500	1,700J	1,900				<200	<200			
Manganese	0.025	1		25-290	160	120	130				<50	<50			
Molybdenum	N/A	2.5-2.6		<2.5-<11	<2.6	<2.6	<2.5				<50	<50			
Nickel	0.11	1		4.2-46	6.9	6.8	7.8				<50	<50			
Potassium	23	100		<300-2,200	340	390	410				<5,000	<5,000			

☐ CT&E Data.
☐ N/A Not available.
☐ J Result is an estimate.

TABLE D-13. OLD DUMP SITE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Old Dump Site (LF19)			Matrix: Soil Units: mg/kg		METALS ANALYSES							
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Environmental Samples				Field Blank		Lab Blanks	
					S01-1.5	S04-2.5 & S06-2.5 (Replicates)						EB05
Laboratory Sample ID Numbers					4301-9	4301-8	4301-10					4303 4301
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg					μg/L
Selenium	1.2	49-52		<7.8-<170	<51	<52	<49				<100	<100
Silver	0.53	2.5-2.6		<3-<110	<2.6	<2.6J	<2.5				<50	<50
Sodium	0.55	5		<160-680	53	60	70				<250	<250-267
Thallium	0.011	2.5-2.7		<0.2-<1.2	<0.25	<0.27	<0.26				<5	<5
Vanadium	0.036	1		6.3-59	5.8	6.6	9.3				<50	<50
Zinc	0.16	1		9.2-95	20	20	21				<50	<50

CT&E Data.
Result is an estimate.

☐ J

TABLE D-13. OLD DUMP SITE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Old Dump Site (LF19)		Matrix: Soil Units: mg/kg									
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples				Field Blanks		Lab Blanks
					2S06-1.5	2S07	2S08-1 & 2S09-1 (Replicates)		AB02	EB06	
Laboratory Sample ID Numbers					1698	1700	1702	1704	315	1688	#5-9693 #3&4-82493 #1&2-9493
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	μg/L	μg/L	mg/kg
DRPH	5-10	50-100	500 ^a	9.55-1,150	<100J ^b	<60J ^b	<50J ^b	<50J ^b	NA	<1,000J ^b	<1,000J
GRPH	0.2	2	100	<0.400-<9.0	22J ^b	<2J ^b	<2J ^b	<2J ^b	<100J ^b	<50J ^b	<2
RRPH (Approx.)	10-20	100-200	2,000 ^a	<480	<200	<120	<100	<100	NA	<1,000	NA
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-<1,500	<0.20	<0.10	<0.10	<0.10			
Benzene	0.002-0.004	0.020-0.04	0.5	<0.020-<0.300	<0.04	<0.02	<0.02	<0.02	<1	<1	<0.02
Toluene	0.002-0.004	0.020-0.04		<0.020-<0.300	<0.04	<0.02	<0.02	<0.02	<1	<4J	<0.02
Ethylbenzene	0.002-0.004	0.020-0.04		<0.020-<0.300	<0.04	<0.02	<0.02	<0.02	<1	<2J	<0.02
Xylenes (Total)	0.004-0.008	0.04-0.08		<0.040-<0.600	<0.08	<0.04	<0.04	<0.04	<2	<5J	<0.04
HVOCs	0.01-0.02	0.1-0.2		<0.5J	<0.2	<0.1	<0.1	<0.1	<1	<1-<5J	<0.1

□ CT&E Data.

■ F&B Data.

■ NA
Not analyzed.

J Result is an estimate.

The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined.

DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

TABLE D-13. OLD DUMP SITE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Old Dump Site (LF19)				Matrix: Sediment Units: mg/kg						
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples		Field Blanks		Lab Blanks	
					SD01	2SD02	AB02	EB05	TB05	
Laboratory Sample ID Numbers					351	1696	315 4303-1	392/332 4303-5	375	#3&4-82493 #5-9593 #3&4-9693 #1&2-82493
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	μg/L	μg/L	μg/L	mg/kg
DRPH	6	60	500 ^a	9.55-1,150	580 J ^b	<60 J ^b	NA	<1,000 ^b	NA	<50J
GRPH	0.2-2.7	2-27	100	<0.400-<9.0	<27 J ^b	2 J ^b	<100 J ^b	<100 J ^b	<50 J ^b	<2J
RRPH (Approx.)	12	120	2,000 ^a	<480	5,800	<120	NA	<1,000	NA	<100
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-<1.500	0.03R	<0.10				
Benzene	0.002	0.02	0.5	<0.020-<0.300	<0.02R	<0.02	<1	<1	<1	<0.02J
Toluene	0.002	0.02		<0.020-<0.300	<0.02R	<0.02	<1	<1	<1	<0.02
Ethylbenzene	0.002	0.02		<0.020-<0.300	0.03R	<0.02	<1	<1	<1	<0.02
Xylenes (Total)	0.004	0.04		<0.040-<0.600	<0.04R	<0.04	<2	<2	<2	<0.04
HVOC (8010 Mod.)	0.002-0.001	0.02-0.1		<0.5J	<0.02R	<0.1	<1	<1	<1	<0.1

□ CT&E Data.

■ F&B Data.

■ NA Not analyzed.

J Result is an estimate.

R The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have not yet been determined.

a DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

b

TABLE D-13. OLD DUMP SITE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Old Dump Site (LF19)		Matrix: Surface Water Units: µg/L		Environmental Sample		Field Blanks		Lab Blanks	
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bk'd. Levels	SW01	AB02	EB05	TB05	Lab Blanks
Laboratory Sample ID Numbers					307 4303-6	315 4303-1	392/332 4303-5	375	#3&4-82493 #1&2-82493 4303
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L
GRPH	10	100		<20	<100J ^b	<100J ^b	<100J ^b	<50J ^b	<50J-<100J
BTX (8020/8020 Mod.)									
Benzene	0.1	1	5	<1	<1	<1	<1	<1	<1
Toluene	0.1	1	1,000	<1	<1	<1	<1	<1	<1
Ethylbenzene	0.1	1	700	<1	<1	<1	<1	<1	<1
Xylenes (Total)	0.2	2	10,000	<2	<2	<2	<2	<2	<2
HVOC 8010	0.1	1		NA	<1	<1	<1	<1	<1
VOC 8260	1	1		<1	<1-7.5U	5.1	<1	NA	<1
SVOC 8270	10	33		<10	<33	NA	<11	NA	<10
TOC	5,000	5,000		<5,000-12,700	34,600J	NA	<5,000J	NA	<5,000
TSS	100	200		<30,000-8,000	51,000J	NA	NA	NA	<200
TDS	10,000	10,000		<352,000-328,000	1,800,000J	NA	NA	NA	12,000

□ CT&E Data.

■ F&B Data.

NA Not analyzed.

The analyte was detected in the associated blank.

Result is an estimate.

Compound is not present above the concentration listed.

GRPH concentrations reported for these samples are equivalent to gasoline range organics (GRO) as defined by ADEC.

□ ■ NA B J U b

TABLE D-13. OLD DUMP SITE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Old Dump Site (LF19)			Matrix: Surface Water Units: µg/L		METALS ANALYSES: TOTAL (DISSOLVED)							Field Blank		Lab Blank
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	SW01							EB05		
Laboratory Sample ID Numbers					4303-6							4303-5		4303
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L							µg/L		µg/L
Aluminum	17.4	100		<100-350 (<100-340)	160 (<100)							<100		<100 (<100)
Antimony	N/A	100	6	<100 (<100)	<100 (<100)							<100		<100 (<100)
Arsenic	5.3	100	50	<100 (<100)	<100 (<100)							<100		<100 (<100)
Barium	1.2	100	2,000	<50-93 (<50-91)	90 (140)							<50		<50 (<50)
Beryllium	N/A	50	4	<50 (<50)	<50 (<50)							<50		<50 (<50)
Cadmium	1.7	50	5	<50 (<50)	<50 (<50)							<50		<50 (<50)
Calcium	34.5	100		4,500-88,000 (4,100-86,000)	170,000 (160,000)							<200		<200 (<200)
Chromium	3.29	50	100	<50 (<50)	<50 (<50)							<50		<50 (<50)
Cobalt	N/A	100		<100 (<100)	<100 (<100)							<100		<100 (<100)
Copper	2.3	50	1,300	<50 (<50)	<50 (<50)							<50		<50 (<50)
Iron	25	100		180-2,800 (<100-1,600)	18,000 (1,800)							200		<100 (<100)
Lead	6.6	100	15	<100 (<100)	<100 (<100)							<100		<100 (<100)

☐ CT&E Data.
N/A Not available.

TABLE D-13. OLD DUMP SITE ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Old Dump Site (LF19)			Matrix: Surface Water Units: µg/L		METALS ANALYSES: TOTAL (DISSOLVED)									
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Environmental Sample						Field Blank		Lab Blank	
					SW01									EB05
Laboratory Sample ID Numbers					4303-6							4303-5		4303
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L							µg/L		µg/L
Magnesium	47.8	100		<5,000-53,000 (2,600-54,000)	72,000 (75,000)							<200		<200 (<200)
Manganese	1.24	50		<50-510 (<50-120)	70 (56)							<50		<50 (<50)
Molybdenum	N/A	50		<50 (<50)	<50 (<50)							<50		<50 (<50)
Nickel	5.5	50	100	<50 (<50)	<50 (<50)							<50		<50 (<50)
Potassium	1,154	100		<5,000 (<5,000)	7,600 (8,700)							<500		<5,000 (<5,000)
Selenium	62.4	100	50	<100 (<100)	<100 (<100)							<100		<100 (<100)
Silver	2.6	50	50	<50 (<50)	<50 (50)							<50		<50 (<50)
Sodium	27.7	100		8,400-410,000 (8,200-450,000)	250,000 (320,000)							<250		267 (267)
Thallium	0.57	5	2	<5 (<5)	<5 (<5)							<5		<5 (<5)
Vanadium	1.8	50		<50 (<50)	<50 (<50)							<50		<50 (<50)
Zinc	8.2	50		<50-160 (<50)	<50 (<50)							<50		<50 (<50)

☐ CT&E Data.
N/A Not available.

TABLE D-14. BLADDER DIESEL SPILL ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Bladder Diesel Spill (SS20)				Matrix: Soil/Sediment Units: mg/kg	METALS ANALYSES										
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Environmental Samples						Field Blank			Lab Blanks	
					S02-0.75	S03-1 & S04-1 (Replicates)		SD01					EB05		
Laboratory Sample ID Numbers					4305-6	4305-5	4305-4	4305-1							4305 4303
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg							µg/L
Aluminum	0.35	2		1,500-25,000	1,600	2,100	2,000	2,000						<100	<100
Antimony	N/A	2-58		<7.8-<230	<52	<53	<53	<58						<100	<100
Arsenic	0.11	5.2-5.8		<4.9-8.5	<5.2	<5.3	<5.3	<5.8						<100	<100
Barium	0.024	1		27-390	13	18	18	20						<50	<50
Beryllium	N/A	2.6-2.9		<2.6-6.4	<2.6	<2.7	<2.7	<2.9						<50	<50
Cadmium	0.33	2.6-2.9		<3.0-<36	<2.6	<2.7	<2.7	<2.9						<50	<50
Calcium	0.69	4		360-59,000	13,000	3,700	33,000	5,600						<200	<200
Chromium	0.066	1		<4.3-47	3.0	4.7	4.0	9.3						<50	<50
Cobalt	N/A	5.2-5.8		<5.1-12	<5.2	<5.3	<5.3	<5.8						<100	<100
Copper	0.045	1		<2.7-45	4.8	4.7	4.9	6.6						<50	<50
Iron	0.50	2		5,400-35,000	6,000	5,800	6,200	5,700						200	<100
Lead	0.13	2.5-2		<5.1-22	<5.2	11	8.5	9.2						<100	<100
Magnesium	0.96	4		360-7,400	5,600J	1,800	17,000	2,900J						<200	<200
Manganese	0.025	1		25-290	69	69	77	69						<50	<50
Molybdenum	N/A	2.6-2.9		<2.5-<11	<2.6	<2.7	<2.7	<2.9						<50	<50
Nickel	0.11	1		4.2-46	5.6	5.3	4.6	7.7						<50	<50
Potassium	23	260-290		<300-2,200	<260	<270	<270	<290						<5,000	<5,000

☐ CT&E Data.
☐ N/A Not available
☐ J Result is an estimate.

TABLE D-14. BLADDER DIESEL SPILL ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Bladder Diesel Spill (SS20)			Matrix: Soil/Sediment Units: mg/kg		METALS ANALYSES									
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Range from 7 DEW Line Installations	Environmental Samples						Field Blank		Lab Blanks	
					S02-0.75	S03-1 & S04-1 (Replicates)		SD01				EB05		
Laboratory Sample ID Numbers					4305-6	4305-5	4305-4	4305-1					4305 4303	
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg					$\mu\text{g/L}$	
Selenium	1.2	5.2-5.8		<7.8-<170	<5.2	<5.3	<5.3	<5.8				<100	<100	
Silver	0.53	2.6-2.9		<3-<110	<2.6	<2.7	<2.7	<2.9				<50	<50	
Sodium	0.55	5		<160-680	52	61	58	80				<250	<250-267	
Thallium	0.011	0.25-0.3		<0.2-<1.2	<0.25	<0.27	<0.28	<0.30				<5	<5	
Vanadium	0.036	1		6.3-59	4.9	6.0	6.3	6.7				<50	<50	
Zinc	0.16	1		9.2-95	13	2.7	21	30				<50	<50	

☐ CT&E Data.

TABLE D-14. BLADDER DIESEL SPILL ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Bladder Diesel Spill (SS20)		Matrix: Sediment Units: mg/kg		Environmental Samples		Field Blanks			Lab Blanks	
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	SD01	SD02	AB02	EB05	TB05	
Laboratory Sample ID Numbers					378 4305-1	381/382	315 4303-1	392/332 4303-5	375	#3&4-82493 #1&2-82493 4303
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	μg/L	μg/L	μg/L	mg/kg
DRPH	5	50	500 ^a	9.55-1,150	<50J ^b	<50J ^b	NA	<1,000 ^b	NA	<50J
GRPH	0.2-9	2-90	100	<0.400-<9.0	<90J ^b	<2F ^b	<100J ^b	<100J ^b	<50J ^b	<2J
RRPH (Approx.)	10	100	2,000 ^a	<480	<100	<100	NA	<1,000	NA	<100
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-<1.500	<0.13	<0.10R				
Benzene	0.002	0.02	0.5	<0.020-<0.300	<0.02	<0.02R	<1	<1	<1	<0.02J
Toluene	0.002	0.02		<0.020-<0.300	<0.02	<0.02R	<1	<1	<1	<0.02
Ethylbenzene	0.002-0.003	0.02-0.03		<0.020-<0.300	<0.03	<0.02R	<1	<1	<1	<0.02
Xylenes (Total)	0.004-0.006	0.04-0.06		<0.040-<0.600	<0.06	<0.04R	<2	<2	<2	<0.04
VOC 8260										
Naphthalene	0.020	0.020		<0.025-<0.500	0.280	NA	<1	<1	NA	<0.020
Toluene	0.020	0.020		<0.025-<0.500	0.097	NA	2.2	2.3	NA	<0.020

☐ CT&E Data.
☒ F&B Data.
☒ NA
☒ J
☒ R
☒ a
☒ b

Result is an estimate.
Result has been rejected.

The action levels for DRPH and RRPH are based on conversations with ADEC; final action levels have yet to be determined.
DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

TABLE D-14. BLADDER DIESEL SPILL ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Bladder Diesel Spill (SS20)			Matrix: Sediment Units: mg/kg								
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples		Field Blanks			Lab Blanks	
					SD01	SD02	AB02	EB05	TB05		
Laboratory Sample ID Numbers					378 4305-1	381/382	315 4303-1	392/332 4303-5	375	#3&4-82493 #1&2-82493 4303	#6-82393 #1&2-82493 4305
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	µg/L	µg/L	µg/L	µg/L	mg/kg
SVOC 8270											
Fluoranthene	0.200	2.40		<0.230-<3.50	1.70J	NA	NA	<11	NA	<10	<0.200
Phenanthrene	0.200	2.40		<0.230-<3.50	1.96J	NA	NA	<11	NA	<10	<0.200
Pyrene	0.200	2.40		<0.230-<3.50	1.26J	NA	NA	<11	NA	<10	<0.200
TOC				32,000-199,000	3.360	NA	NA	<5,000J	NA	<5,000	NA

☐ CT&E Data.
☐ Not analyzed.
☐ Result is an estimate.

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TABLE D-14. BLADDER DIESEL SPILL ANALYTICAL DATA SUMMARY (CONTINUED)

Installation: Barter Island Site: Bladder Diesel Spill (SS20)		Matrix: Surface Water Units: µg/L										
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples			Field Blanks			Lab Blanks	
					SW01	SW02 & SW03 (Duplicates)		AB02	EB05	TB05		
Laboratory Sample ID Numbers					398/388 4305-7	403/406	407/410	315 4303-1	392/332 4303-5	375	4305 4303	
ANALYSES	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	µg/L	
SVOCs	10	10		<10	<10	NA	NA	NA	<11	NA	<10	
TOC	5,000	5,000		<5,000-12,700	31,400	NA	NA	NA	<5,000J	NA	<5,000	
TSS	100	200		<30,000-8,000	84,000J	NA	NA	NA	NA	NA	<200	
TDS	10,000	10,000		<352,000-328,000	903,000J	NA	NA	NA	NA	NA	<10,000	

CT&E Data.

F&B Data.

Not analyzed.

Result is an estimate.

☐ CT&E Data.
☒ F&B Data.
☐ Not analyzed.
☐ Result is an estimate.

TABLE D-15. JP-4 SPILL ANALYTICAL DATA SUMMARY

Installation: Barter Island Site: JP-4 Spill (SS21)		Matrix: Soil Units: mg/kg											
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	Environmental Samples					Field Blanks			Lab Blanks
					S01 Composite	S02	S03	2S04	2S05	AB02	EB05	TB05	
Laboratory Sample ID Numbers					364	366 4301-11	369	1692	1694	315 4303-1	392/332 4303-5	375	#8-82393 #5-9593 #384-9693 #1&2-82493 4303
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/L	μg/L	μg/L	mg/kg
DRPH	5-10	50-100	500 ^a	9.55-1,150	<50J ^b	1,000J ^b	1,300J ^b	<100J ^b	<100J ^b	NA	<1,000 ^b	NA	<50J
GRPH	0.2-0.4	2-4	100	<0.400-<9.0	<4R ^a	300J ^b	280J ^b	<4J ^b	<2J ^b	<100J ^b	<100J ^b	<50J ^b	<2J
RRPH (Approx.)	10-20	100-200	2,000 ^a	<480	<110	<100	<100	<200	<200	NA	<1,000	NA	<100
BTEX (8020/8020 Mod.)			10 Total BTEX	<0.250-<1,500	<0.10R	48.2JN	36J	<0.20	<0.15				
Benzene	0.002-0.004	0.02-0.04	0.5	<0.200-<0.300	<0.02R	3.5JN	<0.04	<0.04	<0.03	<1	<1	<1	<0.02J
Toluene	0.002-0.004	0.02-0.04		<0.200-<0.300	<0.02R	7.3JN	4	<0.04	<0.03	<1	<1	<1	<0.02
Ethylbenzene	0.002-0.004	0.02-0.04		<0.200-<0.300	<0.02R	10JN	10J	<0.04	<0.03	<1	<1	<1	<0.02
Xylenes (Total)	0.004-0.008	0.04-0.08		<0.400-<0.600	<0.04R	27JN	24J	<0.06	<0.06	<2	<2	<2	<0.04
HVOC 8010	0.004-0.02	0.04-0.2		<0.5J	<0.2R	<0.2	<0.04	<0.2	<0.1	<1	<1	<1	<0.02J-<0.1
VOC 8260													
n-Butylbenzene	0.020	0.200		<0.025-<0.500	NA	4.20	NA	NA	NA	<1	<1	NA	<0.020
sec-Butylbenzene	0.020	0.200		<0.025-<0.500	NA	1.94	NA	NA	NA	<1	<1	NA	<0.020
tert-Butylbenzene	0.020	0.200		<0.025-<0.500	NA	0.345	NA	NA	NA	<1	<1	NA	<0.020

☐ CT&E Data.
☒ F&B Data.
☒ Not analyzed.

NA
 J N R a b

Result is an estimate.

The analysis indicates the presence of an analyte for which there is presumptive evidence to make a "tentative identification."

Result has been rejected.

The action levels for DRPH and RRPB are based on conversations with ADEC; final action levels have not yet been determined.

DRPH and GRPH concentrations reported for these samples are equivalent to diesel and gasoline range organics (DRO and GRO) as defined by ADEC.

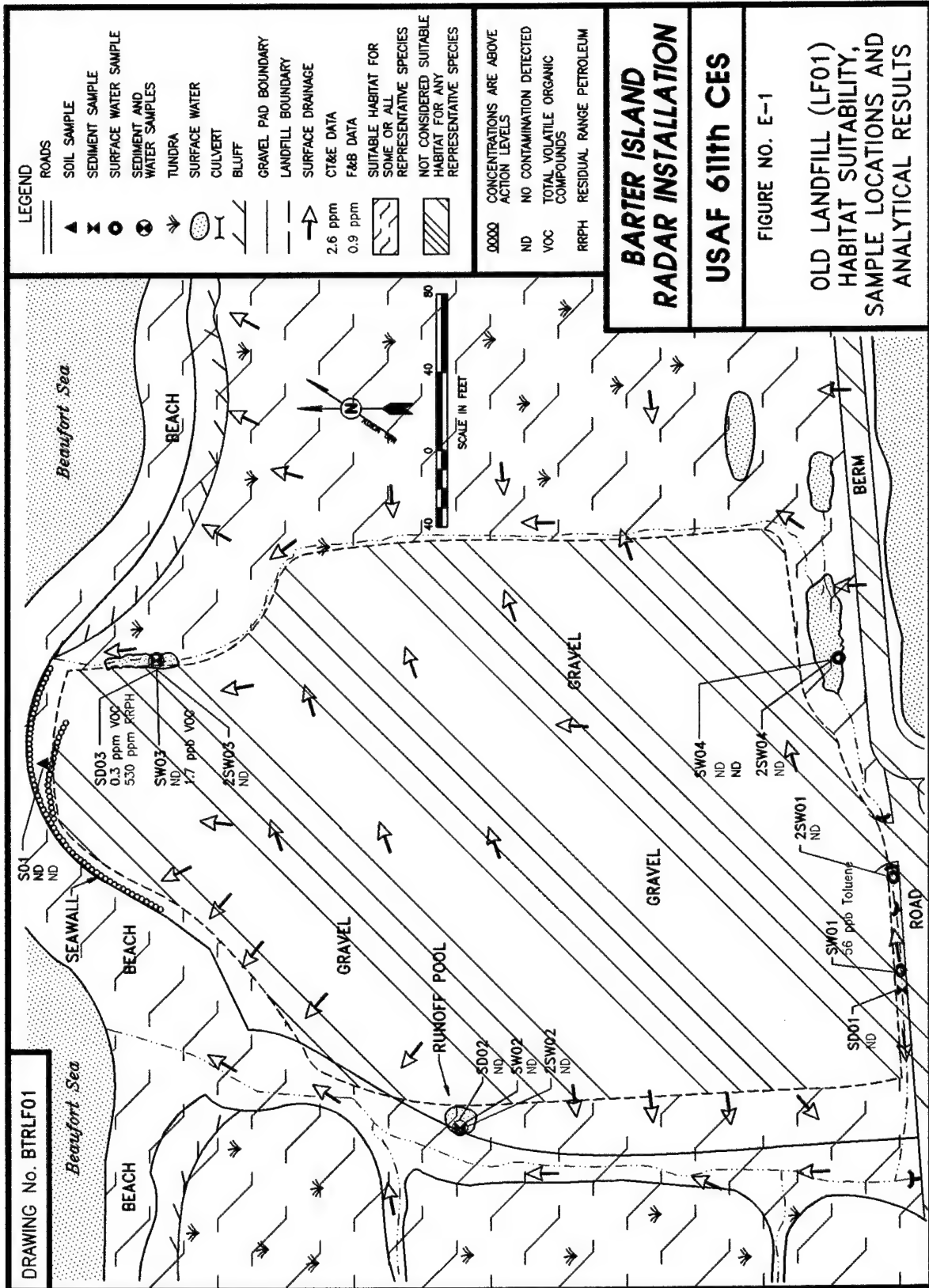
TABLE D-15. JP-4 SPILL ANALYTICAL DATA SUMMARY (CONTINUED)

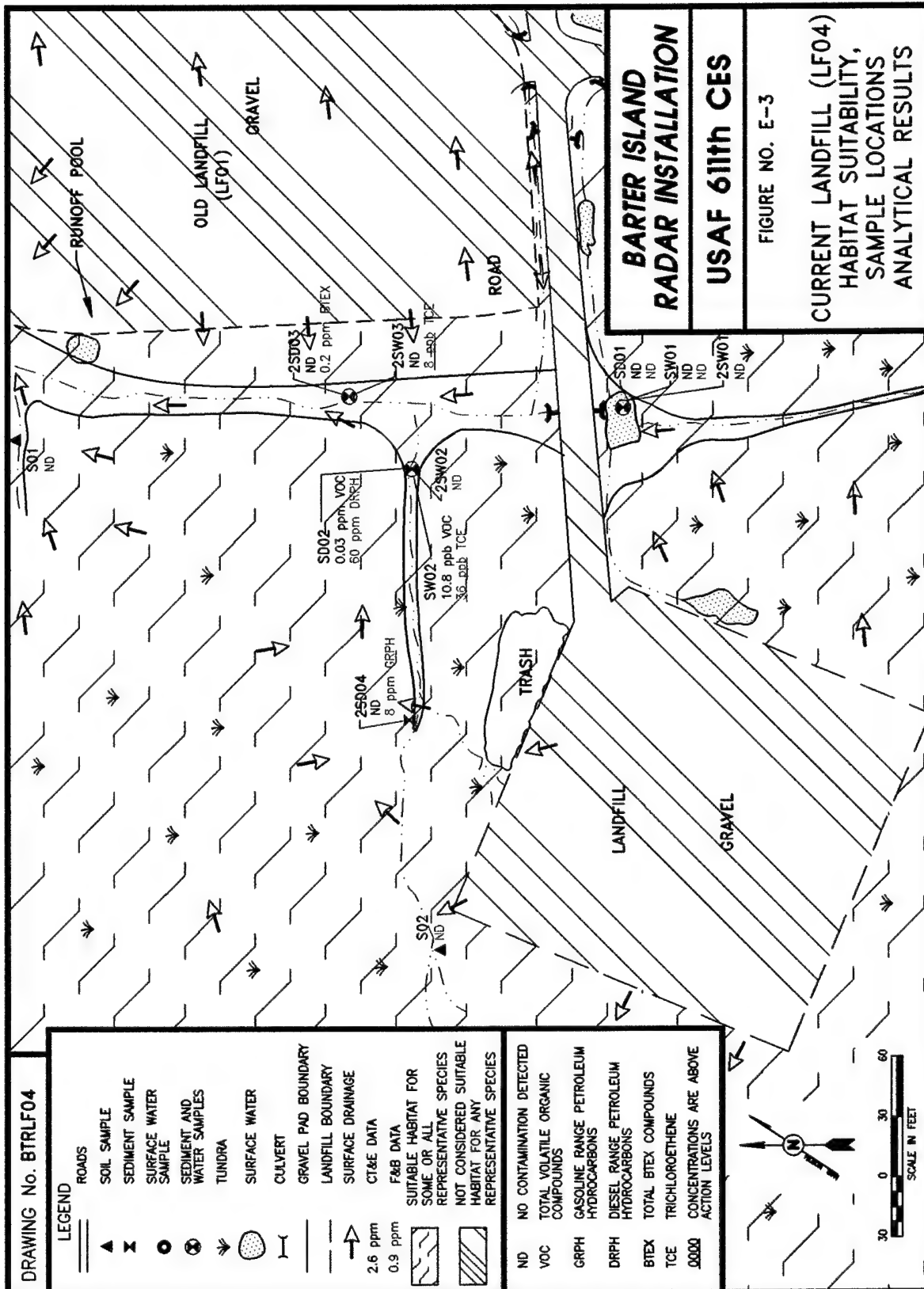
Installation: Barter Island Site: JP-4 Spill (SS21)		Matrix: Soil Units: mg/kg		Environmental Samples							Field Blanks			Lab Blanks	
Parameters	Detect. Limits	Quant. Limits	Action Levels	Bkgd. Levels	S01 Composite	S02	S03	2S04	2S05	AB02	EB05	TB05			
Laboratory Sample ID Numbers					364	366 4301-11	369	1692	1694	315 4303-1	392/332 4303-5	375	4303	#6-82393 #5-9593 4301	
ANALYSES	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/L	mg/L	μg/L	μg/L	mg/kg	
Ethylbenzene	0.020	0.200		<0.025-<0.500	NA	11.8	NA	NA	NA	<1	<1	NA	<1	<0.020	
Isopropylbenzene	0.020	0.200		<0.025-<0.500	NA	2.25	NA	NA	NA	<1	<1	NA	<1	<0.020	
p-Isopropyltoluene	0.020	0.200		<0.025-<0.500	NA	2.56	NA	NA	NA	<1	<1	NA	<1	<0.020	
Naphthalene	0.020	0.200		<0.025-<0.500	NA	5.54	NA	NA	NA	<1	<1	NA	<1	<0.020	
n-Propylbenzene	0.020	0.200		<0.025-<0.500	NA	3.26	NA	NA	NA	<1	<1	NA	<1	<0.020	
Toluene	0.020	0.200		<0.025-<0.500	NA	24.0	NA	NA	NA	2.2	2.3	NA	<1	<0.020	
1,2,4-Trimethylbenzene	0.020	0.200		<0.025-<0.500	NA	18.1	NA	NA	NA	<1	<1	NA	<1	<0.020	
1,3,5-Trimethylbenzene	0.020	0.200		<0.025-<0.500	NA	9.16	NA	NA	NA	<1	<1	NA	<1	<0.020	
Xylenes (Total)	0.040	0.400		<0.050-<1.000	NA	69.0	NA	NA	NA	<2	<2	NA	<2	<0.040	
Total BTEX			15	<0.125-<2.500	NA	104.8	NA	NA	NA						
SVOC 8270															
2-Methylnaphthalene	0.200	0.200		<0.230-<3.50	NA	1.42	NA	NA	NA	NA	<11	NA	<10	<0.200	
Naphthalene	0.200	0.200		<0.230-<3.50	NA	0.274	NA	NA	NA	NA	<11	NA	<10	<0.200	
Pesticides	0.001-0.05	0.01-0.5		<0.001-<0.100	<0.014-<0.51	NA	NA	NA	NA	NA	<0.2-<10	NA	NA	<0.01-<0.5	
TOC					NA	2.200	NA	NA	NA	NA	<5.000J	NA	<5.000	NA	

☐ CT&E Data.
☒ F&B Data.
☐ Not analyzed.
 Result is an estimate.

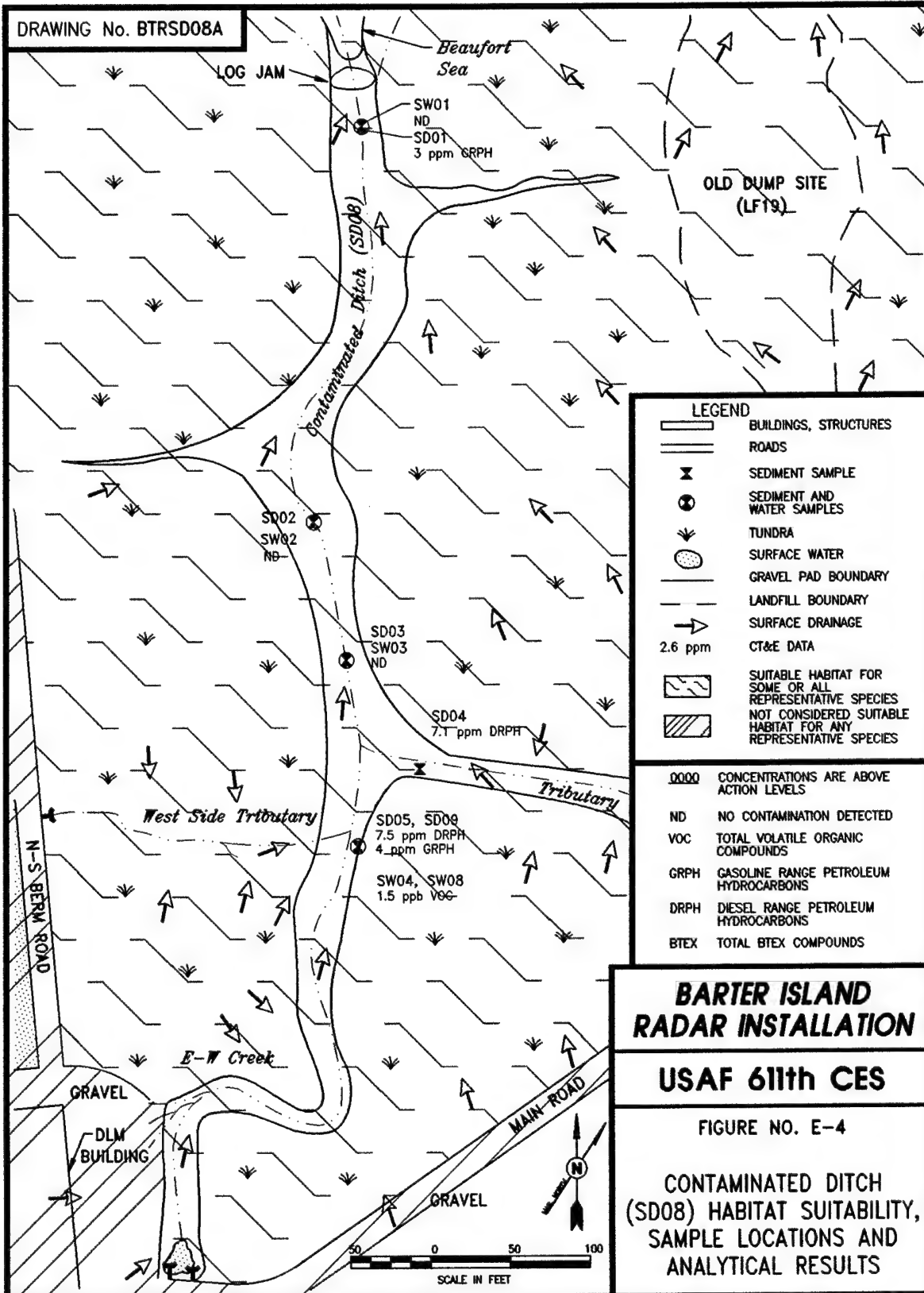
☐ NA
☒ J

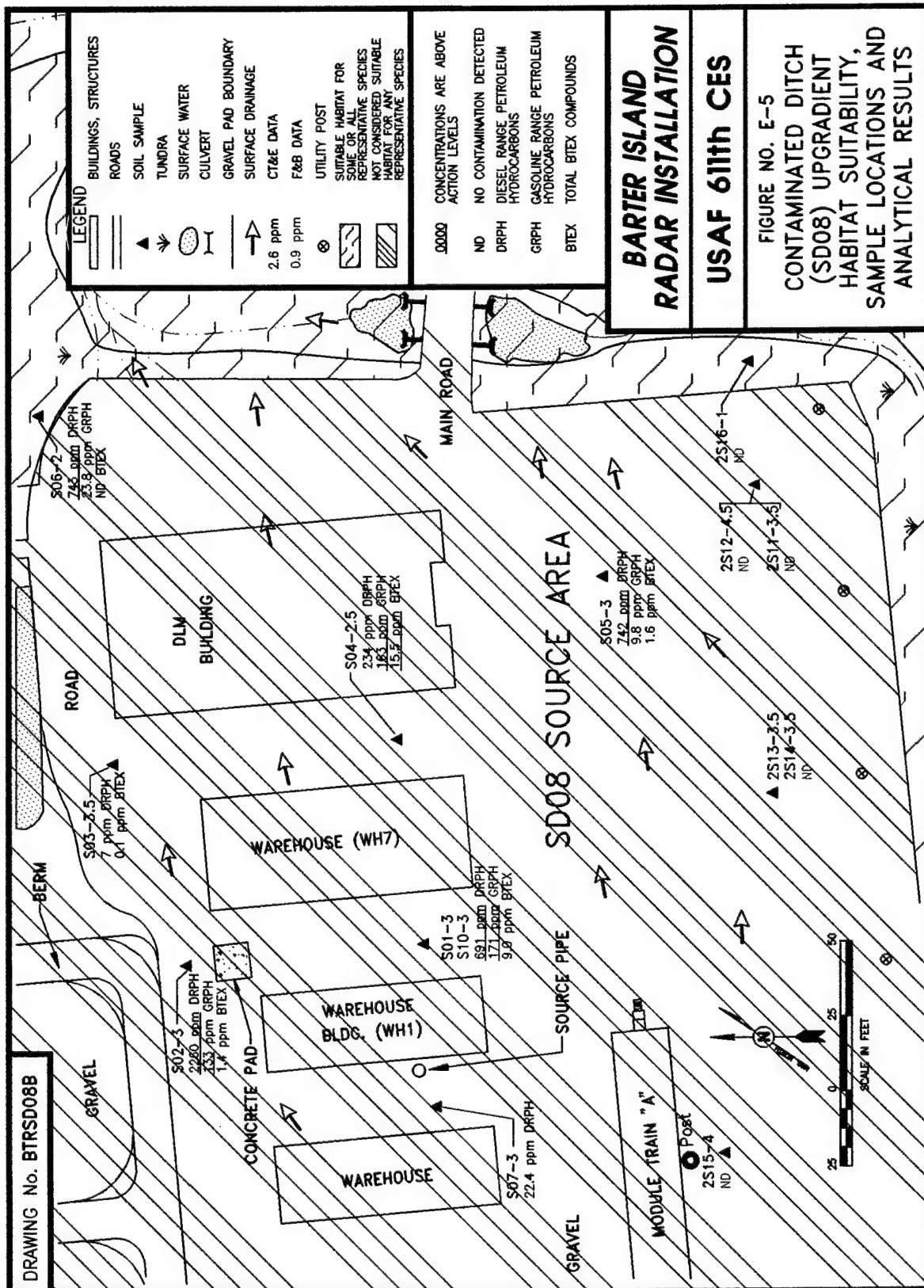
APPENDIX E
SITE HABITAT MAPS



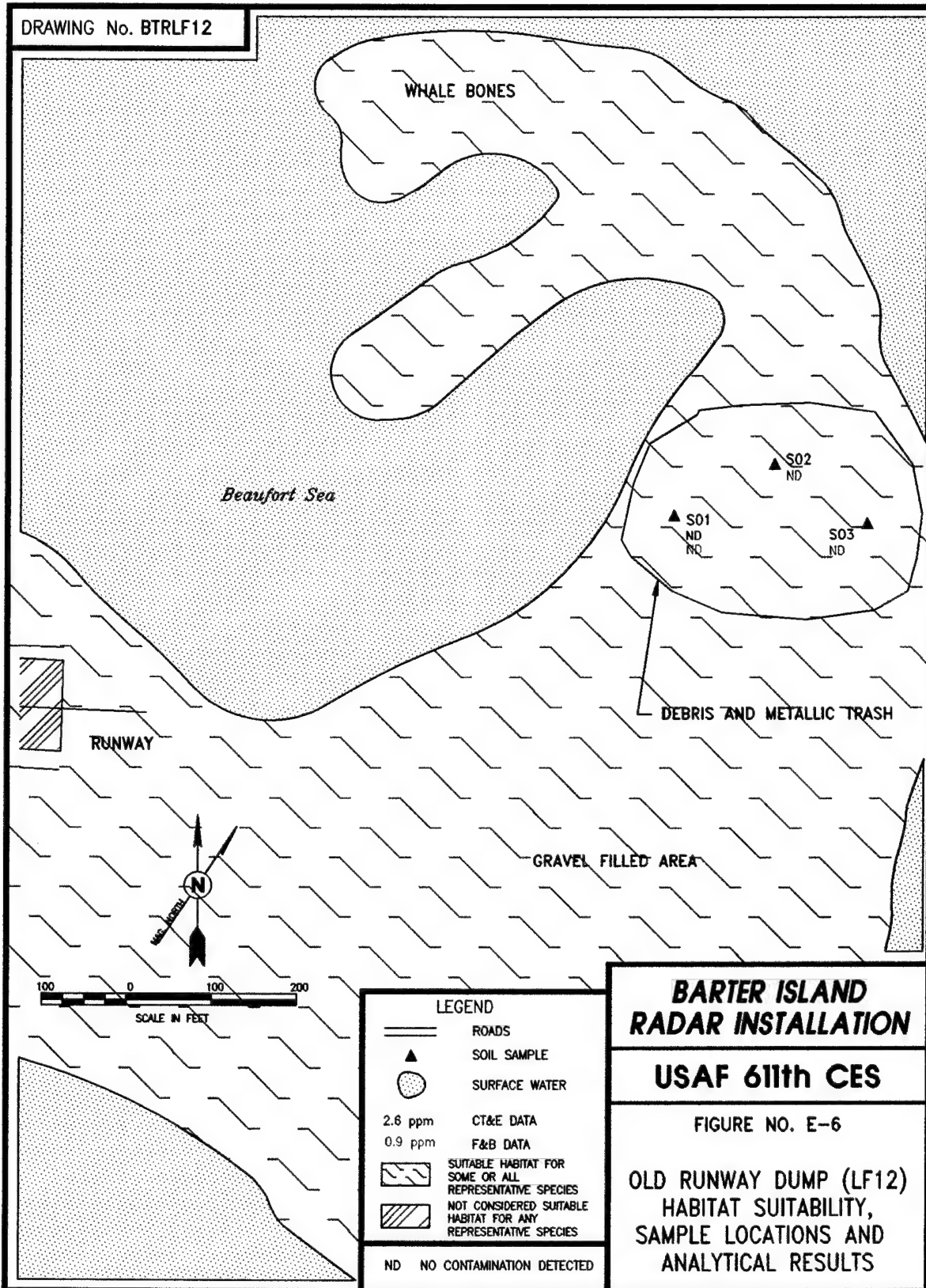


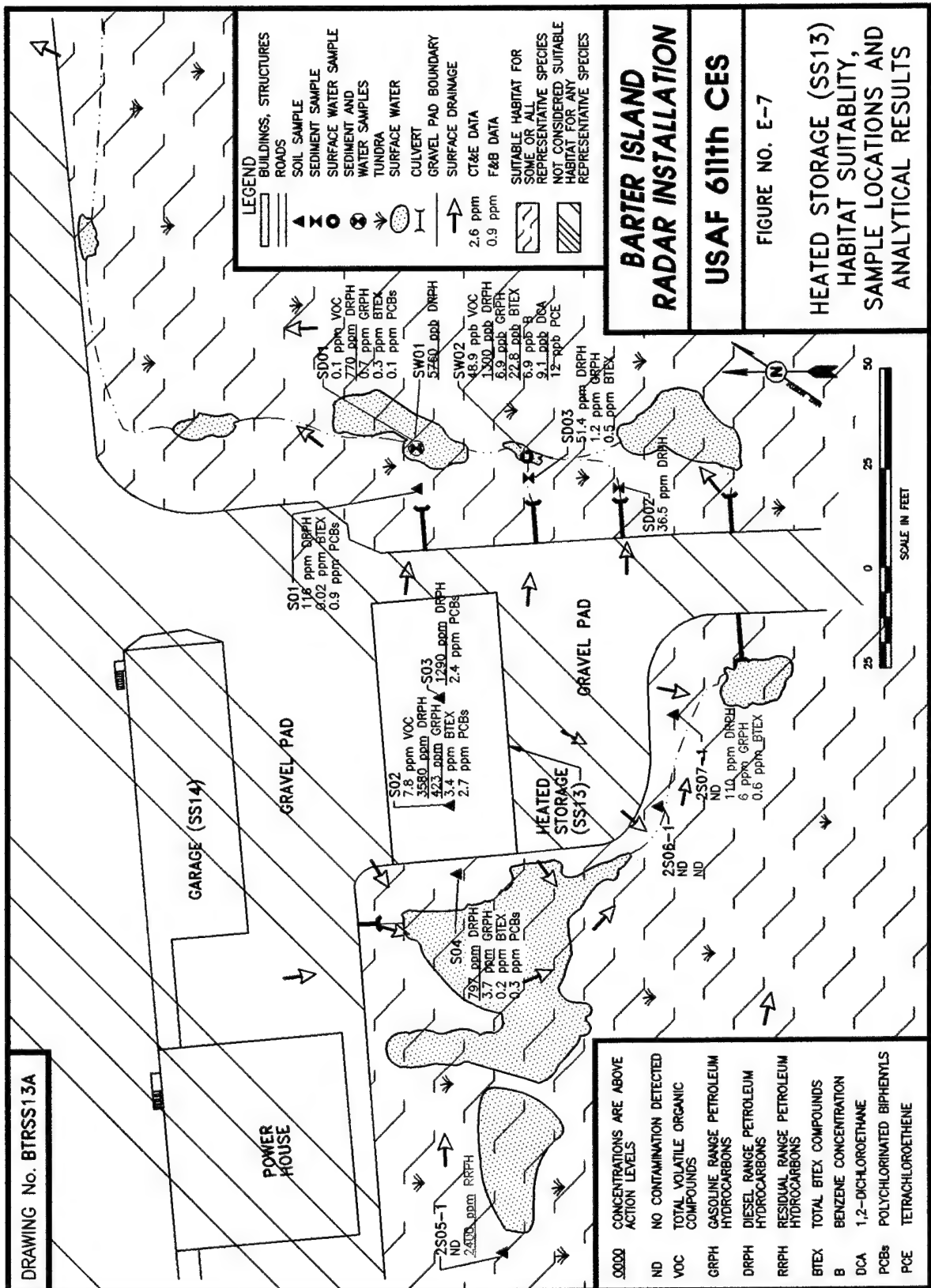
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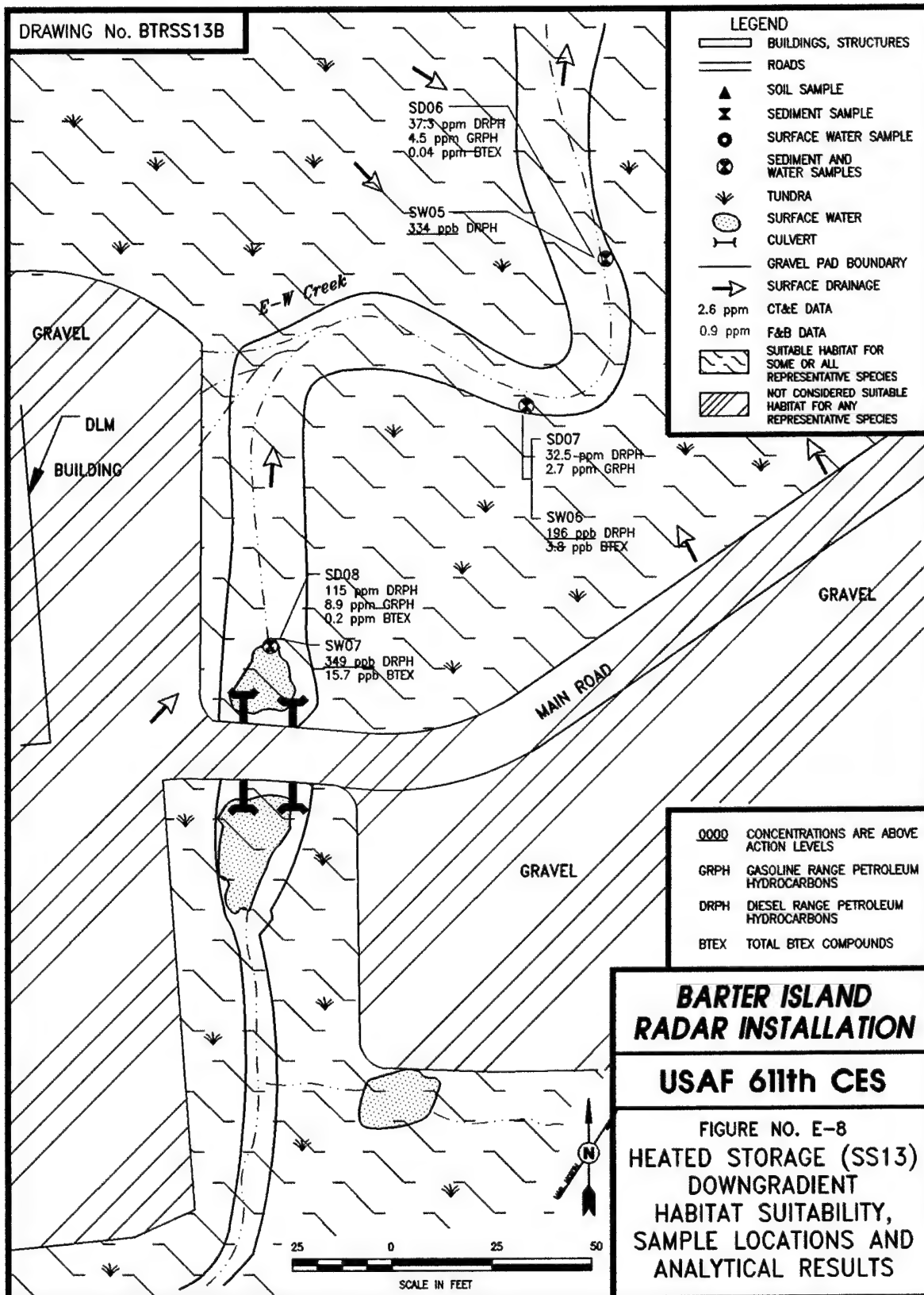


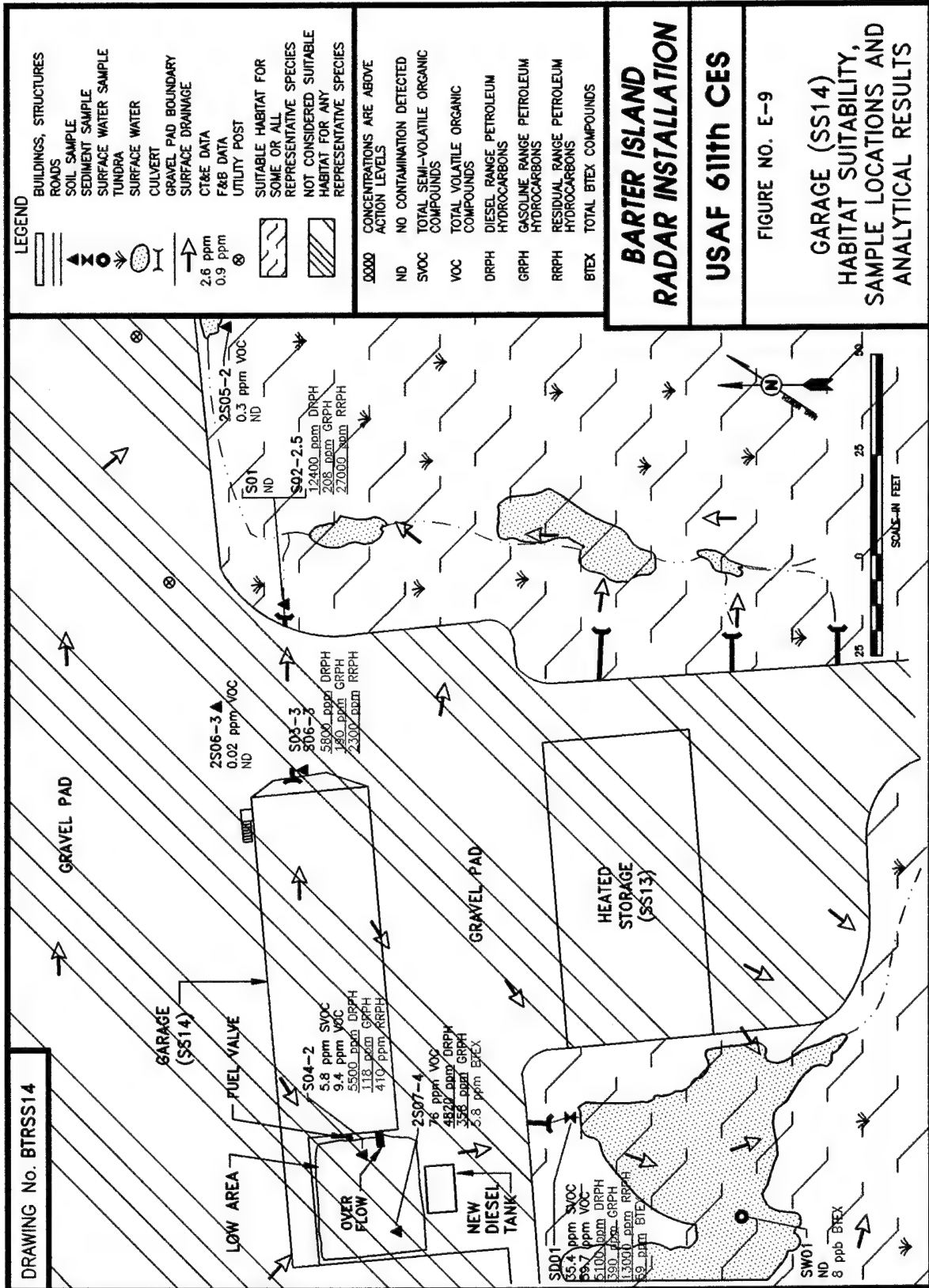


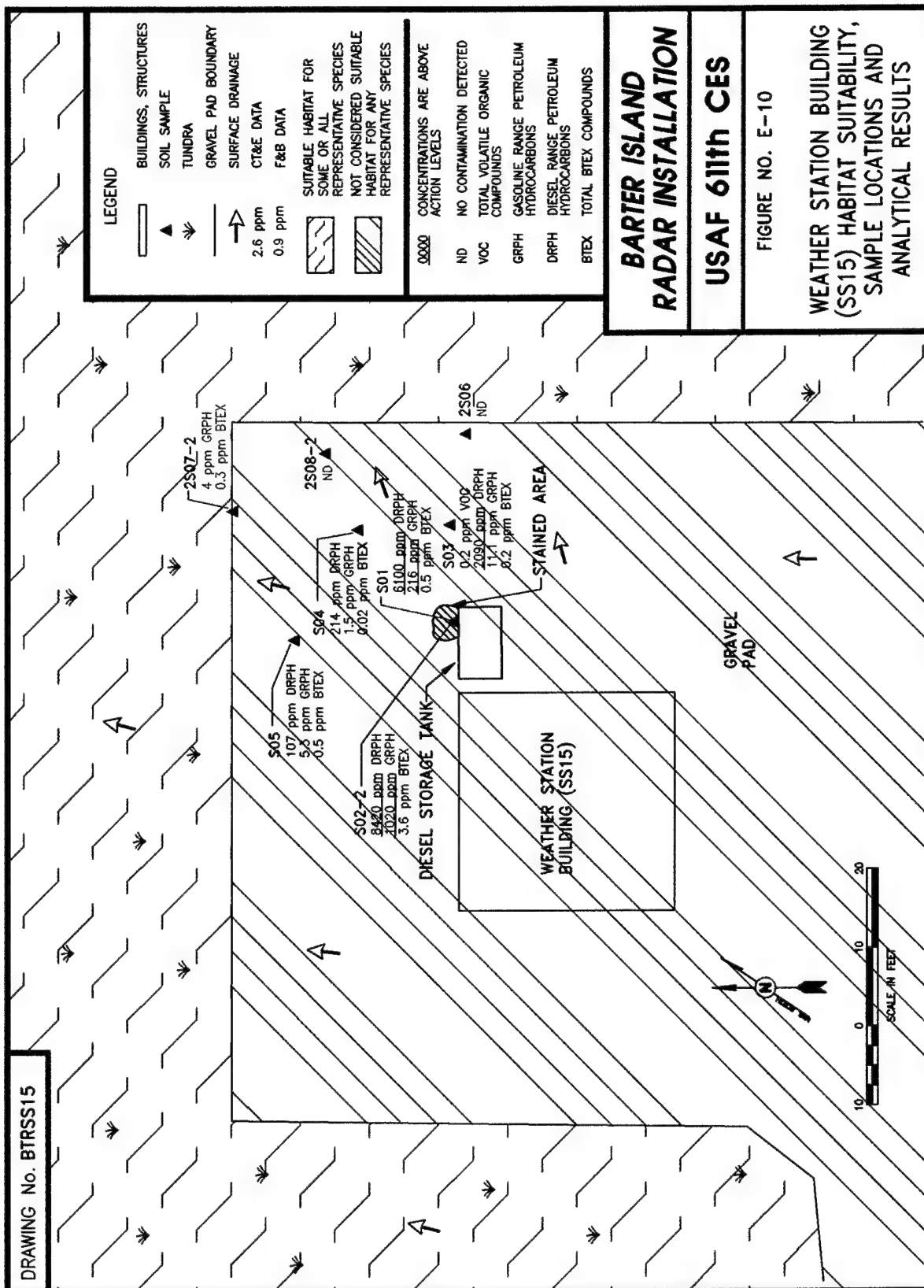
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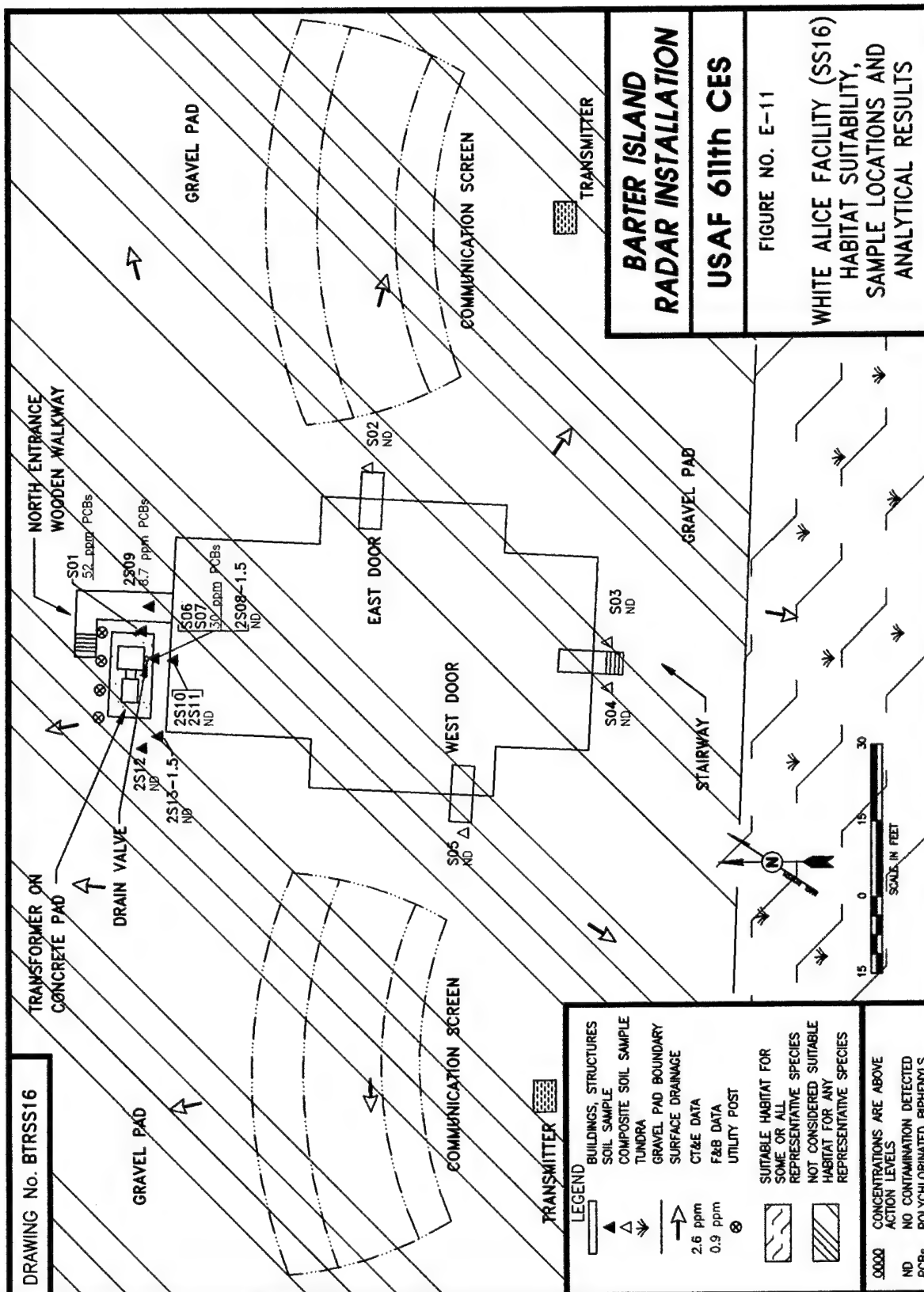


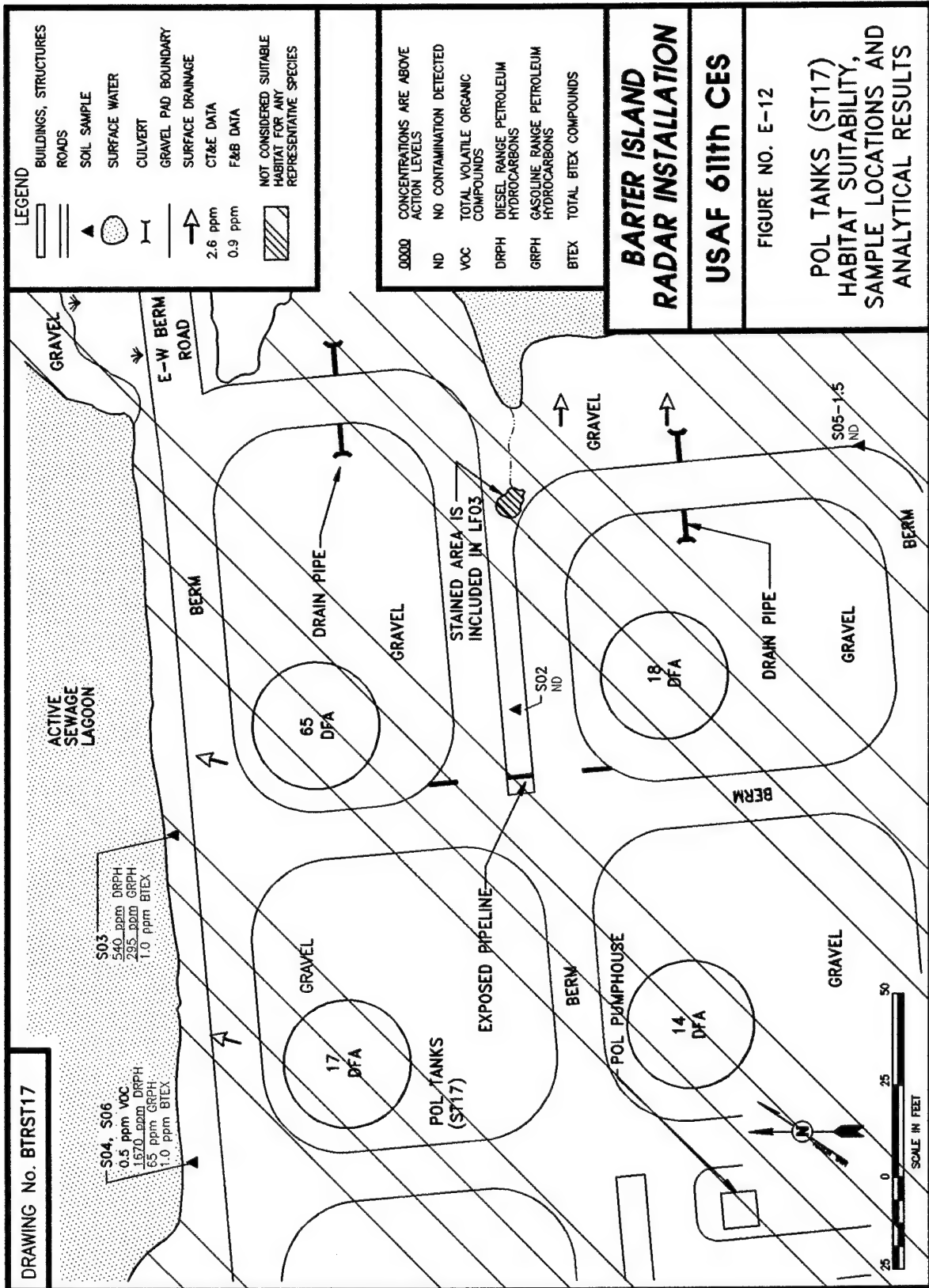


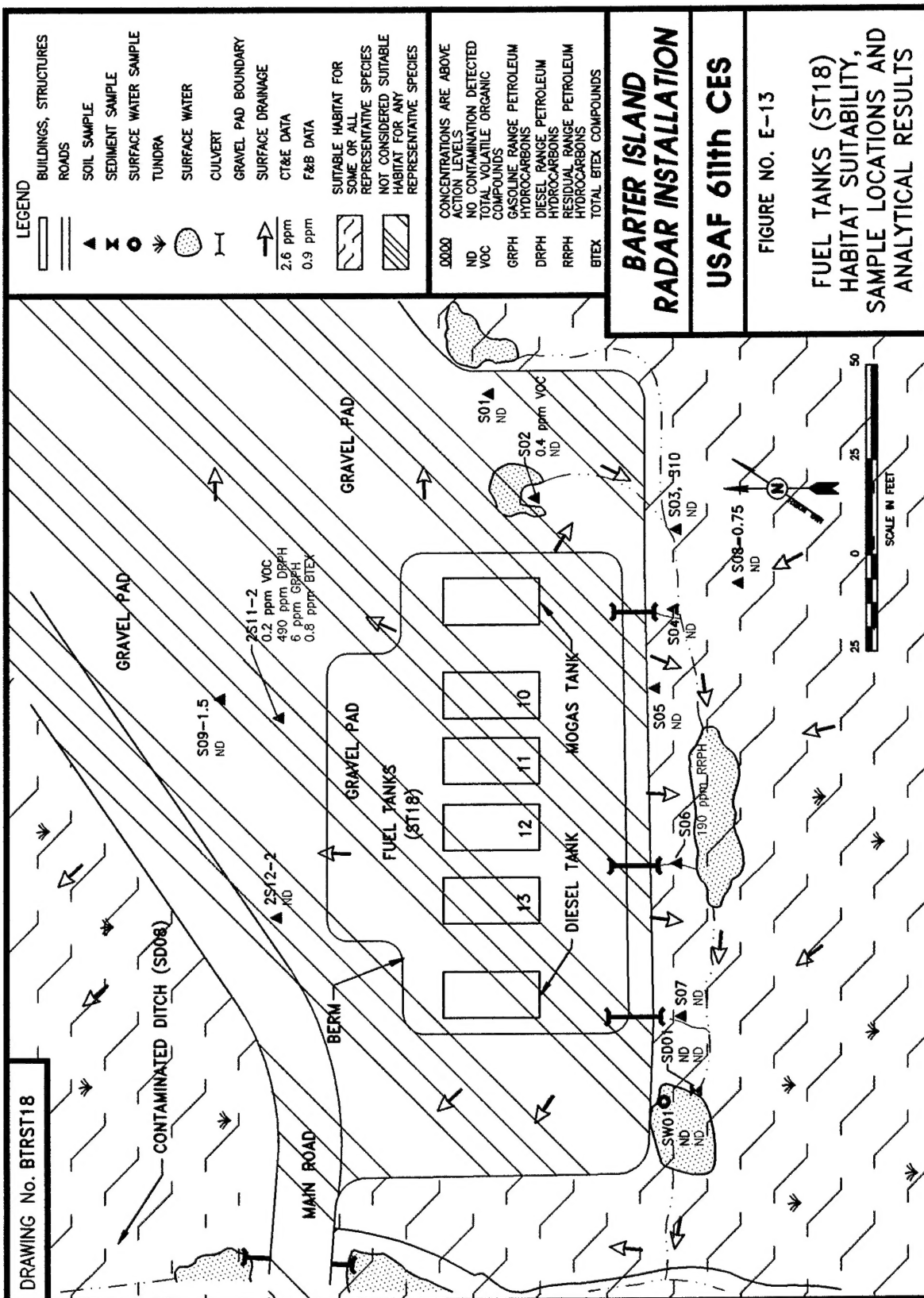




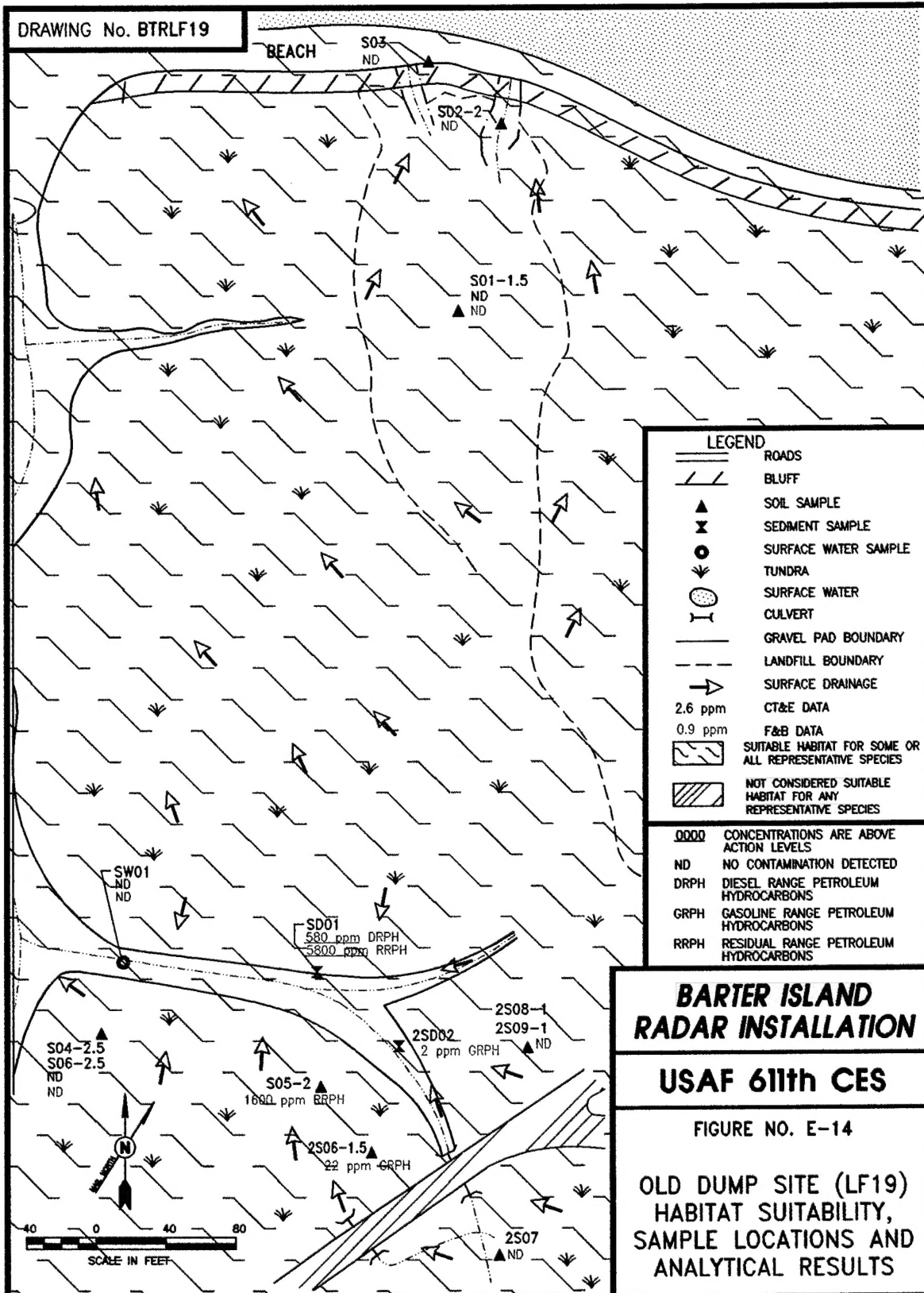








DRAWING No. BTRLF19





DRAWING No. BTRSS21

